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July 11, 2023

PPPO-02-10024209-23

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Dear Mr. Begley and Mr. Weeks:

TRANSMITTAL OF THE METHODS FOR CONDUCTING RISK ASSESSMENTS AND RISK EVALUATIONS AT THE PADUCAH GASEOUS DIFFUSION PLANT PADUCAH, KENTUCKY VOLUME 1. HUMAN HEALTH, DOE/LX/07-0107&D2/R14/V1

Enclosed please find the revised *Methods for Conducting Risk Assessments and Risk Evaluations at the Paducah Gaseous Diffusion Plant Paducah, Kentucky Volume 1. Human Health,* DOE/LX/07-0107&D2/R14/V1 (Risk Methods Document). This revision updates the previously issued document. Updates have been coordinated through the Risk Assessment Working Group. With these revisions, the Risk Methods Document has been updated to promote the development of human health risk assessments in accordance with the most current state and federal guidance. Enclosed also is a redline version of the document.

The U.S. Department of Energy requests acknowledgement of receipt of the subject document no later than July 28, 2023.

If you have any questions or require additional information, please contact Richard Bonczek at (859) 321-7127.

Sincerely,

APRIL LADD Date: 2023.07.11
15:41:50 -05:00'

April Ladd Federal Facility Agreement Manager Portsmouth/Paducah Project Office

Enclosures:

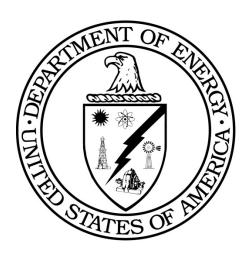
- 1. Methods for Conducting Risk Assessments and Risk Evaluations at the Paducah Gaseous Diffusion Plant Paducah, Kentucky Volume 1. Human Health, DOE/LX/07-0107&D2/R14/V1—Clean
- 2. Methods for Conducting Risk Assessments and Risk Evaluations at the Paducah Gaseous Diffusion Plant Paducah, Kentucky Volume 1. Human Health, DOE/LX/07-0107&D2/R14/V1—Redline

Administrative Record File—ARF ARR

cc w/enclosures:

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Methods for Conducting Risk Assessments and Risk Evaluations at the Paducah Gaseous Diffusion Plant Paducah, Kentucky Volume 1. Human Health



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DOE/LX/07-0107&D2/R14/V1

Methods for Conducting Risk Assessments and Risk Evaluations at the Paducah Gaseous Diffusion Plant Paducah, Kentucky Volume 1. Human Health

Date Issued—June 2023

U.S. DEPARTMENT OF ENERGY Office of Environmental Management

Prepared by
FOUR RIVERS NUCLEAR PARTNERSHIP, LLC,
managing the
Deactivation and Remediation Project at the
Paducah Gaseous Diffusion Plant
under Task Order DE-EM0004895

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PREFACE

This Methods for Conducting Risk Assessments and Risk Evaluations at the Paducah Gaseous Diffusion Plant, Paducah Kentucky, Volume 1. Human Health, DOE/LX/07-0107&D2/R14/V1 (previous versions DOE/LX/07-0107&D2/R13/V1, DOE/LX/07-0107&D2/R12/V1, issued DOE/LX/07-0107&D2/R11/V1. DOE/LX/07-0107&D2/R10/V1, DOE/LX/07-0107&D2/R9/V1, DOE/LX/07-0107&D2/R8/V1, DOE/LX/07-0107&D2/R7/V1, DOE/LX/07-0107&D2/R6/V1, DOE/LX/07-0107&D2/R4/V1, DOE/LX/07-0107&D2/R3/V1, DOE/LX/07-0107&D2/R5/V1, DOE/LX/07-0107&D2/R2/V1, DOE/LX/07-0107&D2/R1/V1, and DOE/OR/07-1506&D1/V1/R1), was prepared in accordance with the requirements under both the Comprehensive Environmental Response, Compensation, and Liability Act (CERCLA) and the Resource Conservation and Recovery Act (RCRA). This document is not meant to be prescriptive, rather it is meant to provide guidance for the completion of risk analyses beyond the guidance found in the most recent revision of the Paducah Site Management Plan (DOE 2022a). Specifically, this document integrates results of comment resolution meetings and technical meetings between the regulatory agencies and the U.S. Department of Energy and provisions in the Federal Facility Agreement (FFA) for the Paducah Gaseous Diffusion Plant (PGDP) (EPA 1998a) and provides methods that should be followed when completing risk analyses to ensure consistency in risk analyses. Risk analyses considered in this document are human health risk assessments and risk evaluations prepared for both informal and formal reports. This document and its appendices, including preliminary remediation goal values, are for use at PGDP and are not applicable to other sites within the Commonwealth of Kentucky.

In accordance with Section IV of the FFA for PGDP, this integrated technical document was developed to satisfy both CERCLA and RCRA corrective action requirements. The phases of the investigation process are referenced by CERCLA terminology within this document to reduce the potential for confusion.



CONTENTS

PR	EFAC	E		iii
FIC	GURE	S		vii
TA	BLES	S		vii
EX	HIBI	ΓS		vii
AC	CRON	YMS		ix
EX	ECU	ΓIVE SU	JMMARY	ES-1
1.	INTI	RODUC	TION	1-1
2.	RISK	X ANAI	YSES DURING SCOPING ACTIVITIES	2-1
	2.1	ANAL	YSES SUPPORTING ACTION PRIOR TO RI/FS	2-11
	2.2		YSES SUPPORTING NO FURTHER ACTION DECISIONS	
	2.3		YSES USED TO PRIORITIZE FURTHER INVESTIGATIONS	
3.	RISK	X ANAI	LYSES DURING THE REMEDIAL INVESTIGATION	3-1
	3.1	ANAL	YSES DURING WORK PLAN DEVELOPMENT AND IMPLEMENTATION	
		(SCRE	EENING RISK ASSESSMENTS)	3-1
		3.1.1	Analyses Appearing in the Introduction of the Integrated RI/FS Work Plan	3-1
		3.1.2	Analyses Appearing in Prior Characterization Chapter of the Integrated RI/FS Work Plan	3-4
		3.1.3	Analyses Appearing in Initial Evaluation Chapter of the Integrated RI/FS Work Plan	3-5
		3.1.4	Analyses Appearing in Remedial Alternatives Development Chapter of the Integrated RI/FS Work Plan	
	3.2	ANAI	YSES FOLLOWING COMPLETION OF THE INITIAL ROUND OF	3-7
	3.2		STIGATION	2 7
	3.3		YSES FOR THE RI REPORT (BASELINE RISK ASSESSMENTS)	
	5.5	3.3.1	Guidance Documents	
		3.3.2	General Methods.	
		3.3.3	Data Evaluation Methods	
		3.3.4	Exposure Assessment Methods	
		3.3.5	Toxicity Assessment Methods	
		3.3.6	Risk Characterization Methods	
		3.3.7	Consideration of Uncertainty in the Risk Assessment	
		3.3.8	Remedial Goal Option Derivation Methods	
4.	RISK	(ANAI	LYSES IN THE PREPARATION OF REMEDY SELECTION DOCUMENTS	4-1
	4.1		ANALYSES DURING THE FEASIBILITY STUDY	
	-	4.1.1	Presentation of Risk Assessment Results in the Feasibility Study	
		4.1.2	Modifications to Risk Assessment Parameters that May Appear in the	4_3

		4.1.3	Risk Analyses during the Identification and Screening of Technologies and	
			Alternatives	4-4
			Risk Analyses during the Detailed Analysis of Alternatives	
	4.2		ANALYSES AFTER THE FEASIBILITY STUDY	
		4.2.1	Risk Evaluation for the Proposed Remedial Action Plan	4-′
		4.2.2	Risk Evaluation for the ROD.	4-8
		4.2.3	Risk Analyses for Residual Risks	4-8
5.	REFI	ERENC!	ES	5-1
AP	PEND	OIX A:	SCREENING LEVELS (CURRENT AS OF OCTOBER 2022)	A-1
AP	PEND	OIX B:	DERIVATION OF PRELIMINARY REMEDIATION GOALS	B-1
AP	PEND	OIX C:	OUTLINE FOR BASELINE HUMAN HEALTH RISK ASSESSMENTS	C- 1
AP	PEND	OIX D:	EXPOSURE EQUATIONS	D-1
ΑP	PEND	IX E:	TECHNICAL INFORMATION (ON CD)	E-1

FIGURES

2.1.	General Approach to Risk-Based Site Scoping	2-2
2.2.	Data Quality Review to Support Risk-Based Site Scoping	
2.3.	Human Health Direct Contact Screening during Risk-Based Site Scoping	
2.4.	Groundwater Protection Screening during Risk-Based Site Scoping	
2.5.	Consideration of Additional Analyses during Risk-Based Site Scoping	
3.1.	Example Risk-Based Conceptual Site Model	
3.2.	Data Life Cycle	3-17
3.3.	Example of Step 3—Evaluation of Sample Quantitation Limits Laboratory Analytical	
	Data	3-19
	TABLES	
2.1.	Significant Chemicals and Radionuclides of Potential Concern at PGDP	2-7
3.1.	Toxicity Equivalency Factors for Carcinogenic PAH Compounds and Dioxins/Furans	3-22
3.2.	Modeling Matrix for Groundwater, Surface Water, and Biota	3-30
	EXHIBITS	
3.1.	Example Two-Way Table for Presentation of Historical Risk Assessment Results	
3.2.	Presentation of Screening Assessment Results in the RI Work Plan	
3.3.	Presentation of Background Comparison in the RI Work Plan	3-7
3.4.	Presentation of Screening Assessment Results to Evaluate Nature of Contamination in	
	Soil after the Initial Round of Sampling	3-8
3.5.	Presentation of Screening Assessment Results to Evaluate Extent of Contamination in	
	Surface Soil after the Initial Round of Sampling	3-9
3.6.	Presentation of Screening Assessment Results to Evaluate Extent of Contamination in	
	Subsurface Soil after the Initial Round of Sampling	3-10
3.7.	Presentation of Screening Assessment Results to Evaluate Nature of Contamination in	
• •	Groundwater after the Initial Round of Sampling	
3.8.	List of Chemicals or Radionuclides of Potential Concern	
3.9.	Summary of Pathway Analysis in the Exposure Assessment	
3.10.	Presentation of Exposure Point Concentrations	
3.11.	Chemical-Specific Parameters	
3.12.	Daily Intakes (Chronic Dose) for Receptor 1	
3.13.	Exposure Route Summary for the Current Use Scenario—Systemic Toxicity	
3.14.	Driving Contaminants Summary for Current Use Scenario—Systemic Toxicity	
3.15.	Summary of Risk Characterization	
3.16.	Summary of Uncertainty Analysis	
3.17.	Presentation of Remedial Goal Options	3-50



ACRONYMS

AOC area of concern

ARAR applicable or relevant and appropriate requirement

bgs below ground surface CAS Chemical Abstracts Service

CDI chronic daily intake

CERCLA Comprehensive Environmental Response, Compensation, and Liability Act

COC contaminant of concern

COPC chemical or radionuclide of potential concern

DAF dilution attenuation factor
DOE U.S. Department of Energy
DQA data quality assessment
DQO data quality objective
ED exposure duration

EE/CA engineering evaluation/cost analysis

ELCR excess lifetime cancer risk

EPA U.S. Environmental Protection Agency

EPC exposure point concentration FFA Federal Facility Agreement

FS feasibility study GI gastrointestinal

HEAST Health Effects Assessment Summary Tables
HHRAWG Human Health Risk Assessment Working Group

HI hazard index HO hazard quotient

IEUBK Integrated Exposure Uptake and Biokinetic

IRIS Integrated Risk Information System

IUR inhalation unit risk

KAR Kentucky Administrative Regulations

K_d distribution coefficient

KDEP Kentucky Department for Environmental Protection

K_p permeability coefficient

KYRHTAB Kentucky Radiation Health and Toxic Agents Branch
MARLAP Multi-Agency Radiological Laboratory Analytical Protocols
MARSSIM Multi-Agency Radiation Survey and Site Investigation Manual

MCL maximum contaminant level MDC minimum detectable concentration

MOC medium of concern

MQC minimum quantification concentration

MQO measurement quality objective

MUSLE Modified Universal Soil Loss Equation

OSWER Office of Solid Waste and Emergency Response

PEGASIS Portsmouth/Paducah Project Office Environmental Geographic Analytical Spatial

Information System

PGDP Paducah Gaseous Diffusion Plant

POC pathway of concern

PRA probabilistic risk assessment PRG preliminary remediation goal

RAGS Risk Assessment Guidance for Superfund

RAO remedial action objective

RAWG Risk Assessment Working Group

RCRA Resource Conservation and Recovery Act

RESRAD RESidual RADioactivity (model)

RfC reference concentration

RfD reference dose

RGA Regional Gravel Aquifer RGO remedial goal option RI remedial investigation ROD record of decision

SADA Spatial Analysis and Decision Assistance

SESOIL Seasonal Soil Model

SF slope factor
SI site investigation
SMP Site Management Plan
SQL sample quantitation limit
SSL soil screening level

SWMM Storm Water Management Model
SWMU solid waste management unit
TEF toxicity equivalence factor
UCL upper confidence limit

UCRS Upper Continental Recharge System
VISL vapor intrusion screening level
VOC volatile organic compound

XRF X-ray fluorescence

EXECUTIVE SUMMARY

This document describes the methods used to prepare the human health risk assessments and risk evaluations needed to complete remedial activities at the Paducah Gaseous Diffusion Plant (PGDP). This document is not meant to be prescriptive, rather it is meant to provide the framework to complete appropriate risk analyses for projects listed in the Paducah Site Management Plan (DOE 2022a) taking into account site-specific conditions at PGDP. The materials and methods presented in this document were developed following agreements reached between the U.S. Department of Energy (DOE) and the regulatory agencies during comment resolution meetings, in the Federal Facility Agreement (FFA), and at technical meetings. In this document, the human health risk analyses that will occur during each phase of remedial activities are discussed, analytical techniques are described, and several analytical tools are presented. By providing this material in a single document, consistency of human health risk assessments and evaluations performed for PGDP is ensured, thereby speeding the completion and review of risk assessments and risk evaluations. This document and its appendices, including preliminary remediation goal values, are for use at PGDP and are not applicable to other sites within the Commonwealth of Kentucky. Any endorsement of this document by Commonwealth agencies is limited to its use at PGDP.

PGDP was placed on the National Priorities List on May 31, 1994. In accordance with Section 120 of the Comprehensive Environmental Response, Compensation, and Liability Act (CERCLA), DOE entered into an FFA with the U.S. Environmental Protection Agency (EPA) and Kentucky on February 13, 1998 (EPA 1998a). The FFA established one set of consistent requirements for achieving comprehensive site remediation in accordance with the Resource Conservation and Recovery Act and CERCLA, including stakeholder involvement. The FFA requires that an evaluation of alternative remedies to address any release be conducted when a baseline risk assessment shows that the cumulative carcinogenic risk for an individual exposed to a given release, based on a reasonable maximum exposure for both current and future land use, is greater than 10⁻⁶, or a baseline risk assessment shows that the noncarcinogenic hazard quotient for an individual exposed to a given release, based on a reasonable maximum exposure for both current and future land use, is greater than 1.¹

This document also discusses some of the methods used to complete radiological dose assessments at PGDP. Radiological dose assessments are conducted to provide information for risk managers and are separate from the risk assessment conducted per the FFA for decision making. The methods for radiological dose assessment are presented generally, and additional discussion should be held with regulatory agencies prior to initiating any radiological dose assessment project that is part of an FFA project.

This document was prepared by the PGDP Risk Assessment Working Group (RAWG). The RAWG is a multiagency, multidisciplinary group tasked with meeting the following goals:

- Produce tools that can be used to prioritize remedial activities at the PGDP.
- Develop methods to complete risk evaluations for the PGDP.

¹ The FFA requires evaluation of alternative remedies if a baseline risk assessment shows noncarcinogenic hazard quotient greater than 1; however, the practice, according to this document and EPA guidance, is based on cumulative hazard index, not hazard from individual chemicals of potential concern.

- Make the results of the risk assessments and evaluations at the PGDP more useful to risk managers.
- Enhance risk communication between the producers of risk assessments and risk evaluations and the users of this information (e.g., risk managers).

Organizations participating in the production of this document and their affiliations are DOE, EPA, Commonwealth of Kentucky Energy and Environment Cabinet, and Commonwealth of Kentucky Radiation Health Branch.

1. INTRODUCTION

The purpose of this document is to present the methods and approaches used for screening level, baseline human health, and residual risk assessments and risk evaluations at the Paducah Gaseous Diffusion Plant (PGDP) and provide resources [such as preliminary remediation goals (PRGs) and radiological dose-based concentrations] for completing those assessments. This document is not meant to be prescriptive, rather it is meant to provide the framework to complete appropriate risk analyses for projects listed in the Paducah Site Management Plan (SMP) (DOE 2022a) taking into account site-specific conditions at PGDP. This document is not intended to replace or modify guidance from the U.S. Environmental Protection Agency (EPA), guidance from the Commonwealth of Kentucky, or any of the tripartite agreements. Analyses of risks and hazards presented by environmental contamination at PGDP are integral to the Federal Facility Agreement's (FFA) primary objective of implementing remedies that minimize, control, or eliminate risks to human health and the environment. These analyses begin during the scoping phase (e.g., during scoping meetings and during, for example, the preliminary assessment/site investigation) when available environmental media and historical information are interpreted and compared with site-specific PRGs and other screening criteria to determine if action may be required at release sites and to plan the timing of that action. These analyses continue during investigation [e.g., the remedial investigation (RI)] when historical information, site-specific PRGs, and other screening criteria are used to focus the work plan on the risk-related problems that must be investigated and may need to be addressed during data collection. Subsequently, the results of the risk analyses are used in decision documents to justify why an action is or is not needed at a site.² A more streamlined approach for risk assessments is sometimes used for removal action decision documents. During the production of the decision documents, the risk analyses also are used to develop the risk-based cleanup levels used in subsequent design activities.

Several major decision points occur during the aforementioned process. These decision points often limit the scope of risk analyses performed during investigation and remedy selection, but allow for interim actions to address important environmental concerns and occur several times during the process.

Risk assessors provide information at the decision points and risk managers use that information to make decisions. Risk assessors and managers and their roles are defined as follows (EPA 1989a).

- Risk Assessor. An individual, team, or organization that generates site- or media-specific risk assessments for use in site-specific decision making. The assessor relies on existing databases and information [e.g., EPA Integrated Risk Information System (IRIS), health assessment documents, and program-specific toxicity information] and media- or site-specific exposure information in characterizing risk. This group also relies, in part, on regulatory agency risk assessment guidelines and program-specific guidance to address scientific policy issues and scientific uncertainties.
- **Risk Manager.** An individual, team, or organization with responsibility for or authority to take action in response to an identified risk. Risk managers *integrate* the risk characterization information provided by the risk assessor with other considerations specified in applicable statutes to make and justify regulatory decisions. Generally, risk managers include lead and regulatory agency managers and decision makers. Risk managers also play a role in determining the scope of risk assessments.

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² There may be scenarios presented pursuant to this document that might not be commensurate with the reasonable foreseeable land use, but may serve as a reference point to decision makers.

This document presents the methods to be used to complete the analyses described herein. In addition, this document discusses many of the analytical tools that can be used to complete this process and discusses the sources of the tools. Materials and methods used to complete scoping activities, including the derivation of risk and radiological dose-based PRGs, the background concentrations of chemicals and radionuclides, and other screening criteria are in Section 2; materials and methods specific to the human health risk assessments, including work plan preparation and baseline human health risk assessment, are in Section 3, "Risk Analyses during the Remedial Investigation"; materials and methods applicable to the feasibility study (FS) risk evaluation, including cleanup level development and consideration of residual risks, are in Section 4.

Radiological dose assessments sometimes are provided to risk managers, as well, and also are discussed within these sections. The approach to radiological dose assessments discussed here is based on EPA guidance (EPA 2000a) and is specific to PGDP. The radiological dose-based concentrations are based on Federal Guidance Report 13 (EPA 1999a) and are not appropriate for other activities such as establishment of authorized limits. The exposure parameters used to derive the radiological dose-based concentrations presented are useful inputs when deriving authorized limits.

This Risk Methods Document discusses determination of cumulative risk for environmental media that are divided into separate operable units. According to the SMP, a final comprehensive site operable unit evaluation will occur following completion of each of the specific operable units at PGDP. The final comprehensive site operable unit will maximize use of the relevant data from previous cleanup activities and document the residual contamination and risk. The comprehensive site operable unit RI will include a sitewide baseline human health and ecological risk assessment to evaluate residual risks and ensure all actions taken to date, when considered collectively, are protective of human health and the environment from a sitewide perspective (DOE 2022b).

2. RISK ANALYSES DURING SCOPING ACTIVITIES

Risk analyses during site scoping activities will be performed to do the following:

- Determine if site risks are so great as to require immediate action prior to RI/FS (i.e., interim action);³
- Determine if site risks are so low as to support a no-further-action decision;
- Prioritize the further investigation of those sites not requiring an interim action or potentially requiring no further action;
- Divide exposure setting into exposure units; 4 and
- Provide information to be used in subsequent work plan development.

General depictions of the methods that will be followed to complete these analyses are shown in Figure 2.1. Figures 2.2, 2.3, 2.4, and 2.5 present specific issues related to the risk screening process (including issues related to radiological dose).

ALs and NALs

- ALs are concentrations of contaminants above which early actions may be warranted after consideration of project- and location-specific conditions.
- NALs are concentrations of contaminants below which no action is generally warranted.

Generally, analyses completed as part of risk-based site scoping will rely on simple comparisons between site contamination data to PGDP-specific PRGs, including risk-based action levels (ALs) and no action levels (NALs),⁵ radiological dose-based concentrations (if a radiological dose assessment is conducted), background concentrations, and potentially applicable or relevant and appropriate requirements (ARARs). Table 2.1 shows the significant chemicals or radionuclides of potential concern (COPCs) at PGDP with the chemical abstract services (CAS) number. Significant COPCs are

chemicals that have been retained as contaminants of concern

(COCs) (sometimes listed as constituents of concern) in prior risk assessments at PGDP. For the purposes of this document, these terms are essentially equivalent. These COPCs therefore are likely to be COPCs for other risk assessments, but the absence of a chemical from the list does not imply that it would not be a COPC at a PGDP site. Risk-based ALs and NALs are presented in Tables A.1 through A.7.

COPCs

Use of the terms "COPCs" and "chemicals" within this document is intended to include radionuclides, as applicable.

³ The report from this point forward will use references to remedial action documents instead of removal action documents for simplicity. If the approach for removal actions differs in the subsequent discussions, these differences will be noted, as appropriate. ⁴ A default exposure unit of 0.5 acres will be used for sites inside the PGDP industrialized area. For a site outside the industrialized area, the size of the exposure unit will be decided during scoping by agreement among the three parties.

⁵ Risk-based ALs are the lesser of the cancer-based values for excess lifetime cancer risk (ELCR) of 1 × 10⁻⁴ and hazard-based values for hazard index (HI) of 3. Risk-based NALs are the lesser of the cancer-based values for ELCR of 1 × 10⁻⁶ and hazard-based values for HI of 0.1. Cancer-based values are based on lifetime scenario for residential and recreational use (i.e., receptor intake is estimated over many years and multiple life stages and is then divided by an averaging time of 70 years to calculate a lifetime average daily dose). Hazard-based values are calculated separately for each lifestage (e.g., child and adult residents; child, teen, and adult recreational users), and the calculated average daily dose is specific to each lifestage. Hazard-based values for children are typically the most protective because of their higher daily intake rates (e.g., soil ingestion rate) paired with smaller body weights.

⁶ For Tables A.1 through A.7, the "a" subtables present ALs and NALs for chemicals; the "b" subtables present the ALs and NALs for radionuclides.

Risk Analyses during Site Scoping at PGDP General Approach

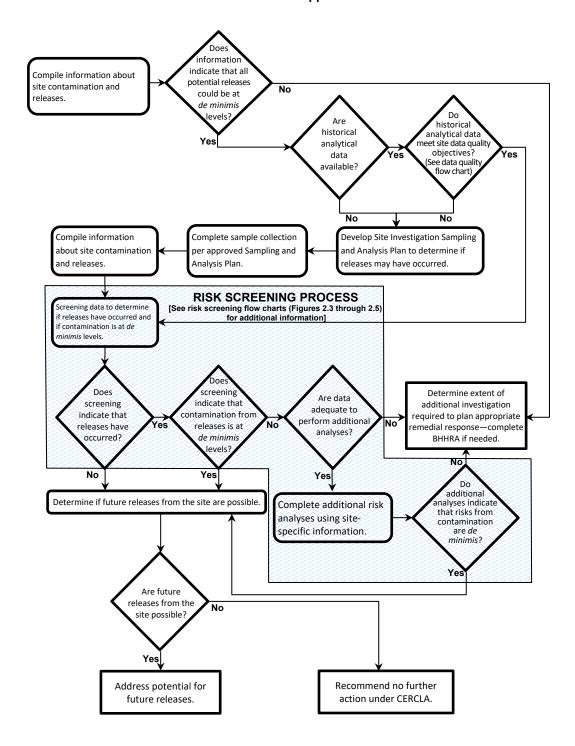


Figure 2.1. General Approach to Risk-Based Site Scoping

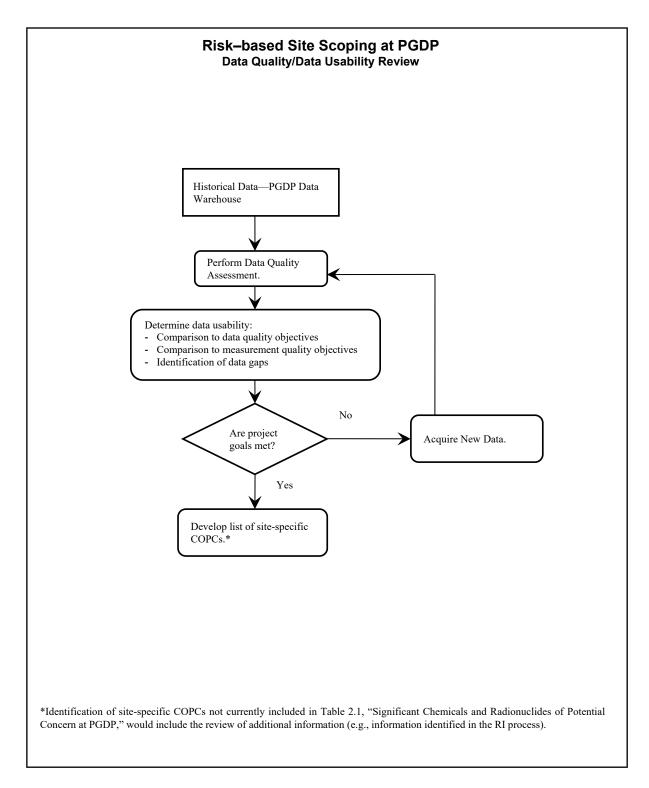


Figure 2.2. Data Quality Review to Support Risk-Based Site Scoping

Risk-Based Site Scoping at PGDP Part 1 – Human Health Direct Contact Screening Accumulate all sampling results; organize information by medium and location of sampling, including depth of sampling. Is contamination found where it is available for direct contact by Yes No a person? Compare maximum detected concentration of potential contaminant to contaminant's background concentration. Go to Part 2 to determine if contamination poses a risk to human health through migration to groundwater. Doesthe maximum concentration exceed the background concentration No Compare maximum detected concentration of contaminant to Doesthe Compare maximum detected maximum concentration of contaminant to values for human health found in concentration human health risk-based regulatory guidance materials (ARARs). exceed the guidance value? concentrations. Yes Does the concentration exceed the direct contact human health risk-based concentration?

NOTE: Guidance values are presented in Appendix A.

Maximum detected

concentration is

unacceptable versus risk-

based concentration for

direct human contact.

Figure 2.3. Human Health Direct Contact Screening during Risk-Based Site Scoping

Maximum detected

concentration is

unacceptable versus

guidance value.

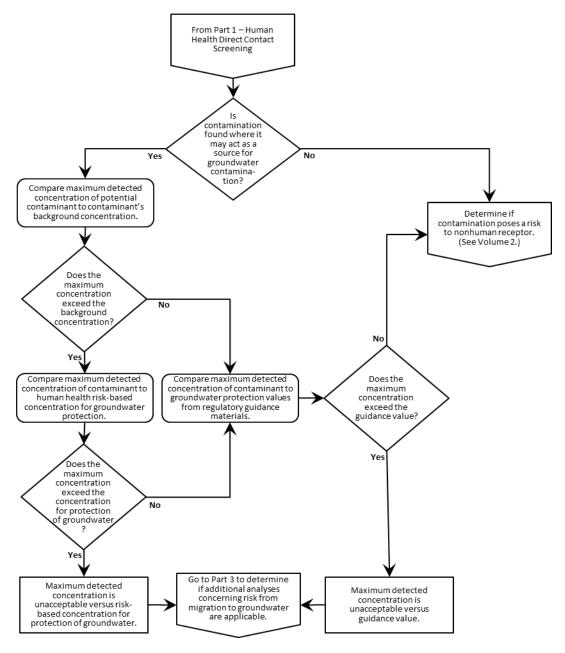
Go to Part 3 to determine

if additional analyses

concerning risk from direct

contact are applicable.

Risk-Based Site Scoping at PGDP Part 2 – Groundwater Protection Screening



NOTE: Guidance values are presented in Appendix A.

Figure 2.4. Groundwater Protection Screening during Risk-Based Site Scoping

Risk-Based Site Scoping at PGDP Part 3 – Consideration of Additional Analyses

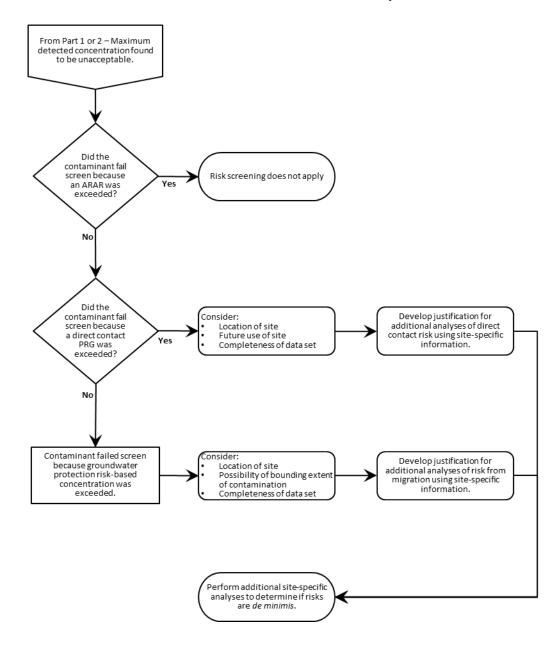


Figure 2.5. Consideration of Additional Analyses during Risk-Based Site Scoping

Table 2.1. Significant Chemicals and Radionuclides of Potential Concern at PGDP^{1,2}

Inorganic Chemicals		Organic Compounds		Radionuclides		
Analyte	CAS Number	Analyte	CAS Number	Analyte	CAS Number	
Aluminum	7429-90-5	Acenaphthene	83-32-9	Americium-241	14596-10-2	
Antimony	7440-36-0	Acenaphthylene	208-96-8	Cesium-137	10045-97-3	
Arsenic	7440-38-2	Acrylonitrile	107-13-1	Neptunium-237	13994-20-2	
Barium	7440-39-3	Anthracene	120-12-7	Plutonium-238	13981-16-3	
Beryllium	7440-41-7	Benzene	71-43-2	Plutonium-239	15117-48-3	
Boron	7440-42-8	Bromodichloromethane	75-27-4	Plutonium-240	14119-33-6	
Cadmium	7440-43-9	Carbazole	86-74-8	Technetium-99	14133-76-7	
Chromium III	16065-83-1	Carbon tetrachloride	56-23-5	Thorium-230	14269-63-7	
Chromium VI ³ Cobalt	18540-29-9	Chloroform 1,1-Dichloroethene	67-66-3	Thorium-232 Uranium-234	7440-29-1	
Copper	7440-48-4	1.2-Dichloroethane	75-35-4 107-06-2	Uranium-235	13966-29-5 15117-96-1	
Fluoride	7440-50-8 16984-48-8	1,2-Dichloroethene (mixed)	540-59-0	Uranium-238	7440-61-1	
Iron	7439-89-6	trans-1,2-Dichloroethene	156-60-5	Oramum-236	/ ++0-01-1	
Lead	7439-92-1	cis-1.2-Dichloroethene	156-59-2			
Manganese	7439-96-5	Dieldrin	60-57-1			
Mercury	7439-97-6	Ethylbenzene	100-41-4			
Molybdenum	7439-98-7	Fluoranthene	206-44-0			
Nickel	7440-02-0	Fluorene	86-73-7			
Selenium	7782-49-2	Hexachlorobenzene	118-74-1			
Silver	7440-22-4	Naphthalene	91-20-3			
Thallium	7440-28-0	2-Nitroaniline	88-74-4			
Uranium	NA	N-Nitroso-di-n-propylamine	621-64-7			
Vanadium	7440-62-2	Pentachlorophenol	87-86-5			
Zinc	7440-66-6	Phenanthrene	85-01-8			
		Pyrene	129-00-0			
		Tetrachloroethene	127-18-4			
		1,1,1-Trichloroethane	71-55-6			
		1,1,2-Trichloroethane	79-00-5			
		Trichloroethene ³	79-01-6			
		Total Dioxins/Furans	1746-01-6 37871-00-4			
		2,3,7,8-HpCDD 2,3,7,8-HpCDF	38998-75-3			
		2,3,7,8-HxCDD	34465-46-8			
		2,3,7,8-HxCDF	55684-94-1			
		OCDD	3268-87-9			
		OCDF	39001-02-0			
		2,3,7,8-PeCDD	36088-22-9			
		1,2,3,7,8-PeCDF	57117-41-6			
		2,3,4,7,8-PeCDF	57117-31-4			
		2,3,7,8-TCDD	1746-01-6			
		2,3,7,8-TCDF	51207-31-9			
		Total Carcinogenic PAHs ³	50-32-8			
		Benz(a)anthracene ³	56-55-3			
		Benzo(a)pyrene ³	50-32-8			
		Benzo(b)fluoranthene ³	205-99-2			
		Benzo(k)fluoranthene ³	207-08-9			
		Chrysene ³	218-01-9			
		Dibenz(a,h)anthracene ³	53-70-3			
		Indeno(1,2,3-cd)pyrene ³	193-39-5			
		Total PCBs Aroclor 1016	1336-36-3			
		Aroclor 1016 Aroclor 1221	12674-11-2			
		Aroclor 1221 Aroclor 1232	11104-28-2 11141-16-5			
		Aroclor 1242	53469-21-9			
		Aroclor 1248	12672-29-6			
		Aroclor 1254	11097-69-1			
		Aroclor 1260	11096-82-5			
		Vinyl chloride ³	75-01-4			
		Xylenes (Mixture)	1330-20-7			
		p-Xylene	106-42-3			
		m-Xylene	108-38-3			
		o-Xylene	95-47-6			
		•				

¹ This list of chemicals, compounds, and radionuclides was compiled from COPCs retained as COCs in baseline risk assessments performed at PGDP between 1990 and 2013 (i.e., DOE 1996a; DOE 1996b; DOE 1999a; DOE 1999b; DOE 2000a; DOE 2001; DOE 2005; DOE 2008; DOE 2010; DOE 2013).

² List may be added to during project scoping based on additional information.

³ Chemical is considered a mutagen (see Table B.6a).

Radiological dose-based concentrations for significant COPCs are presented in Tables A.8 through A.11 in Appendix A. ALs and NALs for contaminants in soil based on radiological dose limits are derived by following EPA guidance (EPA 2000a) and are used for radiological dose assessments at PGDP.

Table A.1 presents risk-based ALs for contaminants in soil and sediment; Table A.2 presents risk-based ALs for contaminants in groundwater; Table A.3 presents risk-based ALs for contaminants in surface water; Table A.4 presents risk-based NALs for contaminants in groundwater; Table A.6 presents risk-based NALs for contaminants in surface water; Table A.7 presents risk-based NALs for contaminants in soil that are protective of groundwater drawn from the Regional Gravel Aquifer (RGA) immediately adjacent to a contaminated area; Table A.8 presents radiological dose-based levels for radionuclide contaminants in soil and sediment; Table A.9 presents radiological dose-based levels for radionuclide contaminants in groundwater; Table A.10 presents radiological dose-based levels for radionuclide contaminants in surface water; and Table A.11 presents radiological dose-based levels for radionuclide contaminants in soil that are protective of groundwater drawn from the RGA immediately adjacent to a contaminants in soil that are protective of groundwater drawn from the RGA immediately adjacent to a contaminated area. Methods used to develop the risk-based and radiological dose-based screening values are presented in Appendix B of this document.

Screening values for the residential scenario are used in data screening to develop the list of COPCs in a baseline human health risk assessment (see Section 3.3.3.2 for additional information). Additional scenarios/receptors are used to determine early action screening.

All groundwater screening is performed using the resident. Of the two receptors (i.e., child and adult), use of the child screening values is more protective for noncarcinogens while the use of lifetime resident screening values is more protective for carcinogens. For chemicals that have screening levels based on both carcinogenic and noncarcinogenic effects, the lower of these two values should be used for screening. Note that values for soil deemed protective of groundwater also are available and are based on the resident only.

The surface water screening values selected are a location-specific decision. For all areas along effluent ditches or along creeks carrying effluent, the industrial worker screening values are appropriate. Additionally, at areas outside the industrialized areas, use of the recreator values is appropriate. Of the three recreator values available, the child recreator values are the smallest and most protective of human health. Note that two different sets of recreator values are available; these are a set for screening shallow water courses under a wading scenario and a set for screening deeper water courses under a swimming scenario. While which of these two recreator screening values to use is a location-specific decision, general guidance should be to use the wading values for most areas. If exposure by a resident to surface water is of concern, use of the recreator values is appropriate, because rates of contact for the recreator were selected assuming that the individual would be a local resident.

Determining which soil or sediment screening value is appropriate is a location-specific decision. For all locations inside the industrialized area at PGDP where surface soil contamination is of concern, use of the industrial worker risk-based screening values is appropriate. [Surface soil is defined as 0–1 ft below ground surface (bgs) (EPA 2018a)]. However, if the scenario involves outdoor maintenance type activities, the outdoor worker risk-based screening values also should be considered. For locations inside the industrialized area at PGDP where contact with surface soil and subsurface soil is of concern (i.e., soil from 0–10 ft bgs or 0–16 ft bgs, as appropriate), use of the excavation worker risk-based screening values is appropriate. For locations, outside the industrialized area where surface soil contamination is of concern,

industrialized area will be developed on a project-specific basis.

⁷ The size of the industrialized area at PGDP has varied historically and continues to change as operational changes are made at the site. Currently, the Paducah Site is situated on approximately 3,556 acres, with approximately 1,450 acres utilized for site operations; approximately 133 acres in acquired easements; and 1,973 acres licensed to the Commonwealth of Kentucky as part of the West Kentucky Wildlife Management Area. An evaluation of whether a location falls within or previously fell within an

screening using the recreator and/or resident risk-based screening values is appropriate. As with the surface water values, the child resident risk-based screening values are the most protective (in terms of protecting human health). Generally, the recreator risk-based screening values are more appropriate for areas along ditches and creeks (i.e., for bank soils), and the resident risk-based screening values are more appropriate for grassy fields. Finally, the outdoor worker risk-based screening values also can be considered for contact with soil in locations outside the industrialized area (i.e., soil from the 0–10 ft bgs or the 0–16 ft bgs, as appropriate) if this scenario is appropriate for the locations considered. If screening needs to consider shorter-term exposures to both surface and subsurface soil in locations outside the industrialized area, excavation worker screening values can be used.

A comparison of analyte concentrations detected in soil and groundwater samples to analyte concentrations detected in background samples will be performed as part of the development of the list of COPCs as shown in Figures 2.3 and 2.4. The values used to represent background are presented in Appendix A. Appendix E also contains a discussion of the derivation of the background values. Only surface soil (0–1 ft bgs), subsurface soil (1–16 ft bgs), and groundwater drawn from the RGA and McNairy Formation will be included in comparison with background concentrations because background values are available only for these media at PGDP (DOE 2000b). The RGA is the lateral flow system that constitutes the shallow Class II groundwater aquifer beneath PGDP and contiguous lands to the north. The McNairy formation flow system is below the RGA.

EPA OLEM Directive 9200.2-141 A addresses anthropogenic background: "Similarly, for anthropogenic contaminant concentrations, the CERCLA program normally does not set cleanup levels below anthropogenic background concentrations (US EPA, 1996; US EPA, 1997b; US EPA, 2000c). The reasons for this approach include cost-effectiveness, technical practicability, and the potential for recontamination of remediated areas by surrounding areas with elevated background concentrations. In cases where areawide contamination may pose risks, but is beyond the authority provided under CERCLA, EPA may be able to help identify other programs or regulatory authorities that are able to address the sources of areawide contamination, particularly anthropogenic (US EPA, 1996; US EPA, 1997b; US EPA, 2000c). In some cases, as part of a response to address CERCLA releases of hazardous substances, pollutants, and contaminants, EPA may also address some of the background contamination that is present on a site due to area-wide contamination." (EPA 2018b). Anthropogenic background levels for organic chemicals [e.g., polycyclic aromatic hydrocarbons (PAHs) in soil; chloroform in groundwater] may be developed on a project-specific basis. Appendix E.7, PAH Contamination and Establishment of Remedial Goals, establishes the concept of anthropogenic background levels that are not specifically associated with siterelated activities. Another example of establishing anthropogenic background levels on a project-specific basis is chloroform being measured in air samples during the Plant Industrial Vapor Intrusion Project (DOE 2021a). The project concluded that, because there were no known environmental sources of

chloroform present near the sampling locations, the source of chloroform in the air samples was likely from an anthropogenic background source (i.e., the interaction of chlorine in leaking potable water containing chlorine and organic matter resulted in the production and off-gassing of chloroform); however, chloroform could be related to historical releases and the disposal of industrial solvents.

Background concentrations for chemicals and radionuclides in soil and RGA and McNairy Formation groundwater to be used during site-scoping activities are presented in Tables A.12 and A.13, respectively. In the background screen for soil and groundwater, the maximum detected concentration of the COPCs will be compared to the values presented in Tables A.12 and A.13.

PEGASIS

The Portsmouth/Paducah Project Office Environmental Geographic Analytical Spatial Information System (PEGASIS) originally was pioneered by the Kentucky Research Consortium for Energy and the Environment, PEGASIS provides dynamic mapping and environmental monitoring data display for PGDP.

PEGASIS is available online at the following link:

https://pegasis.pad.pppo.gov/

Analytes for which the maximum detected concentrations [or maximum activity concentrations for radionuclides with reported values greater than their minimum detectable concentration (MDC)] is less than background will be removed from the data set used in the risk assessment. The background values for soil presented in Table A.12 represent upper tolerance limits of background except as noted in the table footnotes. Additional comparisons of the maximum detected analyte concentration or maximum activity concentration for radionuclides with the range of background values also may be conducted in the uncertainty section of the risk assessment (discussed in Section 3.3.7) to further evaluate if a COPC represents a site contaminant. Because surface water and sediment are transient media in which concentrations and activities can change rapidly, PGDP does not plan to develop surface water and sediment background. Currently, a comparison of the full range of concentrations and activities in upstream versus downstream samples is to be used to determine if a unit or area is releasing contaminants to the environment. Additionally, as part of the analysis, the data adequacy at both the upgradient location and potentially contaminated site must be considered.

To perform the screening analyses during site scoping, available data must be deemed sufficient to determine the potential contamination at a site. Data used during site scoping will be evaluated using the systematic approach presented in Figure 2.2 to ensure that risk analyses employ data of known quality and that the appropriate quantities and types of data are acquired. This systematic approach also is used to evaluate data during an RI, as discussed in Section 3. Detailed discussions related to data quality/data usability review are provided in Section 3.3.3.1.

In presenting the results of risk-based site scoping analyses, several tables should be prepared using a format that allows for easy identification of those chemicals, compounds, and radioisotopes with the potential to contribute to unacceptable levels of risk. If a radiological dose analysis is conducted, similar tables should be prepared to present the results of the radiological dose-based site scoping analysis. To complete the risk-based screening analyses for site scoping, tables will be prepared for soil and sediment, groundwater, and surface water screening. For soil and sediment, up to four tables will be prepared using the risk-based screening levels. These tables offer comparisons among the following:

- Maximum detected concentrations and ALs,
- Maximum detected concentrations and NALs,
- Maximum detected concentrations and levels deemed protective of groundwater, and
- Maximum detected concentrations and established background values for naturally occurring inorganics and radionuclides.

For both groundwater and surface water, two tables will be prepared using the risk-based screening levels. These tables offer comparisons between the following:

- Maximum detected concentrations and ALs and
- Maximum detected concentration and NALs.

In addition, summary tables providing the following information will be prepared for each medium:

- Lists of chemicals and radionuclides analyzed for but never detected;
- A presentation of summary statistics, including a comparison of detected analytes with background;
- Lists of sampling stations that contain a contaminant at a concentration greater than the action screening level; and

• Lists of sampling stations that contain a contaminant at a concentration greater than the no action screening level.

2.1 ANALYSES SUPPORTING ACTION PRIOR TO RI/FS

As discussed in the FFA, interim actions are required at those sites that pose an imminent risk or hazard to human health and the environment. Generally, sites requiring an interim remedial or removal action are those at which contamination with a single or small number of analytes presents a total carcinogenic risk greater than 1×10^{-4} or a systemic toxicity value (i.e., hazard index or HI) greater than one and for which the risk analyses indicate that exposure is occurring under current use patterns. For these sites, the screening risk analyses will be limited to that described here because additional analyses will slow response time; however, to complete later decision documents, estimates of cumulative risk will be developed. [Note: The exact decision point for interim action is a project-specific decision. The values included here are for illustration only. For example, it is possible that a site is a yard that contains source material that might present a principal threat. At such sites, the scoping analyses may not include a risk-based screen. Additionally, note that risks posed to nonhuman receptors (e.g., ecological risk) may call for an interim remedial or removal action even when risks to humans are negligible.] To derive these estimates of cumulative risk, the methods in Equations 1, 2, 3, and 4 will be used. Methods to derive radiological dose estimates are similar and are not presented in the equations. For example, when deriving radiological dose estimates, the dose-based PRG derived using a target dose of 1 mrem/year would replace the "Cancer PRG," and "Target Risk" would be replaced with a target dose of 1 mrem/year.8

Analyte-specific Risk =
$$\frac{MAX}{Cancer PRG} \times Target Risk$$
 Eq. 1

where: MAX = Maximum detected concentration in a medium.

Cancer PRG = The medium-specific risk-based no action screening value for the analyte.

Target Risk = The target risk upon which the risk-based PRG calculation was based (1×10^{-6}) .

NOTE: This relationship is not applicable to non-linear based PRGs [e.g., some vapor intrusion screening levels (VISLs)].

where: Analyte-specific risk is the result from Eq. 1.

Analyte-specific Hazard = $\frac{MAX}{Hazard PRG} \times Target Hazard$ Eq. 3

where: MAX = Maximum detected concentration in a medium.

Hazard PRG = The medium-specific risk-based no action screening value for the analyte.

Target Hazard = The target hazard upon which the risk-based PRG calculation was based (0.1).

⁸ The radiation target dose 1 mrem/year is not a DOE, EPA, or Kentucky standard. Also, as with risk-based PRGs for chemicals and radionuclides, dose-based PRGs are used in project screening only and should not be considered clean-up values.

2-11

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Eq. 4

where: Analyte-specific Hazard is the result from Eq. 3.

[Note: When performing these calculations, total risk and hazard estimates will be developed within medium for only the scenario appropriate to the unit's or area's location and use because the reasonably anticipated future land use at a site is significant in defining source material as a principal or low-level threat waste (EPA 1991a). A total risk (or hazard) over all media may be estimated if exposure to contaminants in multiple media may occur. Also, when summarizing this information, the analytes driving the medium-specific total risk and hazard and the major uncertainties in the estimate will be reported, and a total risk or hazard estimate over all media may be reported if this is deemed appropriate.]

The results provided by these analyses may not be sufficient for documentation of final actions, and additional risk assessment and risk evaluation may be needed to meet reporting requirements. Items not provided by these analyses include the following:

- The identification of use scenarios of concern, including consideration of sensitive subpopulations;
- The identification of pathways of concern (POCs);
- Consideration of risks due to the transformation, degradation, or migration of contamination (although a
 comparison of analyte concentrations in soil to screening values protective of groundwater provides this
 in part); and
- An analysis of uncertainties, including the effect of uncertainties on the resulting risk estimates.

2.2 ANALYSES SUPPORTING NO FURTHER ACTION DECISIONS

No further action can be selected for those sites where it can be demonstrated that no contamination is present that exceeds NALs [i.e., risks are *de minimis* (see Figures 2.1–2.5)] or ARARs. (Note: Non-risk issues also must be considered in making this decision. At some sites without unacceptable risk, a no further action decision may not be appropriate because of non-risk concerns.)

In calculating the risk estimate for this decision, the tables discussed earlier and the equations presented earlier will be used. In summarizing this information, the estimated total risk and hazard from all contaminants under the appropriate use will be reported, and the future risk or hazard posed by contaminant transformation, degradation, and migration will be considered qualitatively. In addition, the uncertainties associated with the screening comparison will be discussed, and the effect of these uncertainties on the total risk and hazard estimates for each scenario will be described. Note: As part of this screening analysis, the total risk or hazard over all media will be presented and discussed to ensure that a no further action decision is appropriate.

2.3 ANALYSES USED TO PRIORITIZE FURTHER INVESTIGATIONS

Remedial activities at PGDP are prioritized to ensure that funds allocated to PGDP for remedial actions are directed toward those units or areas that pose the greatest risk to human health and the environment. This prioritization will ensure that these actions provide the maximum benefits in risk reduction. When necessary,

risk and hazard estimates for prioritization will be calculated using the tables and equations presented earlier. When summarizing this information, the estimated total risk and hazard from all contaminants under both industrial and residential use will be reported, and the potential future radiological doses and risks posed by contaminant transformation, degradation, and migration will be considered qualitatively. In addition, the uncertainties associated with the screening comparison will be discussed, and the effect of these uncertainties on the total risk and hazard estimates for each receptor group will be estimated qualitatively.



3. RISK ANALYSES DURING THE REMEDIAL INVESTIGATION

At PGDP, risk analyses occur at three points during the RI of sites: during the preparation of the RI work plan (and some sampling and analysis plans); following implementation of the initial round of work described in the RI work plan (if needed to plan contingency sampling); and during the preparation of the RI report. Analyses occurring at each of these points are discussed in the following sections. (Note that radiological dose assessments are not specifically described in the following. Generally, if a radiological dose assessment is provided, it will be presented in the same format as the risk assessment.)

3.1 ANALYSES DURING WORK PLAN DEVELOPMENT AND IMPLEMENTATION (SCREENING RISK ASSESSMENTS)

As noted in Section 2.4, the screening analyses performed during the site scoping can be used directly in work plan development to reduce the cost of subsequent RI/FS activities. This section discusses the screening analyses that will be performed as part of work plan development and describes the material that will appear in work plans and sampling and analysis plans. (Note: In the following material, "work plan" is used generically for work plans and for those sampling and analysis plans in which risk screening is of use.)

Generally, in work plans, the majority of the risk-related information will appear as part of the initial evaluation. In the work plan's initial evaluation, the scope, objectives, and methods for the baseline risk assessment will be related; preliminary conceptual site models will be presented; laboratory analytical (or quantitation) limits will be discussed relative to no action screening levels developed specifically for PGDP (i.e., risk-based PRGs in Appendix A); and a preliminary list of COPCs (preliminary COPCs) will be identified. Risk-related information also will appear in the introduction, site characterization summary, and alternatives development description contained in most work plans.

3.1.1 Analyses Appearing in the Introduction of the Integrated RI/FS Work Plan

In the introductory chapter of work plans, the requirements for risk assessments and analyses will be used to help develop the data quality objectives (DQOs) for the RI. DQOs are qualitative and quantitative criteria used to establish requirements for sample collection and analysis and are based on the needs and intended uses of the data. As a primary user of RI data, the consideration of risk analyses is integral to this process.

Development of DQOs follows a series of steps. The seven steps in the process are shown in a flowchart found in EPA QA/G-4, *Guidance on Systematic Planning Using the Data Quality Objectives Process* (EPA 2006a). Similar steps are found in U.S. Department of Energy (DOE) guidance, *Institutionalizing the Data Quality Objectives Process for EM's Environmental Data Collection Activities* (DOE 1994). The purpose and goal of each step are described in the text in EPA QA/G-4, accompanying the flowchart. EPA QA/G-4 also includes a summary of key elements that also may be of use in developing DQOs for specific investigations. The role of risk assessment within each of these steps is briefly discussed in the remainder of this section.

During Step 1, State the Problem, of the DQO process, risk analyses will be used to identify qualitatively the preliminary COPCs, receptors that may be exposed to contaminants, locations at which exposure may occur, and pathways by which contaminants may reach these locations. This information will be used to develop the conceptual site model against which new data collected as part of the RI can be compared. An example conceptual site model is presented in Figure 3.1.

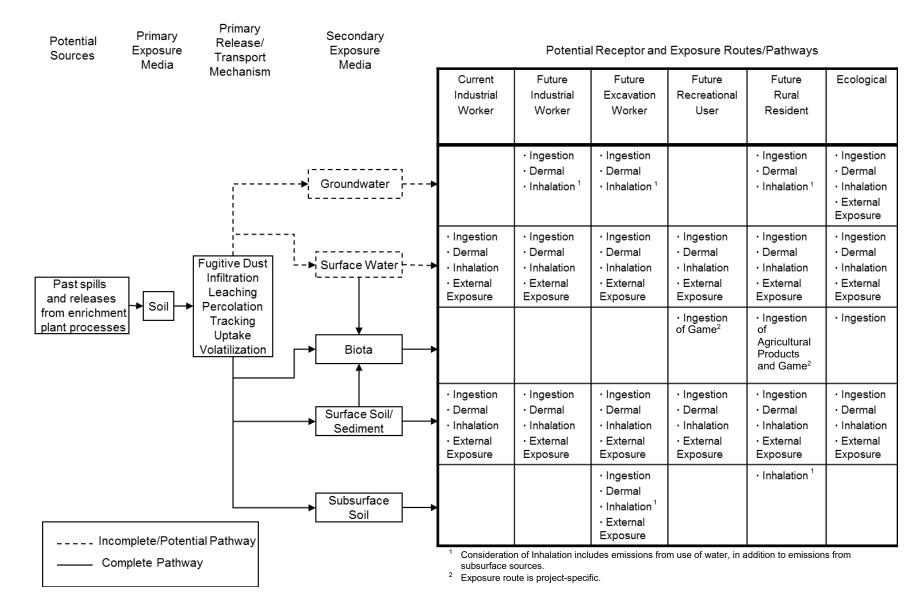


Figure 3.1. Example Risk-Based Conceptual Site Model

Risk analyses also will be used during Step 1 of the DQO process to ensure that the risk management issues are addressed during the investigation. For example, in the approved sampling and analysis plan for Solid Waste Management Unit (SWMU) 2 of Waste Area Grouping 22 (DOE 1996a), the problem is stated as follows:

In the past, uranium and multiple COCs were disposed of at SWMU 2. These contaminants have been shown by previous work to be migrating (vertically and horizontally) from the waste cells and show the potential for subsurface migration from the SWMU to the RGA at concentrations or activities that may pose risk to human health and the environment....

Risk analyses will be used during Step 2, Identify the Goals of the Study, of the DQO process to clearly pose questions that must be addressed during the RI. Generally, questions developed during Step 2 of the process will be related to the contamination concentrations that may remain at or migrate from a site and not pose unacceptable risk. Inputs to these questions include contaminant fate and transport and activity patterns of current and future receptors. For example, in the SWMU 2 sampling and analysis plan (DOE 1996a), primary questions related to risk assessment and risk management included the following:

- Will the contaminants migrate (and how) to the RGA at unacceptable concentrations?
- Is there lateral/vertical contaminant movement in the Upper Continental Recharge System (UCRS)?
- What are the chemical characteristics of the waste?

Risk analyses will be used during Step 3, Identify Information Inputs, of the DQO process to establish the preliminary remedial action objectives (RAOs) that must be achieved to mitigate risk to human health and the environment and to provide information useful in determining which alternatives may achieve these objectives. RAOs are criteria used in the FS to aid in the alternative development and selection process. They are site-specific goals that establish the primary objectives and extent of cleanup required by a CERCLA remediation (EPA 1988) and consider COCs, media of concern (MOCs), and potential exposure pathways. The screening levels presented in Section 2 are concentration goals that will make up a portion of the preliminary RAOs for each project. For all investigations at PGDP, the basis of this portion of the human health RAO is to prevent exposure to contaminated media that results in a cumulative (or total) ELCR greater than 1 × 10⁻⁶ or a cumulative (or total) HI greater than or equal to one. This generalized RAO will be enhanced on a project-specific basis as needed (e.g., to include radiological dose concerns).

Risk analyses will be used during Step 4, Define the Boundaries of the Study, of the DQO process to aid in the determination of the spatial and temporal boundaries within which samples must be collected or to which contaminant concentrations must be modeled. Risk analyses will be used to identify spatial boundaries by delimiting the locations both at a SWMU and away from the SWMU at which exposure to contaminants may occur (i.e., exposure points). Risk analyses will be used to identify temporal boundaries by delineating the present and future receptors that may be exposed to contamination and the periods during which these receptors potentially may be present at the exposure points. This information will be used, in turn, to determine the modeling needs for the RI.

Risk analyses will be used during both Steps 3 and 5, Develop the Analytic Approach to the Decision, to set the risk-based limits inherent in these rules and to identify the data required to determine if these limits may be exceeded, consistent with Section XII of the Paducah FFA (EPA 1998a). A primary decision rule that will be included in all work plans for PGDP will note that action must be considered if the risk or hazard posed by contamination at or migrating from a site exceeds allowable limits of an ELCR greater than 1×10^{-6} or HI greater than or equal to one. For example, in the SWMU 2 sampling and analysis plan (DOE 1996a), the leading decision rule (D1) is as follows:

If any of the constituents shown in Table 5.2 are migrating or could migrate (based on RESRAD for uranium and technetium-99 (99 Tc) and best available 2- or 3-D model for other constituents) from the burial pits, soil matrix, and/or UCRS to the RGA in the future and are found to pose a risk greater than 1×10^{-6} (excess lifetime cancer) or an HI = 1 (noncancer), then an action to control the migration will be evaluated.

Similarly, the following inputs necessary to make this decision are common to all investigations:

- Chemical-specific exposure point concentrations (EPCs) in environmental media, including contaminant concentrations in waste;
- Land-use assumptions (i.e., which scenarios need to be considered);
- Exposure pathways and exposure routes for all current and potential future receptors;
- Exposure units for the investigated area;
- Modeling parameters;
- Risk estimates for each receptor, including sensitive subpopulations, if applicable.

Risk analyses will be used in Step 6, Specify Performance or Acceptance Criteria, by providing the risk-based goals and contaminant concentrations and activities related to these goals that can be used either quantitatively or qualitatively to set decision error limits. As noted previously, consistent with the PGDP FFA, the risk-based goals to be used in all investigations are 1×10^{-6} for ELCR and 1 for HI. For a radiological dose assessment done to provide information for risk managers, the radiological dose-based goal is 1 mrem/year. The concentrations and activities related to these goals are the PRGs presented as the NALs in Section 2.

Risk analyses will be used in Step 7, Develop the Plan for Obtaining Data, to ensure that the sampling strategy proposed for all investigations meets the minimum requirements needed to achieve answers to the risk-related decision rules. To ensure that this is achieved, all sampling proposed as part of all investigations will be critically reviewed against the needs established under the decision rules for the investigation. Sampling that does not provide information useful to answering risk-related decisions will be justified on another basis.

3.1.2 Analyses Appearing in Prior Characterization Chapter of the Integrated RI/FS Work Plan

In the prior characterization chapter of work plans, results of previous risk evaluations performed for the site under investigation or related to the site will be summarized. Generally, these summaries will consist of results from evaluations performed during the Phases I and II Site Investigations (CH2M HILL 1991 and 1992) or baseline risk assessments and screening analyses performed to support earlier decisions at or near the site, such as prioritization activities.

In presenting the information from previous evaluations, **no attempt will be made to correct any errors or update any values contained in the earlier reports**. All information contained in the earlier report will be presented without change; however, any errors or uncertainties affecting the results will be identified. Additionally, because in earlier baseline risk assessments, results were not summarized in a consistent format, an attempt will be made to present the results taken from these earlier reports in two-way tables. [Note: The format for the two-way table is patterned after the format in Exhibits 8-2 and 8-3 of Risk Assessment Guidance for Superfund (RAGS), Part A, (EPA 1989a) and is consistent with the risk

characterization tables found in RAGS, Part D (EPA 1998b).] The exact format for tables presented in RAGS, Part D, is not used for the PGDP risk characterization tables because the Risk Assessment Working Group (RAWG) determined that the tables presented in this Risk Methods Document are adequate to meet the intent of RAGS, Part D. In addition, when summarizing the results of previous assessments, the scenarios, pathways, contaminants, and MOC for each unit or area under investigation will be listed, and major uncertainties affecting the risk assessment results will be noted.

An example of the format for the "two-way table," adapted from Table 5.78 of Appendix L.1 of the approved *Resource Conservation and Recovery Act Facility Investigation/Remedial Investigation Report for Waste Area Grouping 1 and 7 at Paducah Gaseous Diffusion Plant, Paducah, Kentucky* (DOE 1996b), is shown in Exhibit 3.1. The example table shown in the exhibit will be used to summarize risk assessment results because it allows easy identification of scenarios of concern (i.e., value in column entitled "Total Risk," COCs (i.e., values in the column entitled "Chemical-Specific Risk"), and POCs (i.e., values in the row entitled "Pathway Risk"). In addition, the chemicals and pathways driving total risk can be easily identified, and the risk related to exposure to each environmental medium can be easily derived (i.e., by summing the appropriate pathway totals). Finally, the blank cells in the table and the associated explanation for these blanks show where information was insufficient to allow risks to be characterized.

Exhibit 3.1. Example Two-Way Table for Presentation of Historical Risk Assessment Results

SWMU 136 Excess Lifetime Cancer Risks for Future Rural Resident							
Analyte	Ingestion of Groundwater	Dermal Contact with Groundwater		Ingestion of Soil		Chemical- specific Risk	Total Risk
Trichloroethene	2.30E-05	4.17E-06				8.35E-05	
Benzo(a)anthracene				8.78E-09		1.35E-06	
Benzo(a)pyrene				1.20E-07		1.83E-05	
			•				
			•			•	
	•		•			•	
Uranium-238				1.53E-09		3.05E-07	
Pathway Risk	2.32E-05	4.23E-06		1.72E-07			
Total Risk							1.10E-04

Note: The reasons for blank cells are discussed as part of the risk assessment/evaluation. Generally, blank cells will result from unavailable or inadequate data.

3.1.3 Analyses Appearing in Initial Evaluation Chapter of the Integrated RI/FS Work Plan

In the initial evaluation chapter of work plans, the methods to be used to complete the baseline risk assessment for the units or areas under investigation will be discussed, and a preliminary evaluation of historical information, including a comparison of concentrations and activities of analytes in environmental samples with risk-based screening values (e.g., NALs, ALs, chemical-specific ARARs, etc.) and a comparison of analytical limits with background concentrations, will be presented. This information will be used, in turn, to develop the field sampling plan contained in the work plan.

The description of the methods to be used to complete the baseline risk assessments for the units or areas under investigation will follow that presented in Section 3.3 of this document. Generally, this material will delineate clearly the scope and objectives of the baseline risk assessment and briefly describe the activities that will occur during the data evaluation (i.e., identification of COPCs); exposure assessment; toxicity assessment; risk characterization; and remedial goal option (RGO) development stages of the baseline human health risk assessment. This material also will summarize the results that will be obtained from each

stage of the baseline risk assessment. As part of this discussion, conceptual site models for each unit or area under investigation will be presented.

The preliminary evaluation of historical information presented in this chapter of the work plan will summarize the information presented in earlier chapters of the work plan and evaluate this information against the characterization and inventory of wastes, information status of key assessment factors, and release potential from contaminant sources. As part of the characterization and inventory of wastes, comparison tables similar to those discussed in Section 2 will be prepared. Because additional screening criteria may need to be considered, the comparison tables prepared as part of site scoping activities may not be able to be transferred directly to the work plan. An example of the comparison table that will be used in work plans to compare the PGDP screening PRGs to analytical results from soil (and sediment) and groundwater (and surface water) is shown in Exhibit 3.2.

Exhibit 3.2. Presentation of Screening Assessment Results in the RI Work Plan

	Soil (mg/kg or pCi/g)			Groundwater (µg/L or pCi/L)			
Analyte	Maximum ¹	PRG ²	Method Detection Limit ³	Maximum	PRG	MCL ⁴	Method Detection Limit
# 1							
# 2							
•	•	•	•	•	•	•	•
			•			•	
# N							

This value will be the maximum detected value for the medium reported in previous investigations. The qualifier codes attached to the value, if any, will be included with the value.

After completing the comparison table for each site, the analytes that previously were detected or are expected to be present and that have detection limits (MDCs for radionuclides) that exceed the PRGs will be reported. The analytes with detection limits exceeding PRGs will be reported because the quantitation limit (or method detection limit for chemicals or MDC for radionuclides) used for samples providing data for risk assessment should be less than those concentrations that may have an impact on human health or the environment. It is important to note that, although this evaluation may show that some quantitation limits exceed their respective screening criteria, this evaluation alone will not be used to establish the analytical quantitation limits for a project. The analytical limits will be established considering this information and factors such as site history and potential actions.

Material in the comparison tables also will be used to compile a list of preliminary COPCs for each unit or area under investigation. An analyte will be placed on this preliminary list if the concentration or activity concentration of the analyte at a unit or area exceeds one or more of the screening criteria. Note: Unless it can be shown that cross-media contamination is not present, the list of preliminary COPCs will be compiled over all media. If it can be demonstrated that cross-media contamination is not likely, then a list of preliminary COPCs will be compiled for each medium to be investigated during the project. These lists will provide risk managers with information that can be used in the initial selection and screening of alternatives. In addition, this list can be used to target the analyte list for the project to ensure that analytical costs are appropriate for the project.

² The risk-based PGDP screening PRG that appears in this table will be the no action child residential use PRGs taken from Appendix A.

³ This value will be the project-specific value reported in the Quality Assurance Project Plan of the work plan (or the appropriate chapter of sampling and analysis plans). For radionuclides, this column should have the heading "MDC" or "MDO" (minimum detectable quantity) and present MDCs from Multi-Agency Radiological Laboratory Analytical Protocols (MARLAP) guidance.

⁴ The maximum contaminant levels [i.e., maximum contaminant levels (MCLs)] are drinking water standards and will be taken from the most

recent information.

An example of the comparison table that will be used in work plans to compare background values to analytical results for inorganic chemicals and radionuclides in soil and groundwater is shown in Exhibit 3.3. (Note: as discussed earlier, background values are not available for sediment and surface water; therefore, a table comparing analytical results from sediment and surface water to background will not be presented.) This table will be used to justify the analyte list for the project. As with the list of preliminary COPCs, justification of the analyte list is important to ensure that analytical costs are appropriate for the project.

Exhibit 3.3. Presentation of Background Comparison in the RI Work Plan

		ata for S kg or pC			Groundwater Data for SWMU (μg/L or pCi/L) ³			Groundwater Background
Analyte	SWMU 1		SWMU N	Soil Background Concentration (mg/kg or pCi/g) ²	SWMU 1		SWMU N	Concentration (µg/L or pCi/L) ⁴
# 1								
# 2								
# N								

¹ This will be the maximum detected value for soil reported in previous investigations. The qualifier codes attached to the value, if any, will be included with the value.

3.1.4 Analyses Appearing in Remedial Alternatives Development Chapter of the Integrated RI/FS Work Plan

In the remedial alternatives development chapter of work plans, attention will be paid to the importance of risk reduction in remedial alternatives development and to the method to be used to measure risk reduction during the detailed analysis of remedial alternatives. For example, this chapter will note that remedial alternatives are developed to be protective of human health and the environment and that RAOs will consider COCs, POCs, and MOCs. In addition, this chapter will present the nine criteria used in the detailed analysis of alternatives under CERCLA. Most importantly, this chapter will discuss if a qualitative or quantitative detailed risk analysis of alternatives is anticipated and delineate the data that are required to support this risk analysis. (Determining whether a qualitative or quantitative risk analysis of alternatives is needed is important because additional data may need to be collected during the RI to support a quantitative analysis. Additional discussion concerning qualitative and quantitative risk analysis of alternatives is presented in Section 4.)

3.2 ANALYSES FOLLOWING COMPLETION OF THE INITIAL ROUND OF INVESTIGATION

Many RI work plans will contain a description of contingency sampling that may be used to address the uncertainties in environmental contaminant distribution expected to be encountered during the investigation. If this contingency sampling is to be collected as part of a phased investigation, then analyses may be used to allow the three FFA parties to discuss and agree if contingency soil (or sediment) sampling is necessary. In this case, a formal or informal report may be prepared after the completion of the initial round of sampling. In this report, results from the initial sampling and relevant historical sampling may be compared to human health screening criteria (i.e., PRGs) for the expected future use of the area and background concentrations of chemicals and radionuclides. To keep this presentation consistent with that

² The soil background concentration (or activity concentration) will be that presented in Appendix A or updated values.

³ This will be the maximum detected value for groundwater reported in previous investigations. The qualifier codes attached to the value, if any, will be included with the value.

⁴The groundwater background concentration (or activity concentration) will be that presented in Appendix A or updated values.

used in work plan development, this presentation will use comparison tables similar to those presented earlier. Because the extent of soil (or sediment) contamination needs to be considered, as well as the nature of contamination, tables considering the location of samples (horizontal and vertical), in addition to the tables considering the maximum detected analyte concentrations, will be prepared. A spatial plane view presentation of the data also should be provided.

The format of the comparison table to be used to determine if the nature of contamination in soil may pose an unacceptable risk or hazard is in Exhibit 3.4. In this table, the maximum detected concentration or activity concentration in all soil samples collected at a site is compared to the no action PRG for soil exposure for the expected future land use, the groundwater protection PRG, and the background concentration. This table will be used to refine the list of preliminary COPCs and the analytical list for contingency sampling. In this evaluation, an analyte will become a preliminary COPC if its concentration exceeds any PRG and the background concentration or activity concentration.

Exhibit 3.4. Presentation of Screening Assessment Results to Evaluate Nature of Contamination in Soil after the Initial Round of Sampling

	Soil (mg/kg or pCi/g)					
Analyte	. 1		Groundwater Protection PRG ³	Background ⁴		
# 1						
# 2						
	•	•		•		
•						
# N						

¹ This value will be the maximum detected value for soil reported in the current and relevant previous investigations. The qualifier codes attached to the value, if any, will be included with the value.

The format of the comparison table to be used to determine if the nature of contamination in sediment may pose an unacceptable cancer risk or hazard will be similar to that in Exhibit 3.4; however, for the sediment table, neither the groundwater protection PRG nor the background concentration will appear. The groundwater protection PRG will not be included because migration of contaminants from sediment to groundwater is not expected to be a significant migratory pathway. Background concentrations of chemicals and radionuclides will not be included because these data do not exist for sediment. As with the soil table, the sediment table will be used to refine the list of preliminary COPCs and the analytical list for contingency sampling. In this evaluation, an analyte will become a preliminary COPC if its concentration or activity concentration exceeds any risk-based screening criterion.

The format of the comparison table to be used to evaluate the adequacy of initial sampling in delimiting the extent of contamination in surface soil is in Exhibit 3.5. In this table, the analyte concentrations or activities in surface soil samples collected along migration routes or at the periphery of a site are compared to the no action PRG for soil for the expected future land use and the background concentration or activity concentration. Note that the groundwater protection soil PRG is not used in this comparison because that evaluation is performed as part of the subsurface soil evaluation. Generally, surface sampling will be deemed adequate if analyte concentrations and activities in samples collected along migration routes do not exceed both the no action PRGs for soil and background concentrations. In deciding if sampling has adequately determined the extent of contamination, additional factors such as historical information will be considered.

² The PRG will be the no action PRGs for exposure to soil for the appropriate future use taken from Appendix A. If residential use PRGs are used, then the child NAL should be used.

³ The groundwater protection PRG will be the no action PRGs taken from Appendix A. Note: This PRG is protective of groundwater that may be used in the home. A PRG for protection of groundwater used industrially is not relevant to this screening assessment.

⁴ The soil background concentration (or activity concentration) will be that presented in Appendix A or the most recent updated study/report.

Exhibit 3.5. Presentation of Screening Assessment Results to Evaluate Extent of Contamination in Surface Soil after the Initial Round of Sampling

	Soil (mg/kg or pCi/g)				
Analyte	Maximum ¹	PRG ²	Background ³		
# 1					
# 2					
		•			
# N					

This value will be the maximum detected value for soil reported in a sample collected along migration routes or at the periphery of the unit or area in the current investigation. The qualifier codes attached to the value, if any, will be included with the value.

The format of the comparison table to be used to evaluate the adequacy of initial sampling in delimiting the extent of contamination in sediment will be similar to that used for soil (Exhibit 3.5); however, the background concentration or activity concentration will not appear in the sediment table because background values for sediment do not exist. The evaluation of this table will be the same as for soil.

The format of the comparison table to be used to evaluate the adequacy of initial sampling in delimiting the extent of contamination in subsurface soil is in Exhibit 3.6. In this table, the analyte concentrations or activities in subsurface soil samples collected at the periphery of the area under investigation will be compared to the groundwater protection PRGs and background concentrations of chemicals and radionuclides. Note: The no action PRGs for soil are not in this table because these criteria are for contact with contaminated soil, and contact with subsurface soil is not expected. Generally, subsurface sampling will be deemed adequate if analyte concentrations and activities in samples collected at the periphery of the unit or area under investigation do not exceed both the groundwater protection PRGs and background concentrations. In deciding if sampling has adequately determined the extent of contamination, additional factors such as historical information will be considered.

Analyses to evaluate groundwater and surface water sampling in determining the nature and extent of contamination in groundwater and surface water will be similar to those for soil. The format of the comparison table to be used to determine if the nature of contamination in groundwater may pose an unacceptable excess cancer risk or systemic toxicity is in Exhibit 3.7. In this table, the maximum detected concentration or activity concentration in all groundwater samples collected at the site will be compared to the no action PRG for residential use of groundwater, the MCL, and the background concentration or activity concentration. This table will be used to refine the list of preliminary COPCs and the analytical list for contingency sampling. In this evaluation, an analyte will become a preliminary COPC if its concentration exceeds any screening criterion and the background concentration or activity concentration. Comparisons to MCLs will not be used to identify COPCs, but will be provided for information only.

The table used to determine if contamination in surface water may pose an unacceptable cancer risk or hazard will be similar to that in Exhibit 3.7; however, background concentrations of chemicals and radionuclides will not appear in the surface water table because background data do not exist for surface water. The evaluation of this table will match that for groundwater.

² The PRG will be the no action PRGs for the appropriate future use taken from Appendix A.

³ The soil background concentration (or activity concentration) will be that presented in Appendix A or the most recent updated study/report.

Exhibit 3.6. Presentation of Screening Assessment Results to Evaluate Extent of Contamination in Subsurface Soil after the Initial Round of Sampling

Soil (mg/kg or pCi/g)					
Analyte	Maximum ¹	Groundwater Protection PRG ²	Background ³		
# 1					
# 2					
	•		•		
# N					

¹ This value will be the maximum detected value or maximum activity concentration for radionuclides for subsurface soil reported in a sample collected at the periphery of the unit or area in the current investigation. The qualifier codes attached to the value, if any, will be included with the value.

Exhibit 3.7. Presentation of Screening Assessment Results to Evaluate Nature of Contamination in Groundwater after the Initial Round of Sampling

	Groundwater (μg/L or pCi/L)					
Analyte	Maximum ¹ PRG ²		Maximum Contaminant Level ³	Background ⁴		
# 1						
# 2						
•	•	•	•	•		
# N						

¹ This value will be the maximum detected value for groundwater reported in all samples collected around the unit or area during the current and relevant previous investigations. The qualifier codes attached to the value, if any, will be included with the value.

For all investigations except the final RI of the Groundwater Operable Unit, there will be limited evaluation of the extent of existing groundwater contamination during the evaluation of the initial round of sampling. Currently, only the extent of dense nonaqueous-phase liquid contamination (i.e., secondary sources) is addressed during the investigation of the individual units and areas. The method used for the detection of these secondary sources does not rely on risk analysis and will not be discussed here. For the Groundwater Operable Unit investigation, the comparison table used to examine the adequacy of sampling in determining the extent of groundwater contamination will be similar to that in Exhibit 3.7; however, in this evaluation, a table will be prepared for each groundwater sampling location along the suspected periphery of the contaminant plumes. In each of these tables, the maximum detected analyte concentrations and activities will be compared to the no action residential use PRGs, MCLs, and background concentrations. Generally, groundwater sampling will be deemed adequate to determine the extent of contamination if analyte concentrations and activities in samples collected along periphery of the suspected groundwater contaminant plumes do not exceed screening criteria and background concentrations. In deciding if sampling has adequately determined the extent of contamination, additional factors such as historical information will be considered.

The table to be used to determine the adequacy of sampling in determining the extent of surface water contamination also will be similar to that in Exhibit 3.7. As noted earlier, this table will not contain background concentrations of chemicals and radionuclides because background values are not available for

² These values are taken from Appendix A.

³ The soil background concentration (or activity concentration) will be that presented in Appendix A or the most recent updated study/report.

² The PRG will be the no action PRGs in Appendix A for the child.

³ The MCL will be taken from Appendix A or the most recent update.

⁴The groundwater background concentration (or activity concentration) will be that presented in Appendix A or the most recent update.

surface water. Generally, surface water sampling will be deemed adequate to determine the extent of contamination if analyte concentrations and activities in samples collected downstream of a unit or area do not exceed screening criteria. In deciding if sampling has adequately determined the extent of contamination, additional factors such as historical information will be considered.

3.3 ANALYSES FOR THE RI REPORT (BASELINE RISK ASSESSMENTS)

Baseline risk assessments will be prepared to support final actions at PGDP. To ensure consistency among assessments and conformity with agreements reached between DOE and regulatory agencies, all assessments will contain either the material described in succeeding sections or an explanation stating why the material is not presented. Material described herein but not relevant to a particular assessment will be noted in the assessment. The following are specific objectives of the remedial action process to be addressed in this section:

- Delineate the methods PGDP will use in the evaluation, determination, and documentation of baseline risks to human health and the environment at a site; and
- Describe the methods PGDP will use to determine the concentrations and activities of analytes that can remain on-site and still be adequately protective of human health and the environment both on-site and off-site.

In the following sections, the presentation follows the outline to be used in baseline human health risk assessments. Data evaluation methods are discussed in Section 3.3.3, exposure assessment methods are presented in Section 3.3.4, toxicity assessment methods are described in Section 3.3.5, risk characterization methods are delineated in Section 3.3.6, uncertainty in the risk assessment is discussed in Section 3.3.7, and RGO derivation methods are discussed in Section 3.3.8. In addition, the sources used to prepare this material are listed in Section 3.3.1, and general issues are considered in Section 3.3.2.

[Note: The methods for the baseline ecological risk assessment are not considered here. They are described in the companion Ecological Risk Methods Document. Additionally, methods to be used for radiological dose assessment are not presented in detail. The methods for radiological dose assessment generally should follow those used for baseline risk assessments.]

3.3.1 Guidance Documents

The methods discussed in the following sections are consistent with current EPA Region 4 and headquarters risk assessment guidance documents, the Commonwealth of Kentucky Department for Environmental Protection (KDEP) risk assessment guidance, and applicable DOE Orders. In addition, these methods are consistent with agreements reached during meetings among DOE, EPA Region 4, and KDEP risk assessment personnel (DOE 1996c; EPA 1996a; KDEP 1996; RAWG 2000a, RAWG 2000b, 2000c, 2000d, 2000e, 2000f, 2000g, 2007a, 2007b, 2007c, 2012a, and 2012b; Appendix E of DOE 2017; DOE 2018; DOE 2019; DOE 2020, DOE 2021b and DOE 2022b) and strategies and methods developed for human health risk assessments for use at other DOE sites located in EPA Region 4 (e.g., K-25, X-10, and Y-12 in Oak Ridge, Tennessee). Some of these methods are different from those used in earlier risk assessments. References for methods and approach should refer to this methods document and/or the original guidance documents instead of other site-specific project documents to avoid inappropriate references. Many of the documents and other materials used in developing the methods are listed chronologically in the following sections. If newer versions of the listed reference are available, the newer version should be used in place of the specific version listed in the following sections.

3.3.1.1 EPA guidance documents and materials

- Risk Assessment Guidance for Superfund, Volume I: Human Health Evaluation Manual, Parts A, B, C, D, E, and F (EPA 1989a, 1991b, 1991c, 1998b, 2004a, and 2009, respectively) (RAGS, Parts A, B, C, D, E, and F, respectively)
- Exposure Assessment Methods Handbook (EPA 1989b)
- Role of the Baseline Risk Assessment in Superfund Remedy Selection Decisions (EPA 1990a)
- Guidance for Data Usability in Risk Assessment (EPA 1990b)
- Dermal Exposure Assessment: Principles and Applications (EPA 1992a)
- Risk Assessment Guidance for Superfund, Volume I: Human Health Evaluation Manual, Supplemental Guidance, Dermal Risk Assessment (EPA 1992b)
- Supplemental Guidance to RAGS: Calculating the Concentration Term (EPA 1992c)
- Guidelines for Exposure Assessment (EPA 1992d)
- Revisions to Sections 3.3.1 and 3.3.2 of the RAGS, Part B (EPA 1993a)
- Superfund's Standard Default Exposure Factors for the Central Tendency and Reasonable Maximum Exposure (EPA 1993b)
- Guidance Manual for the Integrated Exposure Uptake and Biokinetic (IEUBK) Model for Lead in Children (EPA 1994a)
- OSWER Directive: Revised Interim Soil Lead Guidance for CERCLA Sites and RCRA Corrective Action Facilities (EPA 1994b)
- Soil Screening Guidance: Technical Background Document (EPA 1996b)
- Approach for Addressing Dioxin in Soil at CERCLA and RCRA Sites (EPA 1998c)
- Soil Screening Guidance for Radionuclides: User's Guide and Technical Background Document Final Guidance (EPA 2000b)
- Guidance for Assessing Chemical Contaminant Data for Use in Fish Advisories, Third Edition (EPA 2000c)
- Estimating Dermal and Inhalation Exposure to Volatile Chemicals in Domestic Water (Schaum et al. 1994)
- Risk Assessment Guidance for Superfund: Volume III-Part A, Process for Conducting Probabilistic Risk Assessment (EPA 2001a)
- Supplemental Guidance for Developing Soil Screening Levels for Superfund Sites, Superfund, Office of Solid Waste and Emergency Response (EPA 2002)

- Recommendations of the Technical Review Workgroup for Lead for an Approach to Assessing Risk Associated with Adult Exposures to Lead in Soil (EPA 2003a)
- Human Health Toxicity Values in Superfund Risk Assessments (EPA 2003b)
- Guidelines for Carcinogen Risk Assessment (EPA 2005a)
- Supplemental Guidance for Assessing Susceptibility from Early-Life Exposure to Carcinogens (EPA 2005b)
- Guidance on Systematic Planning Using the Data Quality Objective Process (EPA 2006a)
- Systematic Planning: A Case Study for Hazardous Waste Site Investigations (EPA 2006b)
- National Recommended Water Quality Criteria: 2006 (EPA 2006c)
- 2006 Edition of the Drinking Water Standards and Health Advisories (EPA 2006d)
- Data Quality Assessment: Statistical Methods for Practitioners (EPA 2006e)
- *EPA provisional toxicity values support document* available on request from Technical Support Section, EPA Region 4 (EPA-PROV)
- The 2005 World Health Organization Reevaluation of Human and Mammalian Toxic Equivalency Factors for Dioxins and Dioxin-Like Compounds (Van den Berg et al. 2006)
- Exposure Factors Handbook 2011 Edition (Final Report) (EPA 2011)
- Human Health Evaluation Manual, Supplemental Guidance: Update of Standard Default Exposure Factors (EPA 2014a)
- Determining Groundwater Exposure Point Concentrations (EPA 2014b)
- Probabilistic Risk Assessment Methods and Case Studies (EPA 2014c)
- Probabilistic Risk Assessment to Inform Decision Making: Frequently Asked Questions (EPA 2014d)
- Radiation Risk Assessment at CERCLA Sites: Q&A (EPA 2014e)
- Technical Guide for Assessing and Mitigating the Vapor Intrusion Pathway from Subsurface Vapor Sources to Indoor Air (EPA 2015)
- Recommendations for Sieving Soil and Dust Samples at Lead Sites for Assessment of Incidental Ingestion (EPA 2016a)
- Updated Scientific Considerations for Lead in Soil Cleanups (EPA 2016b)
- Update to the Adult Lead Methodology's Default Baseline Blood Lead Concentration and Geometric Standard Deviation Parameters and the Integrated Exposure Uptake Biokinetic Model's Default Maternal Blood Lead Concentration at Birth Variable (EPA 2017)

- Region 4 Human Health Risk Assessment Supplemental Guidance, EPA Region 4, March 2018 Update (EPA 2018a)
- Update for Chapter 3 of the Exposure Factors Handbook: Ingestion of Water and Other Select Liquids, National Center for Environmental Assessment (EPA 2019)
- Release of Integrated Exposure Uptake Biokinetic Model for Lead in Children Version 2.0 and Revisions to the Default Parameters for the IEUBK Model for Lead in U.S. Children (EPA 2021a)
- ProUCL Version 5.2.0 Technical Guide Statistical Software for Environmental Applications for Data Sets with and without Nondetect Observations (EPA 2022a)
- ProUCL Version 5.2.0 User Guide Statistical Software for Environmental Applications for Data Sets with and without Nondetect Observations (EPA 2022b)

3.3.1.2 Commonwealth of Kentucky guidance documents and materials

- Kentucky Risk Assessment Guidance (KDEP 2002)
- Kentucky Guidance for Ambient Background Assessment (KDEP 2004a)
- Kentucky Guidance for Groundwater Assessment Screening (KDEP 2004b)
- Trichloroethylene Environmental Levels of Concern (KDEP 2004c)

3.3.1.3 DOE guidance documents and materials

- Multi-Agency Radiation Survey and Site Investigation Manual (MARSSIM) (DOE 2000c)
- Optimizing Radiation Protection of the Public and the Environment for Use with DOE O 458.1, ALARA Requirements (DOE 2014)
- Environmental Radiological Effluent Monitoring and Environmental Surveillance (DOE 2015)

3.3.1.4 Other materials

- Meeting Summary for the Risk Assessment/Risk Evaluation Meeting, February 7, 1996, in Atlanta, February 13, 1996, Conference Call (DOE 1996c)
- Guidance for Conducting Risk Assessments and Related Risk Activities for the DOE-ORO Environmental Management Program (Bechtel Jacobs Company LLC 1999)
- Minutes and notes from meetings of the PGDP Human Health Risk Assessment Working Group (RAWG 2000a, 2000b, 2000c, 2000d, 2000e, 2000f, 2000g, 2007a, 2007b, 2007c, 2012a, 2012b, and Appendix E of DOE 2017, DOE 2018, DOE 2019, DOE 2020, DOE 2021b, and DOE 2022b)
- Geochemical and Mineralogical Data for Soils of the Conterminous United States (USGS 2013)
- Programmatic Quality Assurance Project Plan (DOE 2022c)
- Biota Modeling in EPA's Preliminary Remediation Goal and Dose Compliance Concentration Calculators for Use in EPA Superfund Risk Assessment: Explanation of Intake Rate Derivation, Transfer Factor Compilation, and Mass Loading Factor Sources (ORNL 2016)

3.3.2 General Methods

The risk methods document generally follows guidance in EPA's RAGS (EPA 1989a) and Kentucky's *Risk Assessment Guidance* (KDEP 2002); however, there are issues for which the two guidance documents differ. In those cases, the Risk Methods Document reconciles these two different approaches. The document also serves to address site-specific issues where guidance may be lacking and/or to document site-specific agreements among representatives of the RAWG from DOE, EPA, and Kentucky.

3.3.2.1 Format for the baseline human health risk assessment

The outline that will be followed when preparing baseline human health risk assessments for PGDP is provided in Appendix C of this document. This outline is consistent with that in RAGS, Part A (EPA 1989a), and in *Kentucky Risk Assessment Guidance* (KDEP 2002) and includes all sections that must be included in a complete baseline human health risk assessment. As such, some portions of the outline may not be applicable to some baseline human health risk assessments of limited scope; however, any baseline human health risk assessment prepared for PGDP will include the major and second level headings in the order presented. Major headings that will appear in all baseline risk assessments are "Results of Previous Studies," "Identification of Chemicals of Potential Concern," "Exposure Assessment," "Toxicity Assessment," "Risk Characterization," "Uncertainty in the Risk Assessment," "Conclusions and Summary," and "Remedial Goal Options Development." In addition, each baseline human health risk assessment will contain introductory material that delineates the scope and objectives of the assessment.

Examples of the format for tables that will be used in the risk assessment are presented in Exhibit 3.8. List of Chemicals of Potential Concern; Exhibit 3.9. Summary of Pathway Analysis in the Exposure Assessment; Exhibit 3.10. Presentation of Exposure Point Concentrations; Exhibit 3.11. Chemical-Specific Parameters; Exhibit 3.12. Daily Intakes (Chronic Dose) for Receptor 1; Exhibit 3.13. Exposure Route Summary for the Current Use Scenario—Systemic Toxicity; Exhibit 3.14. Driving Contaminants' Summary for Current Use Scenario—Systemic Toxicity; Exhibit 3.15. Summary of Risk Characterization; Exhibit 3.16. Summary of Uncertainty Analysis; and Exhibit 3.17. Presentation of Remedial Goal Options. Shorter summary tables for the body of the report will summarize the following information:

- Land use scenarios and media assessed for each source area;
- Scenarios for which human health risk exceeds *de minimis* levels; and
- A table for each source summarizing the COCs and POCs, as well as the contribution of each COC and POC to the total risk and hazard.

3.3.2.2 Presentation of results from previous studies

In all baseline risk assessments prepared for PGDP, the results will be presented from previous risk assessments and other risk evaluations that are relevant to the unit or area being assessed. These results will be included to allow for a comparison between results of earlier work and the results of the current baseline risk assessment. Differences seen will be discussed in the observations section of the current baseline risk assessment.

The format for presenting the results of the earlier risk assessments will follow that which will be used for reporting previous studies in the RI work plan. This is discussed in detail in Section 3.1.2. For risk evaluations, if any, that are not risk assessments, results will be presented verbatim and without interpretation. Relevant results from these studies also may be used in the uncertainty discussion of the current baseline human health risk assessment.

3.3.3 Data Evaluation Methods

The primary purpose of this section of the baseline human health risk assessment will be to develop the list of COPCs used in the assessment. In this section, the data quality/data usability review, procedures to screen data, a summary of the results of the screening, and a final list of COPCs will be presented. Additionally, this section will provide site-specific characterization data used in the exposure assessment. Methods to complete each of these activities are presented in the following.

3.3.3.1 Data quality/data usability review

The overall goal of the data quality/data usability review is to develop a data set of known quality that is representative of the site and is reproducible. Use of this systematic approach is consistent with EPA guidance (EPA 2006f; EPA 2006e). The data quality/data usability review process (Figure 2.2) incorporates the aspects of data quality/data usability [measurement quality objectives (MQOs)] with an evaluation of planned data uses for each project DQOs to make a determination concerning the suitability of historical/current project data for use in risk assessment. The initial steps of data assessment and data validation generally are completed by a subject matter expert before the results are provided to the risk assessor. The data quality assessment (DQA) examines the data set to ensure that the MQOs have been met and that the data are sufficient and representative of the site or source investigated. Figure 3.2 [from the EPA DQA guidance (EPA 2006f)] is provided to illustrate how DQA fits into the data evaluation process.

3.3.3.2 Procedures to screen or evaluate data to determine COPCs

Data screening to develop the list of COPCs will be performed in the following eight steps.

• Step 1: Evaluation of sample design and locations. Data will be examined to ensure that the samples from which data were derived were collected using sampling methods that are adequate to determine the nature and extent of contamination for the particular unit or area being assessed. Data not from the unit or area under investigation or not useful in determining contaminant migration from the unit or area will not be used quantitatively in the assessment because these data are not representative of the unit or area for which remedial actions are being considered. In particular, when considering groundwater sampling results, only data from samples collected from wells located in contaminant plumes will be used.

For radionuclides, MARSSIM is the guidance used for surface soil sampling for characterization, remedial support surveys, and final status surveys (EPA 2018a).

• Step 2: Evaluation of sampling and analytical methods. Data will be examined to ensure that the sampling methods and analytical methods used in the laboratory are consistent with EPA-approved methods for nonradionuclides. Data for nonradionuclides not from EPA-approved methods will not be used quantitatively in the risk assessment, but may be used qualitatively. Methods for radionuclides will be evaluated during the DQO process to ensure that data quality requirements can be achieved. Also in this step, groundwater and surface water data will be examined, and data from the analyses of filtered water will be deleted from the data set. Only results from unfiltered samples will be used quantitatively in baseline human risk assessments performed at PGDP. Note: Filtered groundwater and surface water data may be used in the uncertainty section of the assessment when discussing data sources and their effects on risk estimates.

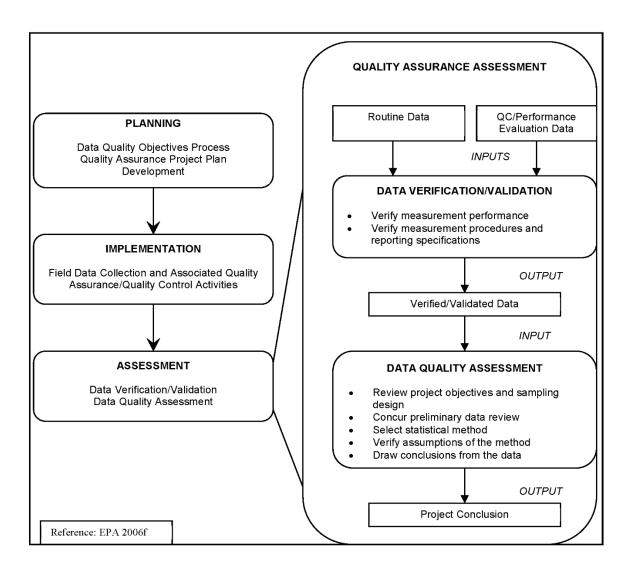


Figure 3.2. Data Life Cycle

For many sites, survey-type data such as X-ray fluorescence (XRF) data and results from polychlorinated biphenyl (PCB) field test kits are available in addition to the laboratory analytical data. The primary use of such data is for site characterization, but these survey-type data also can play a role in risk-based decision making. Survey-type data assist in determining the distribution of COPCs and can be used to identify which sets of laboratory data should be combined to develop site average contaminant concentrations. Potentially, survey-type data also could be combined with lab data in a risk assessment to determine the average concentrations for contaminants, but this would require demonstrating that the lab and survey-type data possess similar detection limits and analytical uncertainty. Collection of survey-type data should follow methods consistent with those developed by EPA (EPA 2018a). Project teams will need to address the uncertainty if detection limits from XRF and PCB field test kits cause EPCs to be inconsistent with maximum detected results. Addressing this uncertainty may include obtaining additional sampling/analytical data. Detection limits for the XRF and PCB test kits optimally should be below levels consistent with expected cleanup levels. In addition, a DQA would need to be completed to show that both types of data sets are comparable and representative of the site conditions. This DOA either could be in the risk assessment or in a report completed prior to or in concert with the risk assessment.

Finally, whenever survey-type data are used for guiding how lab data are handled or are combined with lab data, then the risk assessment would need to have an uncertainty discussion that appropriately identifies (a) how the results of the risk assessment could vary if the survey type data were not used and (b) how the use of the survey data increases or decreases the risk of making an incorrect risk-based decision for a location.

Step 3: Evaluation of sample quantitation limits.

Chemicals. The sample quantitation limits for each analyte and sample will be examined to determine if these limits were below the concentration at which the analyte may pose an unacceptable risk or hazard to human health. If the maximum sample quantitation limit for an analyte (over all samples within a medium) is greater than the concentration that may pose an unacceptable risk or hazard to human health, and the analyte is not detected in any sample, then the data for that analyte will be deemed suspect. Data from these analytes will not be used quantitatively in the risk assessment, but the potential risk or hazard from exposure to media potentially containing these analytes will be examined qualitatively. In developing the qualitative assessment for these data, the maximum quantitation limit for the analyte (in all samples from a medium) will be compared to the appropriate no action residential PRG if historical or process information indicates that the analyte potentially could be present. One-half the maximum quantitation limit for the analyte (in all samples from a medium) will be used in this comparison if historical or process information indicates that the analyte is not expected to be present.

Radionuclides. The analysis for radionuclides will be performed in two steps. In the first step, the MDC/minimum quantification concentration (MQC) for each analyte and sample will be examined to determine if these limits were below the concentration or activity concentration at which the analyte may pose an unacceptable risk (or radiological dose). If the maximum MDC/MQC for an analyte over all samples within a medium is greater than the concentration or activity concentration that may pose an unacceptable risk (or radiological dose) to human health and the analyte is less than the minimum detectable activity concentration MDC/MOC in any samples, then the data for that analyte will be deemed suspect. The MDCs used for radionuclides should be the MDCs established in the MARLAP Manual (EPA 2004b), which provides guidance for evaluating sample quantitation limits (SQLs) for radionuclide data. For radionuclides, all reported values, including negative values, ¹⁰ will be used to derive the EPCs under current conditions.

Survey-type data. When XRF data are used in the derivation of EPCs, all XRF values, including negative values, will be used as reported. Other survey-type data (such as PCB field test kits) should be used in accordance with project-specific review of the data and performance of the method.

See Figure 3.3 for an example of Step 3.

⁹ Radionuclide results reported with an uncertainty that indicates the result could fall below the MDC will be reported as detections or nondetects or otherwise flagged in the data verification/validation and assessment process indicating the detected result is tentative.

¹⁰ Negative results may be reported due to a statistical determination of the counts seen by a detector, minus a background count.

Evaluation of Sample Quantitation Limits

Chemicals:

Consider the following results for Chemicals W, X, Y, and Z. Assume that Chemicals W and Y are site-related contaminants and that Chemicals X and Z are not site-related. Also, let the data qualifier (U) be defined as not detected at the sample quantitation limit (SQL).

Chemical	Sample 1	Sample 2	Sample 3	Sample 4	Screening Value
W	10U	10U	10U	10U	5
X	10U	10U	10U	10U	5
Y	10U	6	10U	10U	5
Z	1U	1U	1U	1U	5

Then, following the rules in Step 3 of the data evaluation process:

- Results for Chemical W are suspect because the maximum SQL overall results (10) is greater than the screening value (5), and Chemical W was not detected in any sample. Because Chemical W is site-related, the qualitative risk analysis of this chemical's potential effect would use the full SQL.
- Results for Chemical X are suspect because the maximum SQL overall results (10) is greater than the screening value (5), and Chemical X was not detected in any sample. Because Chemical X is not site related, the qualitative risk analysis of this chemical's potential effect would use one-half the SQL.
- Results for Chemical Y are not suspect even though the maximum SQL exceeds the screening value because Chemical Y was detected in one sample.
- Results for Chemical Z are not suspect because the maximum SQL is less than the screening value.

For radionuclides, SQLs should be evaluated in accordance with the guidance in the Multi-Agency Radiological Laboratory Analytical Protocols (MARLAP) Manual (EPA 2004b).

Note: Other data qualifiers associated with the data must also be considered during data evaluation. Please see Step 4 of the data evaluation process.

Figure 3.3. Example of Step 3—Evaluation of Sample Quantitation Limits Laboratory Analytical Data

- Step 4: Evaluation of data qualifiers and codes. Generally, the rules presented in RAGS, Part A, Exhibits 5.4 and 5.5 (EPA 1989a) will be used to evaluate all data qualifiers and codes attached to analytical results for chemicals; however, data with a "B" qualifier (i.e., analyte also found in associated blank) will be examined by analyte to ensure that site-related analytes are not eliminated. For other analytes, the "5 and 10X's Rule" described in RAGS, Part A, (EPA 1989a) will be considered. In addition, the method used in data validation to examine blank contamination will be evaluated. If data validation qualified sample results as "U" (i.e., analyte not detected) instead of "B" when blank contamination was present and the analyte passed the "5 and 10X's Rule," then the data will be reevaluated. Specifically, if chemical data are qualified "B," and the value is less than that defined by the "5 and 10X's Rule," then the data will be assumed to be a nondetect and the reported value will be used to derive the EPC.
 - Evaluation of radionuclide data will follow rules agreed upon by the Commonwealth of Kentucky Radiation Health Branch [formerly the Kentucky Radiation Health and Toxic Agents Branch (KYRHTAB)] and DOE (RAWG 2000a through 2000g). The data assessment qualifiers that will appear and their description are as follows:
 - **KYRHTAB-LT:** KYRHTAB has performed an independent data assessment and the results are less than the MDC or detection limit and should not be plotted.

- **KYRHTAB-50:** KYRHTAB has performed an independent data assessment and the radiation counting uncertainty is greater than 50% of the analytical results.
- KYRHTAB-ER: KYRHTAB has performed an independent data assessment and the data present error problems (i.e., no counting uncertainty or zero counting uncertainty).
- KYRHTAB-OK: KYRHTAB has performed an independent data assessment and the data are acceptable for use.
- Step 5: Elimination of analytes not detected. Generally, any chemical not detected in at least one sample from a medium will be deleted from the data set. Any radionuclide for which no analytical results exceed its MARLAP MDC also will be deleted from the project dataset, provided the MDC is an acceptable level for the project. If a chemical analyte is suspected of being present at very low concentrations (i.e., below the quantitation limit) due to cross-media contamination or is suspected of being present based on historical or process information, the analyte may remain in the data set even though the analyte was not detected. In this case, the concentrations used to determine the representative or EPC for the analyte will be the sample quantitation limits for the analyte in the medium. For classes of analytes such as PAHs, PCBs, and dioxins/furans, if one compound is detected at a concentration greater than a screening value and is identified as a COPC, then others in that class will be assumed to be present as well. The method used to analyze these classes of compounds is presented later in this section.
- Step 6: Examination of toxicity of detected analytes. The maximum concentrations and activities of analytes remaining in the data set will be compared to no action residential use risk-based PRGs by medium. The PRGs used in this comparison will be the no action values for the child found in Appendix A. Those analytes with a maximum detected concentration less than each respective no action risk-based PRG will be eliminated from the data set unless the analyte has a bioaccumulation factor for fish equal to or greater than 100 (DOE 1996d). Note: The uncertainty introduced through the application of this screening procedure will be examined quantitatively in the uncertainty analysis portion of the baseline risk assessment. The derivation of the risk-based PRGs used in this comparison is described in Appendix B of this document.
- Step 7: Examination of analyte concentrations of essential nutrients detected in site samples. Analytes not removed from the data set in previous steps will be examined to determine if any are essential nutrients. Seven analytes known to be essential nutrients and known to be toxic only at extremely high concentrations will be removed from the data set on the basis of regulatory guidance (EPA 2018a). These analytes are calcium, chloride, iodine, magnesium, potassium, sodium, and phosphorus. No other analytes known to be essential nutrients will be deleted from the data set on the basis of this screen. Any uncertainty regarding retention of essential nutrient in the list of COPCs will be discussed in the uncertainty section of the risk assessment.
- Step 8: Comparison of analyte concentrations detected in soil and groundwater samples to analyte concentrations detected in background. This comparison is performed as part of the development of the list of COPCs. As a first step, maximum detected concentrations of analytes will be compared to the background concentrations presented in Appendix A. Analytes not detected at a concentration greater than the background concentration will not be retained as COPCs. Analytes detected at concentrations greater than their background concentration may be retained as COPCs, depending upon the outcome of other screening steps. Analytes retained as COPCs, however, may be

3-20

¹¹ These types of decisions (acceptable MDCs) would be a product of the consensus of the FFA parties arrived at during project discussions at the appropriate stage in document development.

considered with the full range of background as part of the uncertainty analysis. This analysis, if completed, will be done to determine if the analyte is generally present at concentrations above its background concentration or if the detected concentrations of the analyte above the selected background concentration is consistent with natural enrichment. The impacts on risk characterization of not retaining an analyte on the basis of the background screen will also be considered in the uncertainty analysis.

During the development of the list of COPCs, concentrations of total carcinogenic PAHs, PCBs, and dioxins/furans (dioxins) will be derived. Total carcinogenic PAHs, total PCBs, and total dioxins will be derived to allow for the correct use of the toxicity screen described in Step 6 and to allow for correct calculation of ELCR from exposure to these organic compounds.

When deriving total carcinogenic PAHs, the toxicity equivalence factors (TEFs) presented in Human Health Risk Assessment Protocol for Hazardous Waste Combustion Facilities (EPA 2005c) will be used. These TEFs are presented in Table 3.1. Note that these TEFs will be applied to the concentrations of detected PAHs in each sample and that the total carcinogenic PAH concentration in a sample will be the sum of the products of each carcinogenic PAH and its TEF. For samples in which PAHs are not detected, the value for the minimum detection limit of the PAHs with TEFs will be used in the calculation of the EPC.

When deriving total PCBs [if this analyte (i.e., Total PCBs) is not reported in the data set], the detected concentrations of each PCB within a sample will be summed. For samples in which no PCBs are detected, the value for the minimum detection limit of the PCBs will be used in the calculation of the EPC. If there are detection limits for PCBs exceeding risk-based concentrations, this issue should be discussed in the uncertainty section. Note that there are no TEFs to use when deriving total PCBs from individual Aroclors. If dioxin-like PCBs are detected at a site, they should be added to the total PCBs after weighting with the TEFs for those compounds in Van Den Berg, et al. 2006.

When deriving total dioxin, the TEFs presented in *Federal Register*: May 10, 2007 (Volume 72, Number 90), *Dioxin and Dioxin-like Compounds; Toxic Equivalency Information* will be used. These TEFs are presented in Table 3.1. Note that these TEFs will be applied to both the concentrations of detected dioxins and furans and to one-half the sample quantitation limit of undetected dioxins and furans, when one or more dioxin or furan is detected. The total dioxin concentration in a sample will be the sum of the products of each dioxin/furan and its TEF. For samples in which no dioxin or furan was detected, the minimum detection limit for 2,3,7,8-tetrachlorodibenzo-p-dioxin (TCDD) will be used as the value for the total dioxin concentration. If there are detection limits for dioxins and furans exceeding risk-based concentrations, this issue should be discussed in the uncertainty section.

Table 3.1. Toxicity Equivalency Factors for Carcinogenic PAH Compounds and Dioxins/Furans

Carcinogenic	Toxicity	Dioxin/Furan Compound ²	Toxicity
PAH Compound ¹	Equivalence Factor		Equivalence Factor
Benzo(a)pyrene	1.0	2,3,7,8-TCDD	1.0
Benzo(a)anthracene	0.1	1,2,3,7,8-PeCDD	1.0
Benzo(b)fluoranthene	0.1	1,2,3,4,7,8-HxCDD	0.1
Benzo(k)fluoranthene	0.01	1,2,3,6,7,8-HxCDD	0.1
Chrysene	0.001	1,2,3,7,8,9-HxCDD	0.1
Dibenzo(a,h)anthracene	1.0	1,2,3,4,6,7,8-HpCDD	0.01
Indeno(1,2,3-c,d)pyrene	0.1	OCDD	0.0003
All other PAHs	0	2,3,7,8-TCDF	0.1
		1,2,3,7,8-PeCDF	0.03
		2,3,4,7,8-PeCDF	0.3
		1,2,3,4,7,8-HxCDF	0.1
		1,2,3,6.7,8-HxCDF	0.1
		1,2,3,7,8,9-HxCDF	0.1
		2,3,4,6,7,8-HxCDF	0.1
		1,2,3,4,6,7,8-HpCDF	0.01
		1,2,3,4,7,8,9-HpCDF	0.01
		OCDF	0.0003

¹ TEFs from EPA 2005c

3.3.3.3 Presentation of data evaluation

A summary of the data evaluation will be provided in both narrative and tables. Tables from each step of the data evaluation process may be presented. The detailed data tables, if voluminous, should appear in an appendix to the risk assessment; however, the summary tables described earlier (see Section 3.3.2.1) should appear in the main text of the assessment. At minimum, a table listing the COPCs for the assessment should appear in the main text. An example of the information that should appear in this summary table is in Exhibit 3.8.

Exhibit 3.8. List of Chemicals or Radionuclides of Potential Concern

Analyte	Frequency of Detection ¹
Site and Medium ²	
Analyte # 1	
Analyte # 2	
•	•
•	•
Analyte # N	

¹This value will be the number of samples in which the analyte was detected over the number of samples in which an analysis for the analyte was performed.

3.3.4 Site-specific characterization information

Several pieces of site-specific characterization information are relevant to virtually all baseline human health risk assessments performed for PGDP because they explain resource use around PGDP. Because this information is in the form of interviews and letters, it generally is not readily available; therefore, this information is included in Appendix E of this document to provide a ready source of these materials. Appendix E, presents the following documentation.

² TEFs from Van Den Berg, et al. 2006

² A list of COPCs will be presented for each site and medium combination.

- Reference to the Phase I Site Investigation results of surface water and groundwater users survey to determine groundwater use near PGDP (CH2M HILL 1991). 12
- Summary of agricultural practices in Ballard County, Kentucky.
- Summary of the agricultural practices in McCracken County, Kentucky.
- Area of crop land in Ballard and McCracken County, Kentucky.
- Recreational use of Bayou and Little Bayou Creeks near PGDP.
- Annual harvests of turkeys and deer, in McCracken and Ballard Counties, Kentucky, and waterfowl in Ballard County, Kentucky.
- Reports entitled "Planning Issues for Superfund Site Remediation" and "Quantitative Decision Making in Superfund: A Data Quality Objectives Case Study" from *Hazardous Materials Control* regarding use of exposure units in risk calculations and remedial decisions.
- A link to Kentucky Risk Assessment Guidance.
- Environmental Indicators flowchart submitted to the Hazardous Waste Branch of the Kentucky Division for Waste Management.
- The table of parameters for probabilistic risk assessment (PRA) from the Southwest Plume Investigation report. This table provides the parameter values used for the PRA in that report, which should be considered for use in other PRAs. The values in the table do not represent specified default values for use in all PRAs.
- Lead-210 and PAHs at PGDP.
- Guidance on development of site-specific soil screening levels and site-specific dilution attenuation factors to be implemented when scoping projects.
- Human health information for the Paducah vapor intrusion evaluation.
- Minutes from the previous year's RAWG meetings.
- A comparison of the geochemical analyses summarized in the draft RI/FS report for the C-400 Complex Operable Unit, with the geochemical parameter distributions utilized in the probabilistic fate and transport modeling presented in the Site Investigation Report for the Southwest Groundwater Plume at the Paducah Gaseous Diffusion Plant, Paducah, Kentucky, DOE/OR/07-2180&D2/R1 (DOE 2007).

3.3.4 Exposure Assessment Methods

The primary purpose of this section of the baseline human health risk assessment will be to report the results of the exposure assessment for each unit or area investigated. In this section, the exposure setting for each unit or area will be characterized, exposure pathways will be identified, exposure will be quantified (i.e.,

¹² Although completed in 1989, these surveys are relevant to current use of surface water and groundwater because these survey results were collected before the current Water Policy was in place; therefore, these survey results represent likely surface water and groundwater use within the Water Policy Box and in adjacent areas in the absence of PGDP-derived contamination.

chronic dose or intake calculated), and chronic doses (or intake) will be presented. Methods to complete each of these steps are discussed in the following sections.

3.3.4.1 Characterize the exposure setting

This section of the exposure assessment or other portions of the document will describe the physical setting of each unit, including meteorology, climate, vegetation, soil type, surface hydrology, groundwater hydrology, and geology. In addition, the surrounding populations will be characterized as needed. Specific note will be given to determining if sensitive subpopulations may be present. In risk assessments in RI reports, the information presented concerning climate, vegetation, soil type, surface hydrology, groundwater hydrology, and geology will be brief, and references will be to material presented in earlier sections of the RI report. (Note: A brief presentation of this material must be included in the baseline risk assessment because the FFA states that the baseline risk assessment is to be written as a stand-alone report.) In baseline risk assessments not in RI reports, the information presented concerning climate, vegetation, soil type, surface hydrology, groundwater hydrology, and geology will be more extensive.

Current and potential future land use and the time frame for future use also will be discussed in this section of the exposure assessment. The most likely future land use will be determined using information in the most recent PGDP SMP; however, because future land use over time is uncertain, the use scenarios considered in the baseline risk assessment will not be governed by that information alone. Use scenarios that will be considered in all baseline risk assessments under future conditions are rural residential, recreational, industrial, outdoor worker, and excavation. Appropriate use scenarios may be evaluated during project scoping.

Finally, this section of the baseline human health risk assessment will integrate the preceding information and declare the unit or area under investigation either as a source or integrator unit and identify exposure points. Definitions used to determine whether the area or unit is a source or integrator are as follows:

- **Source unit.** Those units or areas that may release contaminants to other units or areas.
- Integrator unit. Those units or areas that accumulate contaminants from source units or areas.

Generally, application of these definitions to units and areas to be investigated at PGDP shows that all areas on-site where contamination exists (e.g., the soil and other material at burial grounds, spill areas, and landfills) are source areas. Integrator units identified using these definitions are air, groundwater (e.g., RGA), and surface water (e.g., Bayou and Little Bayou Creek watersheds and the Ohio River).

Also in this section of the exposure assessment, exposure points will be evaluated. For source units, the exposure points that will be evaluated under current conditions are at the unit or area ("hot spots" may be

evaluated separately) and at points downgradient to which contamination may migrate. Downgradient points that will be evaluated for risk communication purposes include at the PGDP industrialized area boundary [i.e., the boundary of the area corresponding to the industrial land use delineated in the SMP (DOE 2022a)]; at the DOE property boundary; and at Little Bayou Creek. Note that for some source units, one

Industrialized Area

Area corresponding to the industrial land use delineated in the Site Management Plan.

or more of these exposure points may not be relevant. The exposure assessment will provide an explanation for exposure points not selected for risk characterization.

For integrator units, exposure points that will be considered are those within the contaminated area (e.g., above the contaminated groundwater plume or along the contaminated ditch) and at areas downgradient. Generally, exposure points that consider migration from a source will consider the time of exposure. For example, for exposure to groundwater both at a source and at the facility boundary, risk or hazard from

exposure to measured concentrations under current conditions and future conditions will be determined. In addition, risk or hazard from exposure to expected future concentrations or activities will be modeled to determine the risk or hazard that may occur under potential future conditions as contaminants migrate from the source to the underlying aquifer. Exposure to contaminants in or migrating to the surface water integrator unit will be handled similarly. The mechanism that will be used to determine the extent of modeling that will be used in a baseline human health risk assessment is discussed later.

3.3.4.2 Identification of exposure pathways

This section of the exposure assessment will delineate the pathways through which the receptors may be exposed under both current and future conditions. For current receptors, these pathways and their parameters should be based on realistic exposures; for future receptors, these pathways and their parameters should be based on reasonable maximum exposure values. The goal of this material will be to provide a complete depiction of all exposure pathways for current and future uses. To achieve this goal, this section will present conceptual site models and supporting text. Also, in this section, each pathway will be described in terms of source, exposure route, exposure point, and receptor. This format will be followed because all four must be present for a complete pathway to exist. Note: Potential pathways not containing all four items will be described as being incomplete, and text justifying their omission from the assessment will be provided. Potential pathways that will be considered in all assessments are described herein.

Exposure assessments in baseline human health risk assessments completed in the past indicate that at least 24 exposure pathways should be considered as potential pathways in all assessments. These pathways are listed below. (Note: Additional pathways, such as contact with buried waste and modeled vapor intrusion from groundwater or soil gas to indoor air, may be reasonable for some units or areas; these pathways are not included.)

- Ingestion of groundwater as a drinking water source;
- Inhalation of volatile constituents emitted from groundwater during household use (including showering);
- Inhalation of volatile constituents emitted from groundwater during construction/excavation activities; 13
- Dermal contact with groundwater while showering;
- External exposure to ionizing radiation emitted by constituents in groundwater while showering;
- Inhalation of volatile constituents emitted from groundwater during irrigation;
- Incidental ingestion of soil;
- Dermal contact with soil;
- Inhalation of particulates emitted from soil;

-

¹³ The Virginia Department of Environmental Quality's Virginia Unified Risk Assessment Model includes a construction worker trench model that can be used to estimate the migration of volatile constituents from soil gas and/or groundwater to the air within a trench (VDEQ 2022). The results of this model can be useful in site screening and project scoping; however, additional information developed on a project-specific basis is expected to be needed for decision making.

- Inhalation of volatile constituents emitted from soil;
- External exposure to ionizing radiation emitted by constituents in soil;
- Incidental ingestion of surface water while swimming or wading in creeks or natural or man-made ponds;
- Dermal contact with surface water while swimming or wading in creeks or natural or man-made ponds;
- External exposure to ionizing radiation emitted by constituents in surface water while swimming or wading in creeks or natural or man-made ponds;
- Incidental ingestion of sediment while swimming or wading in creeks or natural or man-made ponds;
- Dermal contact with sediment while swimming or wading in creeks or natural or man-made ponds;
- External exposure to ionizing radiation emitted by constituents in sediment while swimming or wading in creeks or natural or man-made ponds;
- Consumption of fish taken from creeks or natural or man-made ponds;
- Consumption of vegetables and produce raised in contaminated soil;
- Consumption of irrigated vegetables;
- Consumption of beef from animals contaminated by consuming vegetation (pasture and concentrates) irrigated with contaminated water or grown on contaminated soil, by drinking contaminated water, or ingesting contaminated soil;
- Consumption of dairy products (i.e., milk) from animals contaminated by consuming vegetation (pasture and concentrates) irrigated with contaminated water or grown on contaminated soil, by drinking contaminated water, or ingesting contaminated soil;
- Consumption of pork from animals contaminated by consuming vegetation (concentrates) irrigated with contaminated water or grown on contaminated soil or by drinking contaminated water;
- Consumption of poultry products from animals drinking contaminated water; and
- Consumption of game (i.e., deer, rabbits, and quail) contaminated by consuming contaminated vegetation or soil and ingesting water.

While these pathways have been found to be reasonable in past assessments, not all may be reasonable, or complete, for future assessments; therefore, the decision as to which pathways to quantify will be made on a project-specific basis. In any case, the rationale for the inclusion or exclusion of any of the pathways listed herein will be included in the exposure assessment.

It is important to note that the pathways relating to livestock consumption are not reasonable for most source units. This is because most source units are too small to support livestock in addition to a homestead and garden. Generally, a source unit will be required to be larger than two acres to be considered for livestock production. (This requirement assumes that a minimum of two acres is required for a home and associated garden.) Note: Under this definition, all integrator unit assessments will contain an assessment of risk from

consumption of livestock because the area they cover is greater than two acres. In assessments where livestock consumption is included, the range size for each beef or cow will be two acres per head (Morrison 1959).

Older references, such as *Feed and Feeding* (Morrison 1959), provide exposure scenario rates representative of a local farming scenario, including foraging and production of feed locally, and also include higher rates of water usage than new techniques for livestock production. These scenarios and rates facilitate the calculation concentrations of chemicals in meat that would result from local farming as opposed to large-scale commercial farming that would use purchased feed. Similarly, older references are also used for the raising of laying hens under the farming scenario and fish ingestion rates. Newer references may be considered, as appropriate, for specific projects.

For baseline human health and ecological risk assessments that incorporate larger areas (such as the final sitewide baseline human health and ecological risk assessment), scenarios will be evaluated on a project-specific basis including evaluation of exposure due to unit size (e.g., recreational/hunter scenarios where wild game have a range much larger than 0.5 acres).

Using the characterization information and pathway analysis, a conceptual site model will be developed for each unit or area. The format that will be used for the conceptual site models is that in Figure 3.1. Note: When presenting the conceptual site models for multiple units or areas in a single baseline human health risk assessment, the units or areas may be grouped to reduce the number of figures that need to be presented.

3.3.4.3 Quantification of exposure

To quantify exposure or dose, both the EPC and the exposure factors are required. Here, the EPC can be defined as the concentration or activity concentration of the COPC in the environmental medium ingested, inhaled, contacted, or consumed, and the exposure factor can be defined as the product of the exposure parameters describing the degree of exposure to the environmental medium in terms of duration or frequency of exposure and mass of the receptor.

EPCs under current conditions of all COPCs for which environmental samples were taken will be determined using the following procedure.

- (1) If results from fewer than ten samples are available, then the EPC will be the maximum detected concentration.
- (2) If results from ten or more samples are available, then the most recent version of EPA's ProUCL software will be used to determine the EPC. The value selected as the EPC will be the value recommended by ProUCL, noted as the "Potential UCL to Use." EPA's ProUCL software 14 incorporates a number of different distributional tests that may be used to calculate

From Soils Operable Unit RI Report (DOE 2012):

The representative sampling design for the SWMUs was gridding. In some instances (such as SWMUs/AOCs not grid sampled in summer 2010), when a grid was applied to the SWMUs/AOCs, a grid lacking a sample result resulted. In order to fill a grid lacking a sample result, the average of the grids within the exposure unit with sampling results was used. Attachment D2 [of the Soils Operable Unit RI Report] presents an uncertainty evaluation in determining EPC values using these averages against EPC values calculated without using the averages or the maximum value, as applicable. An example for determining the EPC through averaging is illustrated below.

If the SWMU/exposure unit combination had less than 10 grids, the maximum grid result was used as the EPC. If the SWMU/exposure unit combination had 10 or more grids, the grid values were used to determine the EPC. Grid values were determined following guidance in the work plan. Basically, the maximum detected result from within the grid applies to the grid. If not detected, the minimum detection limit applies to the grid.

If a grid had no result (detect or nondetect) for the COPC, an average of the results for the grids with results was used.

¹⁴ Software is available at www.epa.gov/land-research/proucl-software.

the most appropriate EPC (EPA 2022a). In the current version of ProUCL, the software has computation methods for handling data sets with nondetect values. Unless other determinations are made during project scoping, nondetect values should be handled according to the recommendations in the ProUCL User Guide (EPA 2022b). Additional information regarding the statistics and computation methods used in ProUCL can be found in the User Guide and in the ProUCL Technical Guide (EPA 2022a). Additionally, it is unlikely that the upper confidence limits (UCLs) based upon those methods will exceed the maximum detected value, unless some outliers are present in the data set. The RAWG has concluded that the 95% UCL should be used as the EPC and if the 95% UCL exceeds the maximum detected concentration, then the uncertainty needs to be discussed in the uncertainty section of the risk assessment.

Options to determine the ten or more samples may include use of grid values. It is recommended that a geostatistical approach utilizing Spatial Analysis and Decision Assistance (SADA) or similar software be used to estimate values for empty grids. SADA is available at http://www.sadaproject.net/. Alternately, an average value may be used. An example is shown in the text box [from Soils Operable Unit RI Report (DOE 2012)]. These options should be discussed and agreed to in the planning phases of projects.

In determining the UCL when the medium is soil, data will be segregated into depth intervals relevant to receptors.

- For scenarios in locations inside the industrialized area, the following will be used to estimate the EPC:
 - Excavation worker: data from samples collected from 0 to 10 ft bgs, ¹⁵
 - Outdoor worker: data from samples collected from 0 to 1 ft bgs, and
 - All other scenarios: data from samples collected from 0 to 1 ft bgs.
- For scenarios in locations outside the industrialized area, the following will be used to estimate the EPC:
 - Excavation worker: data from samples collected from 0 to 10 ft bgs, ¹²
 - Outdoor worker performing maintenance-type activities: data from samples collected from 0 to 10 ft bgs, 8 and
 - All other scenarios: data from samples collected from 0 to 1 ft bgs.

In determining the UCL when the medium is groundwater, data from samples from each potable aquifer (i.e., RGA and McNairy Formation) will be used; however, data will be summarized within and not over aquifers, consistent with EPA guidance (EPA 2014b). Note: For the groundwater integrator investigations (e.g., that for the Groundwater Operable Unit), the representative concentration for groundwater may be the average concentration of the samples taken from wells within the contaminant plume if data are sufficient. EPA guidance recommends calculating the 95% UCL of the arithmetic mean as the EPC for risk assessments (EPA 2014b). It is generally desirable to use at least 10 data points for each contaminant (e.g., 5 wells and 2 rounds of data representative of current conditions equate to 10 data points) to compute a 95% UCL. If the 95% UCL is greater than the maximum detected concentration, EPA guidance recommends that the EPC default to the maximum detected concentration for that contaminant. The RAWG has concluded that the 95% UCL should be used as the EPC and if the 95% UCL exceeds the maximum detected concentration, then the

¹⁵ Unless information indicates that results from samples collected at deeper depths (i.e., 0–16 ft bgs in areas where infrastructure is found) should be included in the derivation of the EPC.

uncertainty needs to be discussed. If less than 3 wells are within the core of the plume, maximum detections may be used as the EPC for that contaminant (EPA 2014b). In addition, as with soil, the wells used in each calculation may be grouped so that risk or hazard at differing contaminant concentrations and in various areas may be estimated. Decisions concerning the method that will be used to estimate the concentration of COPCs for the groundwater integrator unit will be made on a case-by-case basis and will be justified in the baseline risk assessment.

Risks from water drawn from the UCRS will not be presented in the main body of the risk assessment because this water source is not considered to be an aquifer due to low yield. However, risks from ingestion of water from this source will be considered at least qualitatively in the uncertainty section of the risk assessment.

Finally, for some samples, duplicate or split-sample analyses may be available. When calculating the representative concentration, the maximum value reported in the duplicate or split-sample analysis will be used. Duplicate and split-sample results will not be averaged when calculating the representative concentration in baseline risk assessments performed for PGDP.

The EPCs and activities used for future conditions will depend on the time frame for which risk or hazard is being quantified. At minimum, for all assessments for PGDP, risk and hazard to potential future users, will be quantified using the current EPCs and activities. In addition, for those sites and areas where future concentrations or activities may increase, modeled concentrations will be used. To determine if modeling is needed, the maximum soil concentrations and activities at the source (over all depths) for each analyte will be compared to the appropriate groundwater protection PRG (PRGs appear in Appendix A). If the maximum soil concentration exceeds the groundwater protection PRG, then future concentrations in groundwater and surface water (if appropriate) will be modeled. Models to be used to determine future concentrations and activities at the source and in groundwater will be based on the modeling matrix presented in Table 3.2. Tier 1 values are existing sets of screening levels used for the initial screening of a site. Tier 2 values also are used for scoping, but account for more specific estimates of model parameters than the default Tier 1 values. Tiers 3 and 4 are models used with primarily site—specific values for site decision making.

Because all models contain significant uncertainty, the baseline risk assessment's analysis of off-site migration also will include risks calculated using current contaminant concentrations [i.e., data collected within the year preceding the model so that the data is representative of modeled conditions, if possible (e.g., if the model is created in 2015, then data collected in 2014 will be used)] at source units in addition to modeled values. This analysis will be included in the uncertainty section of all baseline risk assessments that contain modeling.

In baseline risk assessments for the integrator units, analyte degradation, attenuation, and transformation will be considered in addition to migration when calculating future concentrations, if possible. The analysis of these factors will rely upon the analysis presented in earlier sections of the RI report.

Table 3.2. Modeling Matrix for Groundwater, Surface Water, and Biota

	Values for Soil to Protect Groundwater	Model	Point of Exposure	Notes
JMENTS	Tier 1 (Used for scoping)	Soil Screening Levels (SSLs) and/or RESidual RADioactivity (RESRAD)	At source unit	Value to be used for initial scoping, use dilution attenuation factor (DAF) of 1 for SSLs unless site-specific values are available. Groundwater protection value based on residential use and targets of 1E-6, 0.1, and 1 for risk, hazard,
INVESTIGATION DOCUMENTS				and radiological dose, respectively. If site-specific DAF values are used, then need to justify these values. The depth of water needs to be considered in the calculation.
/SIL		Vapor intrusion model	At source unit (indoor or outdoor, as appropriate)	Appropriate initial vapor intrusion model will use default values.
NVES	Tier 2	Seasonal Soil Model (SESOIL) and/or	At source unit	Includes source delimitation.
	(Used for scoping)	RESRAD		Recognize SESOIL limitations when modeling inorganic COPCs-refine distribution coefficients (K _d s).
	Tier 3	SESOIL and RESRAD suite of codes (including	At source unit and at downgradient points	Uses source delimitation and refined K _d s from above.
	(Enhanced modeling used in decision	RESRAD-OFFSITE)		Use values from this effort to set initial cleanup levels.
	documents if needed)	with Analytical	(Industrialized area, DOE	O d T (d d chopp)
		Transient 1-, 2-, 3- Dimensional Simulation	property boundary, creek, river)	On the Terrace (southern portion of PGDP), different points of exposure will apply.
DECISION DOCUMENTS		of Waste Transport in the Aquifer System	inver)	different points of exposure win apply.
NOIS	Tier 4	Source modeling and three-dimensional	At source unit and at downgradient points	To be used to refine cleanup levels (if needed).
DECI	(Enhanced modeling used in decision and design documents if needed)	finite-difference groundwater model	appropriate to the selected remedy	May be especially important to set monitoring goals.
	- /	(MODFLOW/MT3D/ RT3D)	-	On the Terrace (southern portion of PGDP), different points of exposure will apply.

Table 3.2. Modeling Matrix for Groundwater, Surface Water, and Biota (Continued)

	Values for Soil and Sediment to Protect Surface Water	Model	Point of Exposure	Notes
ENTS	Tier 1 (Used for scoping)	SSLs and/or RESRAD	At source unit	Value to be used for initial scoping by Project Team. Use DAF of 1 for SSLs.
N DOCUMENTS	(Osed for scoping)			Groundwater protection value based on recreational use and targets of 1E-6, 0.1, and 1 for risk, hazard, and radiological dose, respectively.
ATIC				If site-specific DAF values are used, then need to justify these values.
STIG	Tier 2	Modified Universal Soil Loss Equation (MUSLE)	At source unit	Includes source delimitation. Value to be used during follow-up meetings by Project Team.
INVESTIGATION	(Used for scoping)			
S	Tier 3	Storm Water Management Model	At source unit and at downgradient points	Uses source delimitation from above.
UMENTS	(Enhanced modeling used in decision documents if needed)	(SWMM)	(Industrialized area, creek)	Initial cleanup level calculations.
DOC	Tier 4	Enhanced SWMM	At source unit and at downgradient points	To be used to refine cleanup levels (if needed).
DECISION	(Enhanced modeling used in decision and design documents if needed)		appropriate to the selected remedy	May be especially important to set monitoring goals.
DEC			(Industrialized area, creek)	

Table 3.2. Modeling Matrix for Groundwater, Surface Water, and Biota (Continued)

	Values for Soil and Sediment to Protect Biota	Model	Point of Exposure	Notes
ON DOCUMENTS	Tier 1	NONE	NONE	The RAWG determined that development of screening values based on biota modeling would not be appropriate; therefore, these values do not exist.
INVESTIGATION	Tier 2 (Used in Baseline Risk Assessments)	Those contained in current Methods Document, Appendix D	At source unit	Includes source delimitation.
DOCUMENTS	Tier 3 (Enhanced modeling used in Decision Documents if needed)	Those contained in current Methods Document, Appendix D for biota and transport models presented earlier for receiving media.	At source unit and at downgradient points (Industrialized area, creek)	Uses source delimitation from above. Initial cleanup level calculations.
DECISION 1	Tier 4 (Enhanced modeling used in Decision and Design Documents if needed)	Those contained in current Methods Document, Appendix D for biota and transport models presented earlier for receiving media.	At source unit and at downgradient points (Industrialized area, creek)	To be used to refine cleanup levels (if needed). May be especially important to set monitoring goals.

The equations to be used to combine the EPCs and exposure factors to estimate dose will follow the general format presented in RAGS, Part A (EPA 1989a). This general equation is shown in Equation 5. Specific equations are presented in Appendix D of this document. In this appendix, references are presented for each exposure parameter (e.g., CR, BW) included in the equation. Generally, these parameters were taken from guidance documents (e.g., EPA 1989a; KDEP 2002) unless site-specific values are available. (Equations used to derive radionuclide dose are similar to those presented in Appendix D.)

Intake =
$$C \times \frac{CR \times EFD}{BW} \times \frac{1}{AT}$$
 Eq. 5

where: Intake = The chemical dose $[mg/(kg \times day)]$

C = The average concentration contacted over the exposure period. See Eqs. 6 and 7 and associated discussion.

CR = The contact rate or amount of contaminated medium contacted per unit time or event.

EFD = The exposure frequency and duration describing how long and how often exposure occurs.

BW = The average body weight of the receptor over the term of exposure.

AT = The averaging time or period over which exposure is averaged.

In the material in Appendix D, equations that can be used to calculate the concentrations of COPCs in selected biota (e.g., vegetables, fish, game, and livestock) also are presented. Generally, for baseline human health risk assessments for source units inside the industrialized area at PGDP, concentrations of COPCs in biota will be estimated using these equations because biota sampling cannot be performed. (These biota are not present.)

For assessments for source units outside the industrialized area and for integrator unit baseline risk assessments, results from biota sampling may be available. In cases where this information is available, the EPC will be calculated using the methods presented earlier in this section. In cases where this information is not available, the equations presented in Appendix D will be used to estimate the concentrations in biota. (Note: Because concentrations in biota can differ markedly with time of sampling, tissue sampled, species sampled, age of animal, and other factors, the use of analytical results from biota sampling in the risk assessment also may give results that are very uncertain; therefore, the uncertainty in the results calculated using biota analytical results also will be considered completely.)

3.3.4.4 Consideration of vapor intrusion

Analysis of the exposure pathway for vapor intrusion due to volatile organic compound (VOC)-contaminated soils and groundwater will be evaluated on a project-specific basis, as needed. This potential exposure pathway often is considered in order to support possible future industrial missions within the PGDP industrialized area. Redevelopment with the potential for inhabited structures to be located in areas where VOC-contaminated groundwater and soil exist or have existed is considered a reasonable future use. Additionally, areas outside the industrialized area where volatile contaminants may be present (e.g., the Water Policy Area) may be considered.

OSWER Technical Guide for Assessing and Mitigating the VI Pathway from Subsurface Vapor Sources to Indoor Air (EPA 2015) provides technical and policy recommendations on determining if the vapor intrusion pathway poses an unacceptable risk to human health. VISLs can be used to evaluate site analytical data. VISLs are risk-based screening levels used to identify sites or buildings that may pose a health concern through the vapor intrusion pathway. The EPA VISL calculator is located on the Web site https://www.epa.gov/vaporintrusion/vapor-intrusion-screening-levels-visls. Please refer to Table E.10 in Appendix E for vapor intrusion risk information. At sites where subsurface concentrations of vapor-forming chemicals fall below VISLs, no further action or study is warranted as long as the site fulfills the conditions and assumptions of the generic conceptual model underlying the development of the VISLs. Evaluating

these conditions and assumptions requires "basic knowledge of the subsurface source of vapors (e.g., location, form, and extent of site-specific vapor-forming chemicals) and subsurface conditions (e.g., soil type in the vadose zone, depth to groundwater for groundwater sources), which are important elements of the CSM."

Exceeding a VISL generally suggests that unacceptable exposures might occur and that further evaluation of the vapor intrusion pathway is appropriate. Further evaluation could be a human health risk assessment conducted to determine whether the potential human health risk posed to building occupants by a complete vapor intrusion pathway are within or exceed acceptable levels of risk (i.e., EPA CERCLA risk range and Kentucky's target risk ¹⁶), consistent with EPA guidance. The primary purpose of this risk assessment is to provide risk managers with an understanding of the risks to human health posed by vapor intrusion under current and reasonably expected future conditions. Depending on building- and site-specific circumstances, an early action also could be considered. See Sections 3.3 and 7.8 of OSWER Publication 9200.2-154 for additional information on when it may be appropriate to implement mitigation of the vapor intrusion pathway as an early action even though all pertinent lines of evidence have not yet been completely developed.

3.3.4.5 Presentation of the results of the exposure assessment

Several figures and tables will be used to report the results of the exposure assessment in baseline human health risk assessments performed for PGDP. As noted earlier, conceptual site models for each unit, group of units, or area under investigation will be presented, and tables presenting exposure and risk information will be prepared. In addition, this section also will present a summary of the decisions made concerning the selection of pathways to be quantified for each unit or area under investigation; the representative (i.e., exposure point) concentration of COPCs in each medium, including biota; any chemical-specific values used in the calculations; and the daily intakes resulting from the application of the exposure equations.

The material appearing in this summary will be taken from the larger tables presented in the appendix to the risk assessment. Formats to present this summary information are in Exhibits 3.9–3.12.

Exhibit 3.9. Summary of Pathway Analysis in the Exposure Assessment

Potentially Exposed Population	Exposure route, medium, and exposure point ¹	Pathway selected? (yes/no)	Reason for pathway selection or dismissal ²
Time period ³			
Population 1 ⁴			
	Pathway 1		
	Pathway 2		
		•	
		•	
	Pathway N		

¹ Each of the pathways presented in this section will be included.

² A short statement drawn from the discussion in the text will be provided for the decision.

 16 EPA's generally acceptable risk range is 10^{-4} to 10^{-6} for carcinogenic risk and below the HI of 1 for noncarcinogens (EPA 1999b). Kentucky's target risk is defined as 10^{-6} (401 *KAR* 100:030).

³ Summary tables will be prepared for both the current or future time period. If multiple future time periods are assessed, a summary table will be included for each.

⁴ The populations include residential, recreational, industrial, and excavator. Only populations relevant to the time period will be included.

Exhibit 3.10. Presentation of Exposure Point Concentrations¹

COPC ²	Medium 1 ³	Medium 2	• • •	Medium N
Unit or Area 1 ⁴				
Analyte 1				
Analyte 2				
•				
		•		•
Analyte N				

A table will be made for each time period if models are used to estimate future representative concentrations.

Exhibit 3.11. Chemical-Specific Parameters

COPC ¹	Parameter 1 ²	Parameter 2	• • •	Parameter N
Analyte 1				
Analyte 2				
		•	•	•
		•	•	•
Analyte N			• • •	

All COPCs over all units or areas investigated will be presented. A separate list will not be presented for each unit unless unit-specific, chemical-specific parameters are used in the assessment.

Exhibit 3.12. Daily Intakes (Chronic Dose) for Receptor 1¹

COPC ²	Pathway 1 ³	Pathway 2	• • •	Pathway N
Unit or Area 1 ⁴				
Analyte 1				
Analyte 2				
	•	•	•	•
	•	•	•	•
Analyte N				

A separate table will be made for each receptor. If use patterns are assumed to differ between time periods, separate tables for each time period will also be provided.

3.3.4.6 Probabilistic risk assessment

Initially, all baseline risk assessments will be conducted as deterministic (point estimate) risk assessments. COPCs with high variability and uncertainty in exposure concentrations or for which individual exposure parameters greatly influence the risk or hazard estimate may be considered for PRAs. These assessments evaluate the variability and uncertainty in risk estimates, and are used to determine the likelihood of exceeding a risk level of concern. PRAs will be conducted following the guidance in *RAGS Volume III-Part A* (EPA 2001a). Scoping is an extremely important component of a PRA to determine which parameters should vary and to develop appropriate ranges of values for those parameters. Ranges of values for variables in the risk equations that were used in a previous PRA for the Southwest Plume are provided in Appendix E of this document. The values for variables listed in Appendix E are appropriate as

² All COPCs across all media will be presented for each unit or area.

³ All media will be listed. The order will be groundwater, soil, sediment, surface water, and biota if possible. More than one EPC may be derived for a media if different depths are used for exposures under different scenarios.

⁴ Each unit or area will be presented separately, but only one table will be used if possible.

² All chemical-specific parameters will be listed so that the calculations in the assessment can be duplicated by reviewers or users.

² COPCs across all media will be listed for each unit or area.

³ Each pathway included in the assessment will be listed. The order followed will be groundwater pathways, soil pathways, surface water pathways, sediment pathways, and biota pathways, if possible.

⁴A separate presentation will be made for each unit or area; however, only one table will be used if possible.

a starting point for other PRAs, but should be reviewed to ensure they are applicable to a specific project and modified if necessary. Documents using PRA also will need to include additional sections providing explanation of how the PRA was conducted, the interpretation of the results, and the appropriate application of the results to decision making to ensure that the PRA and its results are understandable to both the regulatory agencies and the public. Additional information regarding probabilistic risk assessment can be found in the references listed in Section 3.3.1.1.

3.3.5 Toxicity Assessment Methods

The primary purpose of this section of the baseline human health risk assessment will be to report the toxic effects of the COPCs on exposed populations. In addition, this section will briefly describe the methods used by EPA and in the toxicity assessment, to develop toxicity parameters, delineate the sources used to acquire the toxicity parameters, and present tables summarizing the toxicity information used in the risk assessment. In closing, this section will summarize the amount of toxicity information available on the COPCs in the risk assessment and discuss general toxicity assessment uncertainties. Requirements for each of these activities are discussed below.

3.3.5.1 Toxicity summaries

A toxicity summary for each COPC will be presented in the toxicity assessment. Each summary will contain a short description of the toxic effects of the chemical and the source of the toxicity values. Included in each description will be information on the effects associated with exposure to the chemical; the concentrations at which adverse effects are expected to occur in humans; a brief description of the database used to derive each toxicity value, including the particular study from which the toxicity value used in risk characterization was derived; and the approval status of any toxicity values. Each toxicity summary will conclude with a listing of the toxicity values used in the risk assessment for administered and absorbed dose routes of exposure.

3.3.5.2 Sources of toxicity information

The sources that will be used in developing toxicity information for risk assessments performed for PGDP are listed below. These will be examined in the order presented.

- Tier 1 sources: IRIS (EPA 2016c)
- Tier 2 sources: EPA Provisional Peer Reviewed Toxicity Values
- Tier 3 sources:
 - Health Effects Assessment Summary Tables (HEAST) (EPA 1997; 2001b)
 - Other sources identified in OSWER Directive 9285.7-53
 - Agency for Toxic Substances and Disease Registry toxicological profiles

When compiling toxicity information, provisional and withdrawn values and toxicity values withdrawn from IRIS or HEAST will be included, and provisional values will be clearly identified. If toxicity information is not available from the sources listed above, surrogate chemicals with toxicity values may be identified through consideration of chemical structure and characteristics. Selection of surrogate chemicals may be revised as toxicological information on surrogates becomes available, and will require consultation with and approval from EPA and KDEP. A list of currently approved surrogates for select chemicals is provided in the introduction of Appendix A.

Note: Toxicity values will not be developed for PGDP risk assessments without consultation with the regulatory agencies.

Three additional issues will be addressed when reporting the sources of toxicity information. These are the use of toxicity values for chronic versus subchronic effects, the calculation of toxicity values for absorbed versus administered dose, and the use of oral administered dose toxicity values for the inhalation exposure route. Each of these is discussed herein.

Generally, all risk assessments performed for PGDP will only use toxicity values for chronic exposure when characterizing risk. Although RAGS, Part A, (EPA 1989a) states that toxicity values for subchronic exposure should be used for exposure durations (EDs) less than seven years in length, these will not be used because they are not available for many chemicals (in which case the chronic value should be used). The receptor groups that are affected by this decision are the child rural resident, the recreational user, and the outdoor worker. In no case will toxicity values based on subchronic exposure be used for child or teen receptors. For outdoor workers, toxicity values based in subchronic exposure may be used if the information provided by their use is beneficial in remedial action decision making.

To properly characterize risk from absorbed dose (e.g., dose from dermal absorption across the skin), it is necessary to have toxicity values that are based on absorbed dose. Generally, all toxicity values in IRIS and HEAST are based on administered dose and cannot be used directly with the chronic daily absorbed doses calculated using the exposure equations in Appendix D. To convert administered dose toxicity values to absorbed dose toxicity values, the guidance provided in Risk Assessment Guidance for Superfund, Volume I: Human Health Evaluation Manual. Supplemental Guidance, Dermal Risk Assessment, Interim Guidance (EPA 1992b) will be used. The method delineated in this guidance is depicted in Eqs. 6 and 7. Equation 6 shows that the administered dose toxicity value (absorbed dose slope factor) by dividing by the chemical-specific gastrointestinal absorption efficiency of the respective chemical or compound. Equation 7 shows that the administered dose toxicity value for systemic toxicity [administered reference dose (RfD)] are converted to an absorbed dose toxicity value (absorbed RfD) by multiplying by the chemical-specific gastrointestinal absorption efficiency of the respective chemical or compound.

As stated in RAGS Part E (EPA 2004):

For those organic chemicals that do not appear on the table, the recommendation is to assume a 100% ABS_{GI} value, based on review of literature, indicating that organic chemicals are generally well absorbed (>50%) across the GI tract. Absorption data for inorganics are also provided in Exhibit 4-1 [see text box], indicating a wide range of absorption values for inorganics. Despite the wide range of absorption values for inorganics, the recommendation is to assume a 100% ABS_{GI} value for inorganics that do not appear in this table. This assumption may contribute to an underestimation of risk for those inorganics that are actually poorly absorbed. The extent of this underestimation is inversely proportional to the actual GI absorption.

Absorbed
$$SF = \frac{Administered SF}{GI Efficiency}$$
 Eq. 6

where: Absorbed SF = The absorbed dose slope factor for cancer effects
Administered SF = The administered dose slope factor for cancer effects
GI Efficiency = The chemical-specific gastrointestinal absorption efficiency

Absorbed
$$RfD = Administered RfD \times GI Efficiency$$
 Eq. 7

where: Absorbed RfD = The absorbed reference dose for systemic toxicity
Administered RfD = The administered reference dose for systemic toxicity
GI Efficiency = The chemical-specific gastrointestinal absorption efficiency

The dermal dose derived with this methodology provides an estimate of the contribution of the dermal pathway to the systemic dose. Dermal exposure for baseline risk assessments will follow the Risk Assessment Guidance for Superfund Volume I: Human Health Evaluation Manual (Part E, Supplemental Guidance for Dermal Risk Assessment) (EPA 2004c). The EPA guidance provides specific values for eleven compounds or groups of compounds in Exhibit 3-4 of the dermal guidance. For an alternative evaluation, providing more restrictive values for the dermal-soil pathway, the default values of 25% dermal absorption for VOCs, 10% dermal absorption from soil for all semivolatile organic compounds without specific absorption values specified in RAGS, Part E: and 5% dermal absorption from soil for all inorganic compounds without specific absorption specified in RAGS, Part E, may (based on project-specific goals) be applied to a quantitative risk assessment. This approach is consistent with guidance from KDEP. KDEP-specific values for dermal absorption are provided in Appendix B. See text box for additional information. For the dermalwater pathway, absorption should be calculated using the

In RAGS Part E 2004, Exhibit 4-1, the following gastrointestinal (GI) absorption efficiencies are listed that are below the 5% dermal absorption KDEP has recommended as a default value for inorganics. For these constituents, the dermal absorption value should be modified from 5% to mimic the GI absorption efficiencies, as follows:

Beryllium	0.007 = 0.7%
Chromium III	0.013 = 1.3%
Chromium VI	0.025 = 2.5%
Manganese	0.04 = 4%
Nickel	0.04 = 4%
Silver	0.04 = 4%
Vanadium	0.026 = 2.6%

This is in addition to the chemical-specific dermal absorption fractions listed in RAGS Part E Exhibit 3-4, including:

Arsenic	0.03 = 3%
Cadmium	0.001 = 0.1%

methods described in RAGS, Part E. For inorganic chemicals, the K_p (permeability coefficient) parameter has been identified as one of the major parameters contributing to uncertainty in the assessment of dermal exposures to contaminants in aqueous media. The EPA guidance recommends the use of predicted K_p values. For chemicals that fall outside the Effective Prediction Domain for determining K_p , a fraction-absorbed term should be applied. This Risk Methods Document recommends the EPA default exposure values for all variables for the dermal-water and dermal-soil pathways. These include the residential scenario for water exposure and residential and industrial for soil exposure. For dermal-water exposures, the entire skin surface area is assumed to be available for exposure when bathing and swimming occurs, but the surface area available for a wading scenario includes the portions of the body specified in Appendix D for the dermal equations. Default values for the soil adherence factor also are provided with the equations in Appendix D. The guidance does not include a method for assessing dermal absorption of chemicals in the vapor phase, with the assumption that inhalation will be the major exposure route for vapors.

3.3.5.3 Tables summarizing the toxicity information

To facilitate review of the toxicity assessment, summary tables of toxicity information will be prepared following the examples in the previous sections of this guidance document. Additional tables may be prepared for the main body of the risk assessment, if needed to clarify the toxicity assessment process.

3.3.5.4 Summary of toxicity information available on the COPCs

This section of the toxicity assessment will provide a listing of the chemical classes and the number of chemicals within each class that have toxicity information ordered by medium within the unit or area under investigation. This summary will be presented to illustrate the total amount of toxicity information available to characterize risk in the following section.

3.3.6 Risk Characterization Methods

The primary purpose of this section of the baseline human health risk assessment will be to integrate the dose information developed in the exposure assessment with the effects information presented in the

toxicity assessment to characterize the risk and hazard posed by environmental contamination at PGDP. In this section, the methods used to integrate the information to characterize risk and hazard and the tables and narrative summarizing the risk characterization for each exposure unit under each current and potential future use scenario will be presented. This section will conclude with a listing of use scenarios of concern for each location and a listing of COCs, POCs, and MOCs for each use scenario of concern.

3.3.6.1 Methods used to integrate dose and toxicity

In all baseline human health risk assessments performed for PGDP, the methods outlined in RAGS, Part A, will be used to integrate dose and toxicity information and characterize risk. To characterize risk from inhaled contaminants, the methods outlined in RAGS, Part F will be followed (EPA 2009). The following presents the equations that will be used for these calculations and describes the result of each equation. Note: In this presentation, the calculations for systemic toxicity (i.e., hazard) and cancer risk are presented separately because they differ slightly. Also, note that the values for systemic toxicity are estimates of whether the daily doses from each COPC, from each exposure pathway, and over all pathways and COPCs exceed that which may result in toxic effects in the receptor. However, the values for cancer risk are estimates of the excess cancer incidence that may result from exposure to each COPC, from each exposure pathway, and over all pathways.

Equations 8, 9, and 10 will be used to characterize the potential for systemic toxicity in all baseline human health risk assessments performed for PGDP. The result of Eq. 8 (Eq. 8a for inhalation) is a numeric estimate of the potential for systemic toxicity posed by a single chemical within a single pathway of exposure. The result of Eq. 9 is a numeric estimate of the potential for systemic toxicity posed by all chemicals reaching a receptor through a single pathway. The result of Eq. 10 is a numeric estimate of the potential for systemic toxicity posed to a receptor by exposure to all chemicals over all pathways. (This last value is often called an estimate of "total noncarcinogenic risk.")

$$HQ_i = \frac{CDI_i}{RfD_i}$$
 Eq. 8

where: HQ_i = The hazard quotient, an estimate of the systemic toxicity posed by a single chemical

 CDI_i = The estimate of chronic daily intake (or absorbed dose for some exposure routes) from the exposure assessment (calculated from the chemical intake equations in Appendix D)

 RfD_i = The chronic reference dose for administered or absorbed dose, as appropriate

$$HQ = \frac{EC (\mu g/m^3)}{[RfC (mg/m^3) \times 1,000 (\mu g/mg)]}$$
Eq. 8a

where: HQ_i = The hazard quotient, an estimate of the systemic toxicity posed by a single chemical for inhalation

 EC_i = The exposure concentration for chronic exposure (calculated from the equations in Appendix D)

 RfC_i = The reference concentration for chronic inhalation exposure

$$HI_p = \sum_{i=1}^n HQ_i$$
 Eq. 9

where:

 HI_p = The pathway hazard index, an estimate of the systemic

toxicity posed by all chemicals within a single pathway

 HQ_i = The individual chemical hazard quotients for chemicals reaching the receptor through a single pathway (from Eq. 8 or Eq. 8a)

$$HI_{total} = \sum_{p=1}^{n} HI_{p}$$
 Eq. 10

where: HI_{total} = The total hazard index, an estimate of the systemic toxicity posed by all chemicals over all pathways HI_p = The pathway hazard indices from Eq. 9

Equations 11, 12, and 13 will be used to characterize the potential excess lifetime cancer incidence (i.e., ELCR) in all baseline human health risk assessments performed for PGDP. The result of Eq. 11 (Eq. 11a for inhalation) is an estimate of the increased cancer incidence (i.e., a probability) to a receptor that results from exposure to a single chemical (or radionuclide) within a single pathway for chemicals without identified mutagenic effects. For chemicals with mutagenic effects, the equation may be modified through use of age-dependent adjustment factors combined with age-specific exposure estimates to assess cancer risks, consistent with EPA guidance (EPA 2005b). The result of Eq. 12 is an estimate of the increased cancer incidence (i.e., probability) that results from exposure to all chemicals (or radionuclides) reaching a receptor through a single pathway. The result of Eq. 13 is an estimate of the increased cancer incidence (i.e., probability) that results from exposure to all chemicals (or radionuclides) reaching a receptor over all pathways. (This last value is often called an estimate of "total carcinogenic risk.")

$$ELCR_i = CDI_i \times SF_i$$
 Eq. 11

where: $ELCR_i$ = The chemical-specific excess cancer incidence

 CDI_i = The estimate of chronic daily intake (or absorbed dose) from the exposure assessment (calculated from the chemical intake equations in Appendix D)

 SF_i = The slope factor for administered or absorbed dose, as appropriate

ELCR = EC
$$(\mu g/m3) \times IUR (\mu g/m3)^{-1}$$
 Eq. 11a

where: $ELCR_i$ = The chemical-specific excess cancer incidence

 EC_i = The exposure concentration for chronic exposure (calculated from the equations in Appendix D)

 IUR_i = The unit risk for chronic inhalation exposure

$$ELCR_p = \sum_{i=1}^{n} ELCR_i$$
 Eq. 12

where: $ELCR_p$ = The pathway-specific excess cancer incidence

ELCRi = The chemical-specific excess cancer incidence from Eq. 11 or Eq. 11a

$$ELCR_{total} = \sum_{p=1}^{n} ELCR_{p}$$
 Eq. 13

where: ELCR_{total} = The total excess cancer incidence posed by all chemicals over all pathways

ELCR_p = The pathway-specific excess cancer incidence from Eq. 12

3.3.6.2 Presentation of risk characterization

In the baseline human health risk assessment, risk will be characterized for each exposure unit under each current and potential future use scenario. The results of the characterization will be presented in both tables and as narrative. The tables that will be used for each time, exposure unit, and receptor combination will be consistent with the two-way table presented in RAGS, Part D (EPA 1998b). At this time, scenarios are evaluated independently. Scenarios may be combined if it is determined that it is appropriate to do so to

represent cumulative risk on a site-specific basis. The exact format presented in RAGS Part D is not used for the PGDP risk characterization tables because the FFA team discussed table presentation and agreed that the tables presented in this guidance document are adequate to meet the intent of RAGS, Part D. The narrative that explains this table, which may include summary tables, will present the exposure unit; the receptor, HI_{total} (from Equation 10) or ELCR_{total} (from Equation 13); the primary pathways contributing to HI_{total} or ELCR_{total} (i.e., "driving pathways"); and the primary chemicals contributing to HI_{total} or ELCR_{total} (i.e., "driving chemicals"). An example of a narrative description of risk taken from DOE 1996e is presented below.

Exhibit 3.13 summarizes the HIs for exposure routes for the current industrial worker over all locations. As shown in this exhibit, the total scenario HI (i.e., Location Total in Exhibit 3.13) is greater than 1 for Sectors 5, 6, and 9. For each location, the driving exposure route is dermal contact with soil, which accounts for more than 95% of the total HI. Also, for each location, the inhalation exposure route contributes insignificantly to the location total HI.

Exhibit 3.13. Exposure Route Summary for the Current Use Scenario—Systemic Toxicity*

Scenario and	·	·				
Location	Incidental Ingestion	Exposure Routes for Soil Incidental Ingestion Dermal Contact Inhalation of Vapors/Particles				
		Current industria	l worker			
Sector 1	N/A	N/A	N/A	NIX7		
% of Total	NV	NV	NV	NV		
Sector 2	< 0.1	0.4	NV	0.4		
% of Total	1%	99%	NV	0.4		
Sector 3	< 0.1	0.3	< 0.1	0.2		
% of Total	2%	98%	< 1%	0.3		
Sector 4	< 0.1	1.0	< 0.1	1.0		
% of Total	1%	99%	< 1%	1.0		
Sector 5	< 0.1	1.7	< 0.1	1.0		
% of Total	2%	98%	< 1%	1.8		
Sector 6	< 0.1	1.2	< 0.1	1.2		
% of Total	5%	95%	< 1%	1.2		
Sector 8	< 0.1	1.0	< 0.1	1.0		
% of Total	< 1%	99%	< 1%	1.0		
Sector 9	< 0.1	1.3	NV	1.2		
% of Total	1%	99%	NV	1.3		

N/A indicates that the scenario is not applicable for this location.

Exhibit 3.14 summarizes the contaminants contributing more than 1% of the total systemic toxicity for the current industrial worker over all locations for those locations where the total systemic toxicity for the location exceeds 1. As shown in this exhibit, in each case, metals are the primary driving contaminants; however, PCBs and PAHs are minor contributors for Sector 6.

Exhibit 3.14. Driving Contaminants Summary for Current Use Scenario—Systemic Toxicity

Scenario and Location	Driving Contaminants Over All Exposure Routes	Location Total
	Current industrial worker	
Sector 1	HI < 1	NV
Sector 2	HI < 1	0.4

NV indicates that a value is not available.

^{*}Current convention is to use one significant digit for presentation of hazard indices. Two significant digits are used here when the hazard index is greater than 1 to enable the reader to match the numbers reported in the exhibit with those in its associated risk characterization table. Additionally, use of two significant digits, when the exposure route's value is greater than 1, allows the reader to sum the route values and check the location total.

Exhibit 3.14. Driving Contaminants Summary for Current Use Scenario— Systemic Toxicity (Continued)

Scenario and Location	Driving Contaminants Over All Exposure Routes	Location Total
Sector 3	HI < 1	0.3
Sector 4	HI < 1	1.0
Sector 5	iron (47%); chromium (26%); antimony (22%); uranium (3%)	1.8
Sector 6	chromium (22%); antimony (22%); arsenic (20%); PCB (13%);	
	aluminum (13%); pyrene (2%); fluoranthene (1%)	1.2
Sector 8	HI < 1	1.0
Sector 9	antimony (58%); aluminum (23%); chromium (17%); uranium (2%)	1.3

N/A indicates that the scenario is not applicable for this location.

NV indicates that a value is not available.

HI < 1 indicates that total scenario hazard index is less than 1; therefore, analytes are not listed.

In the tables prepared for risk characterization, all COPCs will be listed, even those that do not have a value. Also, these tables will present the total chemical-specific hazard (or total chemical-specific risk) over all pathways, the total pathway-specific hazard (or risk) over all chemicals, the total hazard or risk over all pathways and chemicals, and the total risk and hazard over all media within the exposure unit (consistent with the Conceptual Site Model).

3.3.6.3 Risk characterization for lead

Risk characterization for lead is a special case. Although it is known that exposure to lead can result in systemic toxic effects and possibly cancer, the approved toxicity values required to estimate potential for systemic toxicity and carcinogenesis are not available. The risk characterization for lead will consist of a comparison of the maximum detected concentration from the site/source to the no action screening levels from EPA and the Commonwealth of Kentucky. The no action screening levels are 400 mg/kg in soil and sediment for the residential and recreational scenarios, 800 mg/kg in soil and sediment for the industrial, and outdoor worker scenarios) and 15 µg/L in groundwater and surface water for all scenarios (residential, recreational, industrial, and outdoor worker). Sites with lead concentrations exceeding these levels will undergo additional analysis for risk using the results of EPA's IEUBK (EPA 2021a; EPA 2021b) for evaluating residential and recreational exposures of children and the results of the EPA Adult Lead Model (EPA 2003a, 2017) for evaluating industrial and outdoor worker exposures. The parameters for use in each of these models are presented in Appendix B. Screening values for lead appear in the tables presented in Appendix A. EPA has published *Recommendations for Sieving Soil and Dust Samples at Lead Sites for Assessment of Incidental Ingestion* (EPA 2016a). Historical data not sampled in accordance with EPA's recommendation should be evaluated as uncertain.

3.3.6.4 Selection of use scenarios, POCs, COCs, and MOCs

Use scenarios, pathways, contaminants, and MOC will be identified for each unit or area under investigation. If any unit or area is divided into exposure units during the exposure assessment, use scenarios, pathways, contaminants, and MOC will be identified for each exposure unit.

In identifying use scenarios, pathways, contaminants, and MOC, specific rules will be followed as discussed below.

• Identification of use scenarios of concern. To determine use scenarios of concern or the basis of risk, risk characterization results for total systemic toxicity (HI_{total}) and total risk ($ELCR_{total}$) will be compared to benchmarks of HI = 1.0 and $ELCR = 1 \times 10^{-6}$. Use scenarios with HI_{total} or $ELCR_{total}$ exceeding either of these benchmarks will be deemed use scenarios of concern. Note: The results in the example narrative provided in Section 3.3.6.2 indicate the teen recreational use scenario is a use

scenario of concern for SWMU 8a ($HI_{total} = 71.5$). This value would be found in the lower right hand corner of a two-way table consistent with RAGS, Part D (EPA 1998b).

- **Identification of POCs.** To determine POCs, risk characterization results for pathway hazard (HI_p) and risk (ELCR_p) over all chemicals within a use scenario of concern will be compared to benchmarks of HI = 0.1 and ELCR = 1×10^{-6} . Pathways within a use scenario of concern exceeding either of these benchmarks will be deemed POCs for the use scenario of concern. Note: The results in the example narrative provided in Section 3.3.6.2 indicate that the POCs for the teen recreational user are dermal contact with surface water ($HI_p = 2.0$), dermal contact with leachate ($HI_p = 0.6$), ingestion of fish ($HI_p = 60.5$), ingestion of sediment ($HI_p = 0.1$), dermal contact with sediment ($HI_p = 8.2$), and ingestion of venison ($HI_p = 0.2$). These values would be found along the bottom margin of a two-way table consistent with RAGS, Part D (EPA 1998b).
- **Identification of COCs.** To determine COCs, risk characterization results for chemical hazard (HQ_i) and risk (ELCR_i) over all pathways *within a use scenario of concern* will be compared to benchmarks of HQ = 0.1 and ELCR = 1×10^{-6} . Chemicals of potential concern within a use scenario of concern exceeding either of these benchmarks will be deemed COCs for the use scenario of concern. [Note: For dioxins and furans, carcinogenic PAHs, and PCBs, the total risk over all congeners (for dioxins and furans) or compounds (for carcinogenic PAHs and PCBs) will be used when determining if these are COCs.] The results in the example narrative provided in Section 3.3.6.2 indicates that the COCs for the teen recreational user are aluminum ($HQ_i = 0.2$), antimony ($HQ_i = 6.1$), arsenic ($HQ_i = 0.2$), cadmium ($HQ_i = 0.6$), iron ($HQ_i = 9.4$), manganese ($HQ_i = 48.4$), strontium ($HQ_i = 0.1$), vanadium ($HQ_i = 4.7$), and zinc ($HQ_i = 1.7$). These values would be found along the right margin of a two-way table.
- Identification of Priority COCs. To determine priority COCs (i.e., those COCs contributing most to cumulative HI and ELCR), risk characterization results for chemical hazard (HQ_i) and risk (ELCR_i) over all pathways within a use scenario of concern will be compared to benchmarks of HQ = 1 and ELCR = 1 × 10⁻⁴. COCs exceeding either of these benchmarks will be deemed priority COCs for the use scenario of concern. [Note: For dioxins and furans, carcinogenic PAHs, and PCBs, the total risk over all congeners (for dioxins and furans) or compounds (for carcinogenic PAHs and PCBs) will be used when determining if these chemicals are priority COCs.]
- **Identification of MOCs.** To determine MOCs, the POCs are reviewed, and those media in these pathways are deemed to be MOC. This is equivalent to screening the total risk and hazard posed by COPCs in the various media against benchmarks of HI = 0.1 and ELCR = 1×10^{-6} . For the results presented in the example narrative in Section 3.3.6.2, the MOCs are surface water, leachate, fish, sediment, and venison.
- Identification of scenarios of concern, POCs, COCs, and MOCs in Radiological Dose Assessment. If a radiological dose assessment is conducted to provide additional information to risk managers, a scenario of concern will be one that has a total radiological dose exceeding the PGDP *de minimis* radiological dose of 1 mrem/year. A COC will be one that has a contaminant-specific radiological dose exceeding 1 mrem/year. A POC will be an exposure route that has a route-specific radiological dose exceeding 1 mrem/year. An MOC will be those media appearing in any POC.

3.3.6.5 Consideration of COPCs for which risk cannot be estimated

For some COPCs, information is insufficient for risk characterization. Generally, risk cannot be characterized for these chemicals because toxicity values are not available. When this occurs in risk assessments performed for PGDP, these COPCs will be deemed COCs during risk characterization, and they will be reported along with the COCs chosen by the rules outlined above.

3.3.6.6 Summary of risk characterization

To provide a summary of risk characterization for each unit or area under investigation, a table will be prepared and included as a summary of risk characterization in all baseline human health risk assessments. This table will follow the format shown in Exhibit 3.15 and list the risk and hazard posed within each use scenario of concern, the percent contribution of each POC to HI_{total} and ELCR_{total}, and the percent contribution of each COC to HI_{total} and ELCR_{total}. A similar table will be prepared to summarize the results of the radiological dose assessment if a radiological dose assessment is conducted for the site.

3.3.7 Consideration of Uncertainty in the Risk Assessment

Uncertainties are associated with each of the steps of the baseline risk assessment. Following a general discussion of uncertainties in risk assessment, this section presents the uncertainties that will be addressed in baseline human health risk assessments prepared for PGDP and provides a format for summarizing this information (when a qualitative uncertainty analysis or sensitivity analysis is performed).

Exhibit 3.15. Summary of Risk Characterization

Use Scenario ¹	Total ELCR ²	COCs ³	% Total ELCR ⁴	POCs ⁵	% Total ELCR ⁶	Total HI ⁷	COCs	% Total HI ⁸	POCs	% Total HI ⁹
# 1										
# 2										
							•	•		•
	•		ě	•	·		•	•	•	•
					•		•			
# N										

¹ All use scenarios will be listed.

The potential effect of the uncertainties on the final risk characterization must be considered when interpreting the results of the risk characterization because the uncertainties directly affect the final risk estimates. Types of uncertainties that must be considered can be divided into four broad categories. These are uncertainties associated with data and data evaluation (i.e., identification of COPCs); exposure assessment; toxicity assessment; and risk characterization. Specific uncertainties under each of these broad categories that will be addressed in baseline human health risk assessments completed for PGDP are listed in the following material.

The exact method that will be used to present the uncertainty analysis in all baseline risk assessments cannot be included here. This is due, in large part, to the fact that the rigor of the uncertainty analysis will depend on the unit or area under investigation, the decisions that must be made for the unit or area, and the uncertainties affecting the risk estimates. At minimum, all baseline risk assessments will contain a qualitative uncertainty analysis that will include a quantitative sensitivity analysis of salient uncertainties. In the qualitative uncertainty analysis, the magnitude of the uncertainty on the risk characterization will be categorized as small, moderate, or large. Uncertainties categorized as small will be those that should not cause the risk estimates to vary by more than one order of magnitude; uncertainties categorized as moderate will be those that may cause the risk estimates to vary by between one and two orders of magnitude; and,

² These values will be those found at the lower right of each unit's two-way table for the scenario of interest.

³ These constituents will be the COCs selected applying the rules listed earlier.

⁴ This value will be calculated by dividing the chemical-specific ELCR (ELCR_i) by the total ELCR (ELCR_{total}).

⁵ These pathways will be the POCs selected applying the rules listed earlier.

⁶ This value will be calculated by dividing the pathway-specific ELCR (ELCR_p) by the total ELCR (ELCR_{total}).

⁷ These values will be those found at the lower right of each unit's two-way table for the scenario of interest.

⁸ This value will be calculated by dividing the chemical-specific hazard quotient (HQ_i) by the total HI (HI_{total}).

⁹ This value will be calculated by dividing the pathway-specific HI (HI_p) by the total HI (HI_{total}).

uncertainties categorized as large will be those that may cause the risk estimates to vary by more than two orders of magnitude.

In the qualitative uncertainty analysis, a note will be made that the uncertainties listed and evaluated are neither independent nor mutually exclusive. It also will be noted that the total effect of all uncertainties upon the risk estimates is not the sum of the estimated effects of each uncertainty evaluated.

3.3.7.1 Uncertainties in data, data evaluation, and identification of COPCs

- Retention of common laboratory contaminants in the list of COPC
- Retention of infrequently detected analytes (i.e., detected in less than 10% of the samples analyzed) in the list of COPCs
- Lack of consideration in temporal patterns when selecting COPCs
- Spatial distribution and number of sampling locations (representativeness)
- Quantitation limits for some analytes exceeding their respective human health risk-based screening criteria (i.e., PRGs)
- Use of historical data ¹⁷ in addition to data collected as part of the RI field investigation
- Removal of analytes from the list of COPCs on the basis of a comparison to background concentrations
- Removal of analytes from the list of COPCs on the basis of comparison to concentrations found in associated blanks
- Removal of analytes from the list of COPCs on the basis of a toxicity screen
- Characterization of EPCs for environmental media under current conditions, including EPCs that are greater than maximum detected values
- Consideration of temporal changes in analyte concentrations and activities
- Use of results from analyses of unfiltered groundwater samples versus filtered groundwater samples
- Use of results from analyses of unfiltered surface water samples versus filtered surface water samples

3.3.7.2 Uncertainties in exposure assessment

• Incorporation of biota fate and transport modeling into risk and hazard estimates (if this type of modeling were performed)

• Uncertainties in modeled concentrations, including the consideration of solubility as defined by differences between contaminant concentrations in filtered and unfiltered water samples

¹⁷ This uncertainty includes use of historical data with qualifiers, as described in Step 4 of Section 3.3.3.2.

- Use of reasonable maximum exposure parameters versus average parameters for all exposure routes and associated pathways
- General issues in the development of conceptual site models
- Consideration of scenarios that involve livestock, dairy, and/or game animals that are potentially exposed to contaminated feed, soil, or water 18
- Consideration of scenarios that involve homegrown vegetables potentially grown in contaminated soil or irrigated with contaminated water¹⁹
- Use of default values from KDEP 2002 when estimating dermal absorbed dose (especially from soil and sediment)
- Difference in gamma walkover survey results and associated laboratory analyses
- Difference in calculation due to use of significant figures

3.3.7.3 Uncertainties in toxicity assessment

- Use of provisional or withdrawn toxicity values
- Difference in risk estimates for trichloroethene (TCE) based on use of KDEP oral slope factor and EPA TCE oral slope factor
- Extrapolation of oral administered dose toxicity values to inhalation dose toxicity values
- Derivation of absorbed dose toxicity values from oral administered dose toxicity values
- Lack of toxicity information, toxicity values, or both for some COPCs
- Use of chronic exposure toxicity values for exposures that are subchronic

3.3.7.4 Uncertainties in risk characterization

- Combination of chemical-specific risk and hazard estimates (ELCR_i and HQ_i, respectively) to derive pathway-specific and use scenario risk and hazard estimates (ELCR_p and ELCR_{total} and HI_p and HI_{total}, respectively) (i.e., effect of chemical mixtures)
- Using mutagenic effects for risk characterization
- Combination of risk estimates from chemical and radioisotope exposure
- Summing cancer risks across pathways and across target organs
- Evaluating presence or absence of Chromium VI when analyte-specific analyses are not available

¹⁸ In addition to the exposure parameters presented herein, *Exposure Factors Handbook 2011 Edition (Final Report)*, and EPA's Radionuclide PRG calculator also contain information on this subject (EPA 2011, EPA 2022c).

¹⁹ See footnote 18.

• Summation of risk and hazard across units or areas under investigation

(Note: Uncertainties regarding the risk characterization are discussed in the accompanying text box.)

3.3.7.5 Summary of qualitative uncertainty analysis

Because uncertainties in the baseline risk assessment must be addressed when screening potential remedial actions, developing revised preliminary remedial goals from RGOs and selecting the final action, the effect of all uncertainties on the risk and hazard estimates will be summarized in a single table. Note: Exhibit 3.16, is most useful when summarizing a qualitative uncertainty analysis; other formats may be used for a quantitative uncertainty analysis.

In addition to the summary table, a narrative (i.e., an Observations section) discussing the joint effects of the various uncertainties on the risk characterization results will be prepared. The overall goal of the narrative will be to focus the list of COCs to those COCs that contribute significantly to the risk and for which the risk estimate or the revised risk estimate in the uncertainty analysis is believed to reasonably reflect the risks posed to receptors under the most likely future use. This narrative in the Observations section will discuss how uncertainties affect the identification of COCs and evaluate scenarios that reflect the most likely future exposure. It also will describe how the inclusion of certain pathways (dermal, food ingestion, etc.) may lead to an overestimate of risks and summarizes which contaminants and/or pathways exceed de minimis levels. The narrative address each of the COCs individually.

Uncertainty in Combining Chemical-Specific Risk and Hazard Estimates and Pathway-Specific Risk and Hazard Estimates

One uncertainty in the risk characterization guidance contained in this document is the method used to combine HQs and chemical-specific ELCRs across pathways and to combine pathway HIs and ELCRs to calculate total HI and ELCR. The method to be used to calculate pathway HIs and ELCRs follows EPA protocols (EPA 1989a). This method calls for the simple addition of HQs and chemical-specific ELCRs to calculate pathway HIs and ELCRs, respectively, and assumes that all effects between chemicals are additive. As explained in EPA 1989a, this assumption is made because information concerning the effects of chemical mixtures is lacking.

The following limitations of this approach for systemic toxicity effects are reported by EPA:

- Little is known about the effects of chemical mixtures; although additivity is assumed, the interaction of multiple chemicals could possibly be synergistic or antagonistic.
- The RfDs and RfCs do not have equal accuracy or precision and are not based on the same severity of effects.
- Dose additivity is most properly applied to compounds that induce the same effect by
 the same mechanism of action. While the approach recommended by EPA is a useful
 screening-level approach, the cumulative systemic toxicity could be overestimated for
 chemicals that act by different mechanisms and/or on different target organs.

The following limitations of this approach for chemical carcinogenesis are reported by EPA:

- Cancer risks (i.e., ELCRs) are based on slope factors that represent an upper 95th
 percentile estimate of potency; the upper 95th percentiles of probability distributions
 are not strictly additive. Summing these risks can result in an overly conservative
 estimate of lifetime ELCR.
- Cancer risks may not be additive. By analogy to systemic toxicity effects, the endpoints may differ, and mechanisms of effect may vary.
- Not all slope factors contain the same weight-of-evidence for human carcinogenicity.
 EPA recognizes this by placing weight-of-evidence classifications on all slope factors.
 Those contaminants with a weight-of-evidence classification of A should probably receive more attention in the selection of a remedial design than contaminants with a B or C classification. Similarly, a contaminant with a B classification should probably receive greater attention than one with a C classification. The simple combination of ELCRs does not take this hierarchy into account.

Uncertainty in Combining Risk Estimated for Chemical Exposure to Those for Risk Estimated for Radioisotope Exposure

Uncertainty associated with adding risks from chemical exposure to those from exposure to radionuclides arises from two sources. First, the slope factors used to characterize the risk from chemicals are derived differently from the slope factors used to characterize risk from radionuclides. This difference results in estimates of chemical exposure risks that may be considered to be upper-bound risk estimates and estimates of radionuclide exposure risks that may be considered to be central tendency (i.e., "best") estimates; therefore, combining chemical exposure and radionuclide exposure risk estimates to estimate total risk for a land use scenario may place too much emphasis on chemical exposure risk. Second, the mechanism by which chemicals may cause cancer varies from the mechanism by which radionuclides may cause cancer. This difference in mechanism of action inflates the uncertainties that assume cancer risks are additive.

Exhibit 3.16. Summary of Uncertainty Analysis

	Estimated Effect ¹							
Description of Uncertainty	Small	Moderate	Large					
Uncertainties related to data, data evaluation, and identification of chemicals of potential concern								
Data uncertainty 1								
Data uncertainty 2								
•	•	•	•					
	•	•	•					
Data uncertainty n								

¹ Definitions of effects are as follows:

- Small—Uncertainty should not cause the risk or hazard estimate to vary by more than one order of magnitude;
- Moderate—Uncertainty may cause the risk or hazard estimate to vary by between one and two orders of magnitude;
- Large—Uncertainty may cause the risk or hazard estimate to vary by more than two orders of magnitude.

3.3.8 Remedial Goal Option Derivation Methods

This section of the baseline human health risk assessment will delineate the methods used to derive and present RGOs. It is important to note that RGOs are not cleanup levels, but are site-specific, risk- or radiological dose-based criteria that may be used to guide the development of revised PRGs in the FS and cleanup levels in the Record of Decision (ROD) by risk managers. Cleanup levels are developed as part of the risk analysis in the ROD (EPA 2018a).

3.3.8.1 Calculation of remedial goal options

Guidance in EPA (2018) directs that multiple RGOs must be calculated for each COC identified in a baseline human health risk assessment. To do this, the goals are calculated by rearranging the exposure equations quantified in the risk assessment so that they solve for a concentration or activity concentration in a medium that results in a specific "target risk," "target hazard," or "target radiological dose." Target risks that will be used to derive RGOs at PGDP are 1×10^{-4} , 1×10^{-5} , and 1×10^{-6} . Target hazards that will be used to derive RGOs are 3, 1, and 0.1. Target radiological doses for all media except groundwater are 1, 12, 25, and 100 mrem/year. For groundwater, the radiological dose targets are 1, 4 (for beta and photon emitters), 12, 25, and 100 mrem/year. As noted above, an RGO must be developed for each COC. Because the selection of a COC is medium- and use scenario-specific, RGOs will be developed for each COC identified for each use scenario of concern at a unit or area. Also, because RGOs must be medium-specific, exposure routes that integrate contaminant contributions from more than one medium (e.g., consumption of vegetables) will be segregated so that each medium contributing to the exposure route is evaluated separately. This segregation will be done by assuming that the concentration or activity concentration of contaminants in the medium not under evaluation is zero.

Two methods may be used to develop RGOs. The first involves rearranging and combining all the exposure equations utilized to determine risk or hazard and using the rearranged equation to calculate the RGO. The second simply uses ratios of concentrations or activities and level of risk, hazard, or radiological dose to derive the RGO. Although the first method is of greater utility because the rearranged equation can be used to directly solve for RGOs, its use involves rearranging a large complex equation in which the chance for error abounds, especially if the estimated contaminant concentrations at the exposure point rely on fate and transport modeling. Similarly, although the second method is simpler mathematically, it can result in an incorrect solution if risk, hazard, or radiological dose determined for COCs at the source in the baseline

² A similar heading will appear for each of the major portions of the baseline human health risk assessment. The other headings are "Uncertainties related to exposure assessment," "Uncertainties related to toxicity assessment," and "Uncertainties related to risk characterization."

human health risk assessment is not linearly and directly related to the concentration or activity concentration of the COCs at the exposure point. Fortunately, the concentration or activity concentration in each of the exposure equations that will be used in baseline human health risk assessments at PGDP (see Appendix D) is linearly and directly related to the resulting risk, hazard, or radiological dose; therefore, the second method will be used in risk assessments at PGDP and is presented in Eqs. 14 and 15. Note: If additional exposure equations beyond those in Appendix D are used in an assessment performed for PGDP, these equations will be checked to ensure that the concentration or activity concentration of COCs is directly and linearly related to risk or hazard.

$$\frac{\text{Conc}_{observed}}{\text{ELCR}_{derived}} = \frac{\text{RGO}}{\text{Target ELCR}}$$
Eq. 14

where: $Conc_{observed}$ = The representative EPC for the COC

ELCR_{derived} = The chemical-specific ELCR of a COC due to exposure to a single medium across all exposure routes

RGO = The remedial goal option

Target ELCR = Either 1×10^{-4} , 1×10^{-5} , or 1×10^{-6}

$$\frac{\text{Conc}_{observed}}{\text{HI}_{derived}} = \frac{\text{RGO}}{\text{Target HI}}$$
 Eq. 15

where: $Conc_{observed}$ = The representative EPC for the COC

HI_{derived} = The chemical-specific HI of a COC from exposure to a single medium across all exposure routes

RGO = The remedial goal option

Target HI = Either 3, 1, or 0.1

As noted, radiological dose-based RGOs will be calculated using similar methods. The targets to be used for all media except groundwater are 1, 12, 25, and 100 mrem/year. For groundwater, the radiological dose targets are 1, 4, 12, 25, and 100 mrem/year.

3.3.8.2 Presentation of remedial goal options

As noted, RGOs must be calculated for each COC within each MOC for each use scenario of concern identified in the baseline human health risk assessment; therefore, many RGOs will be developed in most risk assessments considering multiple units or areas. To simplify the consideration of the RGOs by users of the risk assessment, the format in Exhibit 3.17 will be used to present the RGOs in all baseline human health risk assessments prepared for PGDP. Note: Using this format will result in the preparation of a single table containing all COCs within each MOC for each use scenario of concern; therefore, this table or relevant potions of it can be used directly in the FS.

Exhibit 3.17. Presentation of Remedial Goal Options¹

СОС	Rep. Conc. ²	Regulatory Value ³	ELCR at Conc.4	HI at Conc. ⁵	RGO at HI=0.1		RGO at HI=3	ELCR=		ELCR=	
Scenario and m	Scenario and medium ⁶										
# 1 ⁷											
# 2											
	•		•	•	•	•	•	•	•	•	
# N											

A separate table will be made for each unit or area under investigation.

A separate table following a similar format will be prepared for radiological dose-based RGOs.

3.3.8.3 Revising exposure parameters and calculations in the uncertainty section

As part of the uncertainty analysis for the risk assessment, risk may be recalculated with default exposure factors replaced using exposure parameters consistent with site-specific values. The decision to recalculate risks using these alternative exposure parameters would be a product of the consensus of the FFA parties arrived at during project discussions at the appropriate stage in document development. For example, the ED of 25 years for the outdoor worker may be replaced with a shorter duration of 1 to 5 years that is more likely to reflect the potential exposures at the site. The shorter ED and possibly a revised exposure frequency combined with the other default parameters for the outdoor worker scenario also may be used to produce an excavation worker scenario. Also, risk from dermal exposure to soil/sediment could be evaluated quantitatively to determine the impact of the use of default dermal absorption (ABS) values on the risk characterization. These revised calculations may be considered in the development of revised PRGs and cleanup levels to be used in the preparation of remedy selection documents.

² This value will be the representative concentration used in the calculation of risk or hazard in the baseline human health risk assessment.

³ Regulatory values (taken from ARARs) may not be available for some media.

⁴ This value will be the chemical-specific, medium-specific ELCR presented in the baseline human health risk assessment for the scenario of concern.

⁵This value will be the chemical-specific, medium-specific HI presented in the baseline human health risk assessment for the scenario of concern.

⁶ Each MOC within a scenario of concern will be presented. The current use scenario RGOs will be presented first followed by the options for the most likely future use. The options for the least likely future use will appear last. Also, for the ground and surface water RGO tables, the appropriate MCLs will be listed.

⁷ All COCs should be listed, including those that could not be evaluated quantitatively.

4. RISK ANALYSES IN THE PREPARATION OF REMEDY SELECTION DOCUMENTS

As noted in RAGS, Part C, (EPA 1991c) and in A Guide to Preparing Superfund Proposed Plans, Records of Decision, and Other Remedy Selection Documents (EPA 1999b), risk analyses are an integral part of the remedy selection documents (e.g., FS, Proposed Plan, and ROD). The role of risk evaluations in these documents is discussed in this section. Risk evaluations that appear in other documents, including site investigation (SI) documents and engineering evaluations/cost analyses (EE/CAs), should be equivalent in data quality and content to risk assessments in the documents described in this section. Risk assessments in SI and EE/CA documents may vary from those described in the following section depending on how that risk assessment is used in decision-making for the specific project. A more streamlined approach for risk assessments is sometimes used for removal action decision documents.

Risk evaluations begin in the development and screening stage of the FS, extend through the detailed analysis of alternatives in the FS, and are reported in varying level of detail in the Proposed Plan and ROD. The primary goal of risk analyses here is to provide risk managers with the information needed to choose among specific remedial alternatives and to verify that a cleanup level was achieved. Generally, if a piece of risk information is not needed to choose among alternatives or to verify cleanup, it does not need to be generated; however, it should be noted that it is not uncommon for additional risk analyses to occur after the completion and signing of a ROD (e.g., during the design and implementation of the chosen remedy and after the implementation is complete). Generally, additional analyses occur because additional information relevant to the chosen remedy is acquired. Because the need for and form of these analyses is determined on a project-specific basis, the analyses that may occur after the completion of the FS are not discussed in detail here. The information provided in Sections 2 and 3 should be used to guide any additional work to ensure technical adequacy.

4.1 RISK ANALYSES DURING THE FEASIBILITY STUDY

Risk analyses impact four significant portions of the FS. These are the reporting of baseline or screening risk assessment results (including any radiological dose assessment); the evaluation of the risk analyses to determine the need for remedial action; the identification, development, and screening of technologies and alternatives; and support of the detailed analysis of alternatives. These areas are discussed in Sections 4.1.1, 4.1.2, 4.1.3, and 4.1.4, respectively.

4.1.1 Presentation of Risk Assessment Results in the Feasibility Study

Section 7, Summary and Conclusions, of the baseline human health risk assessment often may be copied directly to the FS report to summarize the identified risks that the FS will need to address. Additionally, following guidance in EPA 1999b, the tables consistent with RAGS, Part D, or relevant parts of them can be inserted directly into the FS. The material placed in the FS will contain a summary of the methods used to identify the COPCs and to complete the exposure assessment, toxicity assessment, and risk characterization, including the identification of significant uncertainties affecting the risk results. In addition, the risk characterization summary tables (Exhibit 3.15) and the relevant portions of the RGO summary tables (Exhibit 3.17) can be transported directly to the FS report. Electronic copies of this material will be made available to the authors of the FS report to simplify the reporting of this information and ensure consistency between the RI and FS reports.

As noted in RAGS Part C (EPA 1991c), the primary use of the baseline risk assessment from the RI is to identify the need for remedial action in the absence of any action. A risk evaluation of remedial alternatives will follow the same general steps as a baseline risk assessment. For some FS reports, recalculation of risk or radiological dose estimates may be required to differentiate between remedial alternatives; these additional risk evaluation activities should be conducted within the scope of Chapter 2 of RAGS Part C (EPA 1991c).

The overall objective of the detailed analysis of alternatives is to obtain and present information that is needed for decision makers to select a remedial alternative for a site. The risk evaluations conducted to support the FS are, in effect, residual risk evaluations that determine whether a technology is capable of achieving PRGs. To support alternative analysis, these residual risk evaluations may consider non-default exposure scenarios and impacts of engineering and institutional controls and may use performance-based criteria (i.e., remove all affected media with concentrations greater than a target level).

Most of the time, it will be sufficient for an FS detailed analysis to indicate whether an alternative has the potential to achieve the PRGs, rather than to quantify the risk that will remain after implementation of the alternative. If more detailed information concerning long-term risk is needed to select an alternative (e.g., to determine the more favorable of two otherwise similar alternatives), then it may be useful to determine whether one alternative is more certain to achieve the PRGs than the other, whether (or to what extent) one may be able to surpass (i.e., achieve lower concentrations than) the PRGs, or whether one may be able to achieve the goals in a shorter time.

Thus, an FS risk evaluation that identifies the post-remedy residual risk may need to be coupled with an implementability, certainty, or permanence evaluation to identify the factors that may be needed to be described further in the remedial design to ensure the achievement of remedial goals. As noted in RAGS Part C, Chapter 2, the presence of the five-year review process focuses the degree of these evaluations. For example, if a remedy includes capping of contaminated soils, then the potential future exposures due to cap failure may include direct contact with soils and leaching of contaminants to groundwater. However, the worst-case situation of complete containment system failure does not necessarily need to be evaluated because it is unlikely to occur, because five-year reviews are conducted at all sites where wastes are managed on-site above concentration levels that allow for unrestricted use and unlimited exposure.

The level of risk evaluation to be conducted in the FS should be determined and agreed to by the three FFA parties during scoping for the FS. Situations where risk estimates may need to be updated for the FS report include the following:

- The time between the completion of the RI report and the preparation of the FS report is such that additional information not considered in the RI report becomes available (e.g., additional samples and/or updated toxicity information/values).
- It is determined that the remedial technologies will produce new contaminants that were not present at the site under baseline conditions.
- The decision to include in the FS more advanced modeling from the matrix in Table 3.2 (including probabilistic risk assessment) in the FS than was used in the RI in order to provide refined estimates of risk necessary for determining the long-term or short-term effectiveness of remedial options or the differences in residual risk between remedial options.

Revised PRGs consistent with the alternatives will be derived in the FS. These revised PRGs will utilize the site-specific information in the RI report and the risk assessment in their calculation.

If additional risk evaluations are required in the FS, then these computations will follow the methods outlined in Section 3. Most importantly, the exposure equations presented in Appendix D will be used for all risk computations that appear in the FS report, and the methods presented in Section 3.3.8 for RGO development will be followed.

In FS reports, the summary of the risk evaluation results will be followed by an evaluation of these results. This evaluation will consider the risk estimates, their basis, and the uncertainties deemed relevant to selection of a remedy. This evaluation will provide the focus for RAO development later in the FS report. The information that follows identifies typical decisions made when determining the need for remedial action in the FS report.

4.1.2 Modifications to Risk Assessment Parameters that May Appear in the Feasibility Study

The baseline human health risk assessment identifies whether remedial action is necessary and provides a basis for evaluating the proposed remedial alternatives; therefore, the baseline human health risk assessment typically will not change in the FS.

The uncertainty section of the baseline human health risk assessment will identify whether an uncertainty is small, moderate, or large (i.e., uncertainties categorized as small do not affect the risk estimates by more than one order of magnitude; those categorized as moderate may affect the risk estimates by between one and two orders of magnitude; uncertainties categorized as large may affect the risk estimate by more than two orders of magnitude; see Section 3.3.7). The FS should evaluate the uncertainty in more detail and may recalculate risk values as determined by agreement of the three parties to support the alternative evaluation better.

Calculation of short-term risks during the detailed analysis of remedial alternatives (see Section 4.1.3) may require significant recalculation of risks and hazards from the baseline risk assessment to account for differences between the exposures to current workers and off-site residents and the default values used for the baseline risk assessment in the RI. For example, current industrial workers and current off-site residents do not consume groundwater from the facility for drinking. In addition, current industrial workers have lower dermal exposure and shorter duration of exposure than is assumed for future industrial workers under a default exposure scenario. Outdoor workers also will have lower exposures than the default parameters due to the use of personal protective equipment and engineering controls. These differences need to be accounted for in the evaluation of risk in the FS, and these evaluations shall be incorporated in the discussion of the detailed analysis of remedial alternatives.

4.1.2.1 Land use considerations for determining appropriate response actions to protect future potential receptors

Land use is an important consideration when determining appropriate response actions based on potential future receptors. Uncertainties associated with future land use are due to the inability to predict if existing controls will be in place in the future and the reliability of implementing additional controls. There may be scenarios developed pursuant to this document that may not be commensurate with the reasonable foreseeable land use, but may serve as a reference point to decision makers. Consequently, the results of the baseline human health risk assessment will not be modified when determining potential risks to future receptors. However, additional risk evaluation (beyond the baseline risk assessment) of scenarios may be used to support development of alternatives developed in the FS report. The ability of these alternatives to ensure protection of potential future receptors will be evaluated in the FS, using risk evaluation as appropriate. Protection may be accomplished through continuation of existing controls in some instances or through the application of new controls. Consequently, potential future scenarios will be evaluated in the FS alternative evaluation to supply decision makers with the information needed to choose appropriate

remedial actions. The information that follows provides examples of scenarios that may be evaluated for future receptors in the FS report.

Site-specific exposures for current industrial workers and the inability to predict potential future exposures have been discussed earlier. For a future industrial worker, the risks to a default industrial worker as determined in the baseline human health risk assessment will be used when estimating risks to determine the need for action. This evaluation includes potential risks as a result of contact with contaminated RGA groundwater, which also is a possibility in the future. Additional evaluations may be developed, however, for the future industrial worker to include an evaluation of the impacts of continuation of the existing institutional controls (i.e., controls and procedures that limit access to affected soil and groundwater and provide an alternative water source); continuation of or application of new controls and procedures (i.e., continuation of current industrial scenario); assuming contact with contaminated RGA groundwater (i.e., no separate water source); and default exposure (i.e., no controls or procedures) without contact with contaminated RGA groundwater (i.e., assuming a separate water supply). Any set of exposure parameters agreed to during scoping may be used to develop a scenario. That scenario may be subjected to additional risk evaluation in the FS to identify the remedial action drivers, irrespective of whether or not that scenario is itself realistic.

Future recreational users and residential users inside the DOE property boundary (including area within the restricted access area, but not the surrounding West Kentucky Wildlife Management Area) may be further assessed in the FS report. The risk managers first will review the results of the baseline risk assessment that assumes that no controls would be in place to restrict a future on-site recreational user or resident from contact with surface contamination. As with the industrial worker, however, the risk managers can identify scenarios to be subjected to additional risk analysis that do place restrictions on exposure to be used in considering alternative FS scenarios to best identify the remedial action drivers.

Modeling during the baseline human health risk assessment typically involves a large degree of uncertainty. For this reason, modeling parameters may be reevaluated during the preparation of the FS report, as discussed in the modeling matrix presented in Table 3.2, if needed to reduce uncertainty and aid in choosing between the proposed remedial alternatives. For the same reason, the FS may consider use of probabilistic models for risk assessment in place of the deterministic models used during the RI if these additional analyses are deemed necessary through scoping agreements by the three parties.

4.1.2.2 Identification of use scenarios, pathways, contaminants, and MOC for decision making purposes

Following evaluation of the results and uncertainties in the baseline human health risk assessment, additional risk evaluation performed to support the FS, and finalization of risk management decisions, a list of use scenarios, pathways, contaminants, and MOC for decision making purposes will be developed.

In the FS report, each item of concern will be identified based on the guidance presented in Section 3.3.6.4.

4.1.3 Risk Analyses during the Identification and Screening of Technologies and Alternatives

During the identification and screening stage of the FS, a range of remedial alternatives is identified, and each alternative is evaluated with respect to effectiveness, implementability, and cost (EPA 1991c). As part of the evaluation of effectiveness, human health risks to the community (e.g., short- and long-term health risks from releases during remediation and after remediation, respectively) and remediation workers (i.e., short-term health risks during remedial activities) will be considered. At PGDP, this evaluation will be performed qualitatively to be consistent with guidance in RAGS, Part C.

4.1.4 Risk Analyses during the Detailed Analysis of Alternatives

The overall objective of the detailed analysis of alternatives is to obtain and present the information needed by risk managers to select a remedial alternative for a site (EPA 1991c). Risk analysis affects three of the selection criteria against which alternatives are evaluated: long-term effectiveness, short-term effectiveness, and overall protection of human health and the environment.

Generally, the human health risk analyses performed during the FS will follow the same procedures as the baseline human health risk assessment. Unlike the baseline human health risk assessment, where the purpose is to estimate the risk posed by environmental contamination in the absence of any action, the purpose of the FS risk analyses is to determine by how much the various remedial alternatives reduce risk or to evaluate short-term risks brought about through remedy implementation (e.g., air stripper emissions).

Consistent with RAGS, Part C, (EPA 1991c), at PGDP the risk analyses performed during the detailed analysis of alternatives may be either qualitative or quantitative. In most cases, a qualitative analysis will be sufficient as indicated in RAGS, Part C; however, a quantitative analysis may be required in some cases. The decision about whether a qualitative or quantitative analysis of alternatives is needed will be made using guidance in RAGS, Part C. In this guidance, EPA notes that the type of analysis that is required depends on (1) whether the relative short-term or long-term effectiveness is an important consideration in selecting the alternative and (2) the "perceived risk" associated with the alternative. Where perceived risk is high, a quantitative risk evaluation would be more appropriate. In RAGS, Part C, EPA defines "perceived risk" as that leading to the belief by site engineers, risk assessors, and neighboring communities, including workers, that an alternative either may not be adequately protective or lead to increased risk. Specific parameters that will be taken into account at PGDP when examining "perceived risk" and determining if a quantitative analysis is required include the following (adapted from RAGS, Part C):

- Proximity of populations to the unit or area;
- Presence of highly or acutely toxic chemicals;
- Technologies with high release potential, either planned or unplanned;
- High uncertainties in the nature of releases;
- Multiple contaminants or exposure routes or both affecting the same receptor;
- Releases from neighboring units or areas, including uncontrolled releases from units or areas not yet addressed;
- Releases that occur over a long period; and
- Level of community concern.

4.1.4.1 Qualitative risk evaluations

As noted herein, a qualitative analysis will be sufficient for most units or areas. In this type of analysis, the risk evaluation will qualitatively evaluate each alternative against the RAOs defined during the FS. In all cases, the qualitative analysis will evaluate whether the alternative can reduce exposure to probable and potential receptor populations to acceptable levels. In many evaluations, this will involve qualitatively

determining if an alternative is effective in reducing contaminant concentrations at a unit or area to the cleanup level (i.e., the RGO or revised PRG consistent with the alternative being evaluated).²⁰

In other cases, this will involve determining if an alternative is effective in changing activity patterns of receptors so that the rate of contact by receptors to the contaminated materials is reduced, resulting in a lowered exposure. Finally, the qualitative risk evaluation in the detailed analysis of alternatives for PGDP will examine the potential for an alternative to produce new contaminants that were not at a unit or area during the RI.

In developing the risk evaluation portion of the qualitative detailed analysis of alternatives, several sources of information will be used. These sources are listed below [adapted from RAGS, Part C, (EPA 1991c)] and include information from the baseline or screening risk assessment (as modified during the risk management to determine the need for action), treatability studies, and results at other sites. Material from the risk assessment includes the following:

- The exposure setting, including exposed populations and future land use;
- The exposure pathways, including sources of contamination, COCs, fate and transport of chemicals (i.e., migration, degradation, and transformation), and exposure points;
- General exposure considerations, including rate of contact, exposure frequency, and ED;
- Exposure concentrations, including temporal effects;
- Estimates of chemical intake and uptake;
- Toxicity information, including uncertainty in toxicity values; and
- Methods used to quantify risks from exposure to media containing multiple chemicals and radionuclides.

Material found in treatability studies that will be used in the qualitative risk evaluation includes the following:

- Effectiveness at reducing potential for exposure, either through reduction in contaminant concentrations and activities or through making the medium containing the contaminant unavailable for contact;
- Potential for short-term emissions; and
- Potential for production of new contaminants.

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²⁰ "Preliminary remediation goals...may be revised...based on the consideration of appropriate factors including, but not limited to: exposure factors, uncertainty factors, and technical factors. Included under exposure factors are: cumulative effect of multiple contaminants, the potential for human exposure from other pathways at the site, population sensitivities, potential impacts on environmental receptors, and cross-media impacts of alternatives. Factors related to uncertainty may include: the reliability of alternatives, the weight of scientific evidence concerning exposures and individual and cumulative health effects, and the reliability of exposure data. Technical factors may include: detection/quantification limits for contaminants, technical limitations to remediation, the ability to monitor and control movement of contaminants, and background levels of contaminants. The final selection of the appropriate risk level is made when the remedy is selected based on the balancing of criteria...." [taken from the National Contingency Plan Preamble: Exposure, Technical, and Uncertainty Factors (55 Fed. Reg. 8717, March 8, 1990)]. Also, see RAGS Volume 1, Part B, Section 2.3 and 2.8 (EPA 1993a) and OSWER Directive 9355.0-30, "Role of the Baseline Risk Assessment in Superfund Remedy Selection Decisions" (EPA 1990a).

Materials found when examining results from other sites that will be used in the qualitative risk evaluation include the following:

- Actual contaminant reductions achieved;
- Conditions in which the technology was not effective; and
- Actual release rates of current or new contaminants.

4.1.4.2 Quantitative risk evaluations

Methods for quantitative risk evaluations during the detailed analysis of alternatives will follow the same procedures as the baseline human health risk assessment. Unlike the baseline human health risk assessment, where the goal is to estimate the potential risk posed by environmental contamination in the absence of any action, the goal of the FS risk analyses is to determine to what extent the various remedial alternatives reduce risk such that unacceptable levels of risk are not posed by residual environmental contamination.

4.2 RISK ANALYSES AFTER THE FEASIBILITY STUDY

After the FS is completed, a remedy is proposed in the Proposed Plan and documented in the ROD. Following this, the remedy is designed and implemented and, depending on the remedy, the site either is deleted or is placed within the group for which five-year reviews are required. This section discusses the risk evaluation activities that will occur during and after preparation of the Proposed Plan. These risk evaluation activities should be consistent with EPA guidance in the *Guide to Preparing Superfund Proposed Plans, Records of Decision, and other Remedy Selection Decision Documents* (EPA 1999b). Some of the material presented here was taken from RAGS, Part C (EPA 1991c).

4.2.1 Risk Evaluation for the Proposed Remedial Action Plan

Typically, no new risk evaluations will take place during the preparation of the Proposed Plan. The material presented in the Proposed Plan should be taken from the supporting FS. This includes a summary of site risks, the site COCs, and, if applicable, the revised PRGs for the selected alternative or a description of the basis for them (i.e., risk or radiological dose target). Consistent with EPA 1999b, the material presented in the "Summary of Site Risks" section of the Proposed Plan primarily will be presented as narrative and limited to approximately three paragraphs. Key information from the baseline risk assessment (or other FS risk evaluations) that will be presented includes all the following:

- Major COCs in each medium
- Land- and groundwater-use assumptions
- Potentially exposed populations under current and future use scenarios
- Major pathways and routes of exposure
- Summary of risk characterization

The risk section of the Proposed Plan also will contain a text box of standard language from the Proposed Plan/ROD guidance (EPA 1999b). This standard language will contain a definition of risk assessment and the meaning of the results from a risk assessment. The risk section of the Proposed Plan will conclude with language similar to the following text taken from EPA 1999b.

It is the lead agency's current judgment that the Preferred Alternative identified in this Proposed Plan, or one of the other active measures considered in the Proposed Plan, is necessary to protect public health or welfare or the environment from actual or threatened

releases of pollutants or contaminants from this site. These pollutants or contaminants may present an imminent and substantial endangerment to public health or welfare.

If new information becomes available during the public comment period, then additional analysis of the alternatives, or possibly the site risks, may be needed. (Note: These analyses will encompass all alternatives and be performed qualitatively to the extent possible.)

4.2.2 Risk Evaluation for the ROD

The primary risk evaluation-related activities that will occur during the ROD will be to document the results of the risk assessment and the risk evaluation portions of the comparison of alternatives performed in the FS and to document the derivation of the chemical-specific cleanup levels. Consistent with EPA guidance in both *Guide to Preparing Superfund Proposed Plans, Records of Decision, and other Remedy Selection Decision Documents* (EPA 1999b) and RAGS, Part C (EPA 1991c), the appropriate risk evaluation materials will be discussed in relation to three of the nine CERCLA alternative analysis criteria: long-term effectiveness, short-term effectiveness, and overall protection of human health and the environment. The discussion of overall protection of human health and the environment will consider, to the extent possible, any residual risks that may remain after the alternative is implemented. Specific information to be presented includes the following:

- Chemical-specific cleanup levels to be attained at the conclusion of the response action;
- Corresponding chemical-specific risk levels;
- Areas of attainment for cleanup levels for groundwater being addressed; and
- Lead agency's basis for the cleanup levels (e.g., risk calculation, ARARs, background, etc.).

To the extent possible, the "Summary of Site Risks" section of the ROD will be presented following the outline contained in EPA 1999b; therefore, this material will include the following:

- A statement of basis for taking action and
- A brief summary of the relevant portions of the risk assessment.

Additionally, this section will focus on the risk drivers as defined in the FS and the exposure scenarios and pathways driving the need for action. The conceptual site model (which should be presented in the *Summary of Site Characteristics* section of the ROD) will be used to support the presentation of site risks.

The standard language to be used for the statement of basis for action will be similar to that which also appears in the Proposed Plan. For the ROD, this statement will appear at the beginning of the site risk summary instead of at the end.

In most cases, the tabular format in EPA 1999b will be used to present risk assessment/evaluation results in the ROD; however, additional tables or tables of a slightly different format may be used to explain the risk assessment/evaluation results, as needed. Note that the primary purpose for including the detailed risk characterization tables in an appendix of the baseline risk assessment is to streamline the preparation of these tables for the FS and ROD.

4.2.3 Risk Analyses for Residual Risks

As noted in RAGS, Part C, (EPA 1991c) analyses to examine residual risks may be required for some locations after implementation of a remedy. Additionally, as discussed in the SMP (DOE 2022a), after completion of all investigations and remedial actions at PGDP, the FFA requires that PGDP determine the residual risks remaining at the facility. Finally, the five-year review of some sites may require additional

residual risk analyses. These residual risk analyses should be conducted consistent with guidance on the five-year review process from both EPA (EPA 2001c; EPA 2003c) and DOE (DOE 2002).

The methods to be used to complete the analyses of residual risks at most units will be qualitative. If quantitative, these analyses will be consistent with the methods in either Section 2 or that in Section 3 of this document. Additionally, any quantitative analyses will be consistent with Section 3.3.4 of RAGS, Part C (EPA 1991c). Generally, these analyses will determine the risks remaining after remediation due to contamination remaining at or migrating from sources (or multiple sources). In these analyses, the measured concentrations and activities of contaminants remaining at the various sources units and in the integrator unit will be used. The cleanup levels in the ROD for the various source units and areas in the integrator units should not be used in these analyses.

Other issues that will be considered when evaluating residual risk will be the following:

- Concentrations and activities of new analytes formed as a result of remedial activities or because of natural processes;
- Changes in land use or proposed future use since the completion of the baseline risk assessment;
- Updated toxicity values; and
- Reduction of migration because of engineered controls and expected future performance of these controls.



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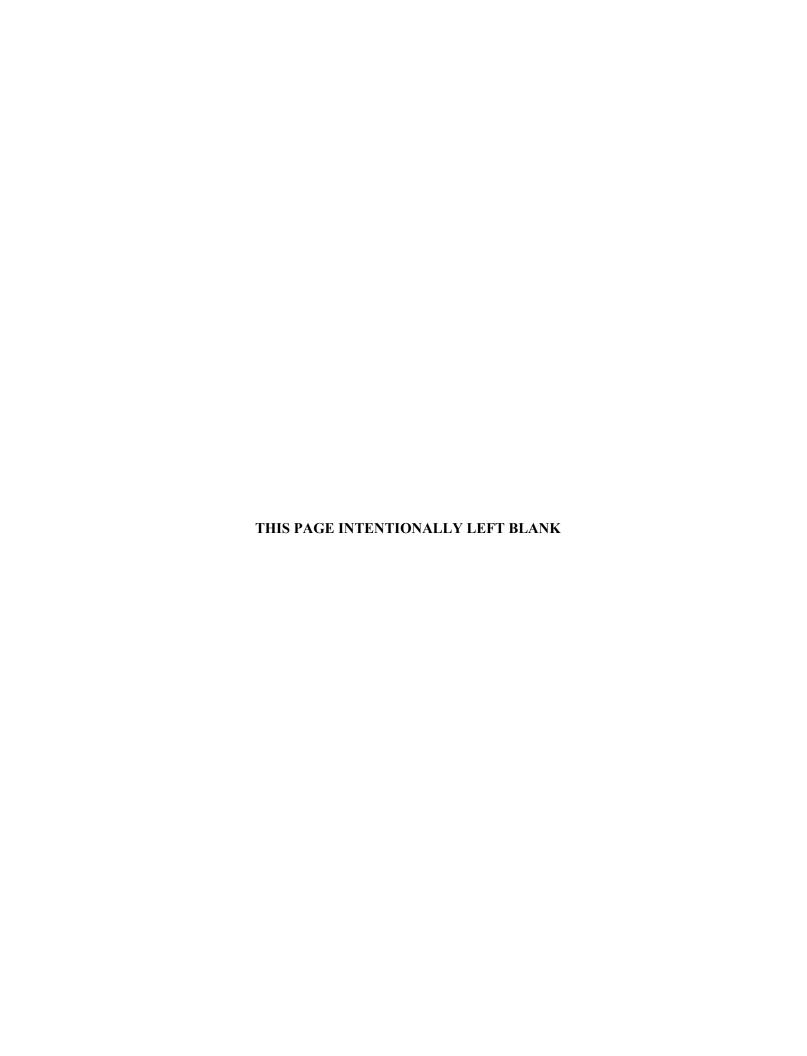
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APPENDIX A

SCREENING LEVELS (CURRENT AS OF OCTOBER 2022)



SCREENING LEVELS

This appendix presents lists of values that can be used during screening and baseline human health risk assessments at the Paducah Gaseous Diffusion Plant (PGDP). These values include risk- and dose-based values for soil, sediment, groundwater, and surface water; background values for soil and groundwater; and regulatory values. Risk-based action levels (ALs) and no action levels (NALs), and radiological dose-based concentrations for significant chemicals or radionuclides of potential concern (COPCs) are presented in Tables A.1 through A.11. ALs and NALs for contaminants in soil based on radiological dose limits are derived by following U.S. Environmental Protection Agency (EPA) guidance as described throughout this document. ALs are defined as concentrations of contaminants above which early actions may be warranted after consideration of project- and location-specific conditions. NALs are defined as concentrations of contaminants below which no action is generally warranted.

All information is current as of the production dates listed in this document, and all values were calculated using the best available information. Methods used to derive the risk- and dose-based values are presented in Appendix B. The screening values presented in this appendix were developed specifically for PGDP and may not be applicable to sites outside that facility. Values are provided in these tables for significant COPCs for PGDP. Values for other COPCs can be obtained using the Risk Assessment Information System (RAIS) Chemical Preliminary Remediation Goal (PRG) online calculator and the EPA Radionuclide PRG calculator, as modified using PGDP-specific inputs (RAIS 2022, EPA 2022).

Please consider the following notes before using the values presented in this appendix.

- (1) ALs are the lesser of a hazard-based value calculated using a target hazard index (HI) of 3 and a cancer-based value calculated using a target excess lifetime cancer risk (ELCR) of 1E-04. Also, see Section 2, Risk Analysis During Scoping Activities, for additional discussion.
- (2) HI values are calculated separately for each receptor. Cancer risks for receptors within a scenario are combined to give one ELCR. For the residential scenario, the cancer risk reflects the adult and child combined. For the recreational scenario, the cancer risk reflects the combined risk to adult, child, and teen.
- (3) ALs are calculated using only direct exposure pathways. Please see Appendix B for a listing of exposure parameters included in the AL calculations. Because not all of the ALs are calculated using PGDP default exposure parameters (e.g., see note 9a), these ALs should be used as benchmarks only. Cumulative risk calculations should not be based upon these ALs.
- (4) NALs are the lesser of a hazard-based value calculated using a target HI of 0.1 and a cancer-based value calculated using a target ELCR of 1E-06. If more than five COPCs are identified for the site, it also may be appropriate to generate NALs based on 1E-07 risk to account for additivity of risk. These NALs were calculated using the exposure parameters listed with the exposure equations in Appendix D. These parameters also are listed in Appendix B. Because the exposure parameters are consistent with the PGDP default exposure parameters, these NALs can be used to derive cumulative risk estimates in addition to their use as benchmarks. Also, see Section 2, Risk Analysis During Scoping Activities, for additional discussion.
- (5) Background values for soil and groundwater presented in this appendix are provisional. Soil background values, except as noted, were derived as detailed in DOE 1996 and DOE 1997a and have been approved by EPA and the Commonwealth of Kentucky as representative background concentrations. Groundwater background values were derived from a study presented in the

Groundwater Operable Unit Feasibility Study Report (DOE 2000) but have not been approved by EPA and the Commonwealth of Kentucky as representative background concentrations. These background values for soil and groundwater have not been agreed to for all uses by the PGDP Risk Assessment Working Group; therefore, these background values are subject to change should other values be more appropriate.

- (6) Soil screening levels for chemicals for protection of groundwater were derived using RAIS and PGDP-specific inputs. The Soil Screening Level (SSL) values based upon a dilution attenuation factor (DAF) of 1 should be considered to be "no action values." "Action" SSLs have not been selected to date for the PGDP. In addition to the SSLs at a DAF of 1, SSLs at a DAF of 20 also are included.
- (7) Regulatory values are for planning purposes only and may not be applicable to conditions at PGDP. Maximum contaminant level (MCL) values are included with the Risk Methods Document's groundwater ALs and NALs. A qualified regulatory specialist should be consulted before using these values for other purposes.
- (8) The outdoor worker scenario is defined as the person exposed to surface soil (i.e., 0–1 ft) inside the industrialized area and surface and subsurface soil (i.e., 0–10 ft or 0–16 ft, as appropriate) outside the industrialized area, for an exposure duration of 25 years for 185 days/year. The excavation worker scenario is defined as the person exposed to surface and subsurface soil (i.e., 0–10 ft or 0–16 ft, as appropriate) for an exposure duration of 5 years for 185 days/year.
- (9) COPC-specific notes for risk-based and dose-based screening values:
 - a) General—Several soil/sediment screening values (especially those on the AL tables) are listed with a value of 100,000. This value was assigned to the COPC because the screening value derived for the contaminant exceeded the upper limit value deemed reasonable by the PGDP Risk Assessment Working Group; therefore, the screening value was reduced to an upper limit value (100,000 mg/kg or pCi/g). If the COPC's environmental concentration exceeds the upper limit value, then additional risk evaluations for the COPC should be performed before accepting the results of a simple comparison. Saturation limits (Csat) are limit values in soil or sediment. The Csat "corresponds to the contaminant concentration in soil [or sediment] at which the absorptive limits of the soil particles, the solubility limits of the pore water, and saturation of soil pore air have been reached," (Regional Screening Level User's Guide, Section 5.12). Because of the variability in soils for the characteristics of absorption capacity on the particles, in the pore water and pore air, the use of Csat saturation limits to compare the ALs is not recommended.

Surface water and groundwater screening values (especially those on the AL tables) may exceed the solubility limit for the analyte; a comparison has not been performed.

- b) Chromium—The screening value for Chromium VI presented in these tables should be used for both Chromium VI and "Total Chromium" results unless it is determined on a project-specific basis that Chromium VI is not present.
- c) Lead—The screening values for lead were selected by the PGDP Risk Assessment Working Group. These values were not derived using the methods presented in Appendix B. NALs are 400 mg/kg for soil/sediment for the resident and the recreator scenarios and 800 mg/kg for the industrial worker and outdoor worker scenarios. These values represent the current screening values provided by the Kentucky Department for Environmental Protection. ALs for soil/sediment are set preliminarily equivalent to the NALs. Sites at which the 400 mg/kg concentration in soil is exceeded should be evaluated using site specific Integrated Exposure Uptake Biokinetic (IEUBK)

modeling for a level resulting in a child exceeding a target blood level of 2.5 μ g/dL [the Commonwealth of Kentucky's recommended blood lead level (Section B.3)] and a target blood level of 5 μ g/dl and Adult Lead Model (ALM) modeling for an adult exceeding the same target blood lead levels. Parameters for the ALM model should be developed for each site. Lead NALs and ALs for groundwater and for surface water are unchanged from those agreed to by the PGDP Risk Assessment Working Group in the 2001 version of this document.

- d) Carcinogenic polycyclic aromatic hydrocarbons (cPAHs)—[These organic compounds include benz(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene, chrysene, dibenz(a,h)anthracene, and indeno(1,2,3-cd)pyrene.] The PGDP Risk Assessment Working Group has determined that these compounds should be evaluated as a group using the PAH (Total) screening values. Please see Section 3.3.3.2, Step 8, of the main text of the methods document for guidance on deriving total PAH concentration from results for individual compounds.
- e) Polychlorinated biphenyls (PCB)—These organic compounds include those listed as Aroclors in the screening tables. The PGDP Risk Assessment Working Group has determined that the cancer effects of these organic compound mixtures should be evaluated as a group using the PCB (Total) screening values. (The screening value associated with the highest risk value is to be used.) Please see Section 3.3.3.2, Step 8, of the main text of the methods document for guidance on deriving total PCB concentration from results for individual mixtures.
- f) Dioxins/Furans—(These organic compounds include the following chlorinated dioxins and furans: 2,3,7,8-TCDD; 1,2,3,7,8-PeCDD; 2,3,4,7,8-PeCDD; 2,3,5,7,8-PeCDD; 2,3,6,7,8-PeCDD; 1,2,3,4,7,8-HxCDD; 1,2,3,5,7,8-HxCDD, 1,2,3,6,7,8-HxCDD; 2,3,4,5,7,8-HxCDD; 2,3,4,6,7,8-HxCDD; 1,2,3,4,5,7,8-HpCDD; 1,2,3,4,6,7,8-HpCDD; 2,3,5,6,7,8-HxCDD; 2,3,7,8-TCDF; 1,2,3,7,8-PeCDF; 2,3,4,7,8-PeCDF; 2,3,4,5,6,7,8-HpCDD; OCDD; 1,2,3,4,7,8-HxCDF; 2,3,4,5,7,8-HxCDF; 1,2,3,5,7,8-HxCDF, 1,2,3,6,7,8-HxCDF; 2,3,4,6,7,8-HxCDF; 2,3,5,6,7,8-HxCDF; 1,2,3,4,5,7,8-HpCDF; 1,2,3,4,6,7,8-HpCDF; 2,3,4,5,6,7,8-HpCDF; and OCDF.) The PGDP Risk Assessment Working Group has determined that these organic compounds should be evaluated as a group using the Dioxins/Furans (Total) screening values. Please see Section 3.3.3.2, Step 8, of the main text of the methods document for guidance on deriving the total dioxins/furans concentration from results for individual compounds.
- g) Radionuclides—Three source and decay assumptions are presented for radionuclides: (1) peak risk—infinite time frame, (2) peak risk—1,000-year time frame, and (3) secular equilibrium. The resulting values and the source and decay assumptions for each value are presented in corresponding tables. Information on the time intervals required to reach the peak risk for each radionuclide for the relevant receptor exposure scenarios is provided in Appendix B.

Radionuclides—Dose targets are (1) 1 millirem (mrem)/year (from the National Council on Radiological Protection and Measurements Report No. 116, *Limitation of Exposure to Ionizing Radiation*, Section 17, "Negligible Individual Dose," and ANSI/Health Physics Society standard N13.12); (2) 12 mrem/year (from "Radiation Risk Assessment at CERCLA Sites: Q & A" Office of Solid Waste and Emergency Response No. 9200.4-40, June 13, 2014); (3) 25 mrem/year (derived from the public dose limit of 100 mrem/year limit in DOE Order 458.1 and considering as low as reasonably achievable principles); and (4) 100 mrem/year. A value of 4 mrem/year for beta particles and photon emitters is used for groundwater (from the radionuclides section of https://www.epa.gov/ground-water-and-drinking-water/national-primary-drinking-water-regulations). As with risk-based ALs and NALs for COPCs, dose-based screening levels are used in project screening only and should not be considered clean-up values. Also, the dose-based screening levels differ from the derived concentration standards in DOE's *Derived Concentration*

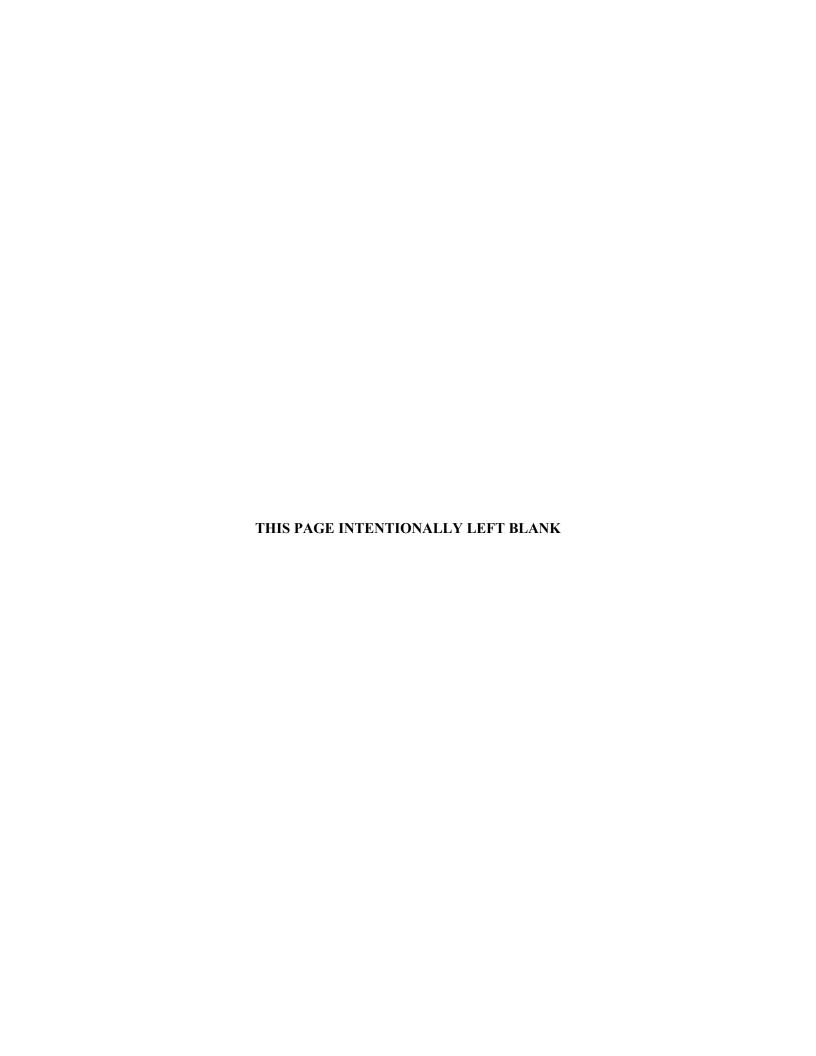
Technical Standard (DOE-STD-1196-2021) because methods used to calculate these two sets of values differ.

- h) Per- and polyfluoroalkyl substances (PFAS)—Screening values for PFAS are not included in this document. Current information related to risk assessment methods for these substances can be found at EPA's PFAS website (see https://www.epa.gov/pfas).
- i) Surrogate chemicals—Selection of surrogate chemicals is based on toxicological information [e.g., the EPA CompTox Chemicals Dashboard (https://comptox.epa.gov/dashboard/)]. Surrogates currently approved by the PGDP Risk Assessment Working Group include the following:
 - acenaphthene for acenaphthylene,
 - fluorene for carbazole,
 - pyrene for benzo(e)pyrene,
 - pyrene for benzo(g,h,i)perylene, and
 - pyrene for phenanthrene for baseline risk assessment and acenaphthene for phenanthrene as part of uncertainty evaluations/discussions.

Due to the nature of Appendix A, not all acronyms are defined within the text. An acronym list is provided on page A-9.

TABLES

A.1a.	Soil/Sediment Action Levels for Significant COPCs at PGDP	A-11
A.1b.	Soil/Sediment Action Levels for Significant Radionuclide COPCs at PGDP	A-20
A.2a.	Groundwater Action Levels and Primary MCLs for Significant COPCs at PGDP	
A.2b.	Groundwater Action Levels and Primary MCLs for Significant Radionuclide COPCs at	
	PGDP	A-25
A.3a.	Surface Water Action Levels for Significant COPCs at PGDP	A-26
A.3b.	Surface Water Action Levels for Significant Radionuclide COPCs at PGDP	A-35
A.4a.	Soil/Sediment No Action Levels for Significant COPCs at PGDP	A-37
A.4b.	Soil/Sediment No Action Levels for Significant Radionuclide COPCs at PGDP	A-46
A.5a.	Groundwater No Action Levels and Primary MCLs for Significant COPCs at PGDP	A-48
A.5b.	Groundwater No Action Levels and Primary MCLs for Significant COPCs at PGDP	A-51
A.6a.	Surface Water No Action Levels for Significant COPCs at PGDP	A-52
A.6b.	Surface Water No Action Levels for Significant Radionuclide COPCs at PGDP	A-61
A.7a.	Risk-Based SSLs for Protection of RGA Groundwater for Significant COPCs at PGDP	A-63
A.7b.	Risk-Based SSLs for Protection of RGA Groundwater for Radionuclide COPCs at PGDP	A-66
A.8.	Dose-Based Soil/Sediment Screening Levels for Site-Related Radionuclides at PGDP	A-67
A.9.	Dose-Based Groundwater Screening Levels for Site-Related Radionuclides at PGDP	A-70
A.10.	Dose-Based Surface Water Screening Levels for Site-Related Radionuclides at PGDP	A-71
A.11.	Dose-Based SSLs for Protection of RGA Groundwater for Site-Related Radionuclides at	
	PGDP	A-72
A.12.	Background Concentrations for Surface and Subsurface Soil at PGDP	A-74
A.13.	Background Concentrations for Groundwater Drawn from the RGA and	
	McNairy Formation at PGDP	A-75



ACRONYMS

AL action level ALM Adult Lead Model

ANSI American National Standards Institute

ATSDR Agency for Toxic Substances and Disease Registry

BaP benzo(a)pyrene

CAS Chemical Abstracts Service

CERCLA Comprehensive Environmental Response, Compensation, and Liability Act

CLP Contract Laboratory Program

COPC chemical or radionuclide of potential concern cPAHs carcinogenic polycyclic aromatic hydrocarbons

Csat saturation limit

DAF dilution attenuation factor DOE U.S. Department of Energy ELCR excess lifetime cancer risk

EPA U.S. Environmental Protection Agency

GW groundwater HI hazard index

IEUBK Integrated Exposure Uptake Biokinetic
IRIS Integrated Risk Information System
K_d chemical-specific distribution coefficient

MCL maximum contaminant level

N/A not available NAL no action level

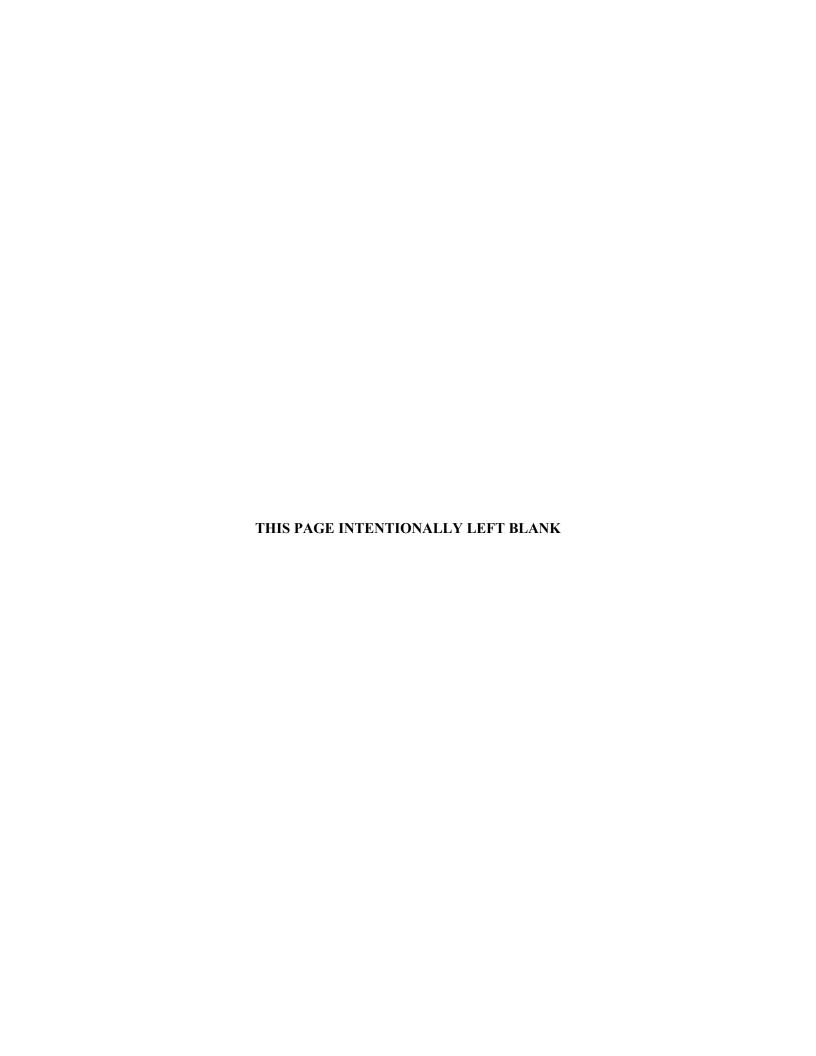
NASA National Aeronautics and Space Administration

PAH polycyclic aromatic hydrocarbon

PCB polychlorinated biphenyl

PFAS per- and polyfluoroalkyl substances
PGDP Paducah Gaseous Diffusion Plant
PRG Preliminary Remediation Goal
RAIS Risk Assessment Information System
RCRA Resource Conservation and Recovery Act

RfC reference concentration
RfD oral reference dose
RGA Regional Gravel Aquifer
RGO remedial goal option
SSL Soil Screening Level
UTL upper tolerance limit



CAS			Ou	tdoor Work	ker	Exc	avation Wo	rker	Industrial Worker		
Number	Analyte	Units	Cancer	Hazard	Action	Cancer	Hazard	Action	Cancer	Hazard	Action
7429-90-5	Aluminum	mg/kg	-	1.00E+05	1.00E+05	-	1.00E+05	1.00E+05	-	1.00E+05	1.00E+05
7440-36-0	Antimony (metallic)	mg/kg	-	3.96E+02	3.96E+02	-	3.96E+02	3.96E+02	-	2.80E+03	2.80E+03
7440-38-2	Arsenic, Inorganic	mg/kg	7.48E+01	3.60E+02	7.48E+01	3.74E+02	3.60E+02	3.60E+02	1.60E+02	7.71E+02	1.60E+02
7440-39-3	Barium	mg/kg	-	1.00E+05	1.00E+05	-	1.00E+05	1.00E+05	-	1.00E+05	1.00E+05
7440-41-7	Beryllium and compounds	mg/kg	1.00E+05	1.97E+03	1.97E+03	1.00E+05	1.97E+03	1.97E+03	1.00E+05	1.35E+04	1.35E+04
7440-42-8	Boron And Borates Only	mg/kg	-	1.00E+05	1.00E+05	-	1.00E+05	1.00E+05	-	1.00E+05	1.00E+05
7440-43-9	Cadmium (Diet)	mg/kg	1.00E+05	7.62E+01	7.62E+01	1.00E+05	7.62E+01	7.62E+01	1.00E+05	1.83E+02	1.83E+02
16065-83-1	Chromium(III), Insoluble Salts	mg/kg	-	1.00E+05	1.00E+05	-	1.00E+05	1.00E+05	-	1.00E+05	1.00E+05
18540-29-9	Chromium(VI)	mg/kg	1.83E+02	2.96E+03	1.83E+02	9.14E+02	2.96E+03	9.14E+02	1.23E+03	2.08E+04	1.23E+03
7440-47-3	Chromium (Total) ^a	-	-	-	-	-	-	-	-	-	-
7440-48-4	Cobalt		1.00E+05	2.95E+02	2.95E+02	1.00E+05	2.95E+02	2.95E+02	1.00E+05	2.06E+03	2.06E+03
7440-50-8	Copper	mg/kg	-	3.96E+04	3.96E+04	-	3.96E+04	3.96E+04	-	1.00E+05	1.00E+05
16984-48-8	Fluoride	mg/kg	-	3.96E+04	3.96E+04	-	3.96E+04	3.96E+04	-	1.00E+05	1.00E+05
7439-89-6	Iron	mg/kg	-	1.00E+05	1.00E+05	-	1.00E+05	1.00E+05	-	1.00E+05	1.00E+05
7439-92-1	Lead ^b	mg/kg	-	-	8.00E+02	-	-	8.00E+02	-	-	8.00E+02
7439-96-5	Manganese (Non-Diet)	mg/kg	-	2.32E+04	2.32E+04	-	2.32E+04	2.32E+04	-	1.00E+05	1.00E+05
Various	Mercury, Inorganic Salts	mg/kg	-	2.96E+02	2.96E+02	-	2.96E+02	2.96E+02	-	2.10E+03	2.10E+03
7439-98-7	Molybdenum	mg/kg	-	4.92E+03	4.92E+03	-	4.92E+03	4.92E+03	-	3.51E+04	3.51E+04
7440-02-0	Nickel Soluble Salts	mg/kg	1.00E+05	1.96E+04	1.96E+04	1.00E+05	1.96E+04	1.96E+04	1.00E+05	1.00E+05	1.00E+05
7782-49-2	Selenium	mg/kg	-	4.92E+03	4.92E+03	-	4.92E+03	4.92E+03	-	3.51E+04	3.51E+04
7440-22-4	Silver	mg/kg	-	4.92E+03	4.92E+03	-	4.92E+03	4.92E+03	-	3.51E+04	3.51E+04
7440-28-0	Thallium (Soluble Salts)	mg/kg	-	9.87E+00	9.87E+00	-	9.87E+00	9.87E+00	-	7.02E+01	7.02E+01
N/A	Uranium (Insoluble Compounds) c	mg/kg	-	2.95E+03	2.95E+03	-	2.95E+03	2.95E+03	-	2.04E+04	2.04E+04
7440-61-1	Uranium (Soluble Salts) c	mg/kg	-	1.97E+02	1.97E+02	-	1.97E+02	1.97E+02	-	1.40E+03	1.40E+03
7440-62-2	Vanadium and Compounds	mg/kg	-	4.95E+03	4.95E+03	-	4.95E+03	4.95E+03	-	3.45E+04	3.45E+04
7440-66-6	Zinc and Compounds	mg/kg	-	1.00E+05	1.00E+05	-	1.00E+05	1.00E+05	-	1.00E+05	1.00E+05
83-32-9	Acenaphthene	mg/kg	-	3.03E+04	3.03E+04	-	3.03E+04	3.03E+04	-	4.14E+04	4.14E+04
208-96-8	Acenaphthylene ^d	mg/kg	-	3.03E+04	3.03E+04	-	3.03E+04	3.03E+04	-	4.14E+04	4.14E+04
107-13-1	Acrylonitrile	mg/kg	8.93E+01	2.66E+02	8.93E+01	4.46E+02	2.66E+02	2.66E+02	1.24E+02	2.01E+02	1.24E+02
120-12-7	Anthracene	mg/kg	-	1.00E+05	1.00E+05	-	1.00E+05	1.00E+05	-	1.00E+05	1.00E+05
71-43-2	Benzene	mg/kg	5.19E+02	1.28E+03	5.19E+02	2.59E+03	1.28E+03	1.28E+03	5.31E+02	1.33E+03	5.31E+02
117-81-7	Bis(2-ethylhexyl)phthalate ^e	mg/kg	3.79E+03	1.14E+04	3.79E+03	1.90E+04	1.14E+04	1.14E+04	5.80E+03	1.74E+04	5.80E+03
75-27-4	Bromodichloromethane	mg/kg	1.59E+02	7.89E+03	1.59E+02	7.93E+02	7.89E+03	7.93E+02	1.30E+02	1.00E+05	1.30E+02
86-74-8	Carbazole	mg/kg	2.65E+03	-	2.65E+03	1.33E+04	-	1.33E+04	4.06E+03	-	4.06E+03
56-23-5	Carbon Tetrachloride	mg/kg	3.14E+02	1.59E+03	3.14E+02	1.57E+03	1.59E+03	1.57E+03	2.96E+02	1.84E+03	2.96E+02
67-66-3	Chloroform	mg/kg	1.78E+02	3.12E+03	1.78E+02	8.90E+02	3.12E+03	8.90E+02	1.39E+02	3.21E+03	1.39E+02

CAS			Ou	tdoor Work	ær	Exc	avation Wo	rker	Ind	ustrial Wor	ker
Number	Analyte	Units	Cancer	Hazard	Action	Cancer	Hazard	Action	Cancer	Hazard	Action
75-71-8	Dichlorodifluoromethane (Freon-12) ^e	mg/kg	-	1.48E+03	1.48E+03	-	1.48E+03	1.48E+03	-	1.10E+03	1.10E+03
75-34-3	Dichloroethane, 1,1-e	mg/kg	1.90E+03	1.69E+04	1.90E+03	9.52E+03	1.69E+04	9.52E+03	1.58E+03	1.36E+04	1.58E+03
107-06-2	Dichloroethane, 1,2-	mg/kg	2.26E+02	5.19E+02	2.26E+02	1.13E+03	5.19E+02	5.19E+02	2.09E+02	4.17E+02	2.09E+02
75-35-4	Dichloroethylene, 1,1-	mg/kg	-	3.78E+03	3.78E+03	-	3.78E+03	3.78E+03	-	3.00E+03	3.00E+03
540-59-0	Dichloroethylene, 1,2- (Mixed Isomers)	mg/kg	-	8.88E+03	8.88E+03	-	8.88E+03	8.88E+03	-	6.30E+04	6.30E+04
156-59-2	Dichloroethylene, cis-1,2-	mg/kg	-	9.34E+02	9.34E+02	-	9.35E+02	9.35E+02	-	1.20E+03	1.20E+03
156-60-5	Dichloroethylene, trans-1,2-	mg/kg	-	1.17E+03	1.17E+03	-	1.17E+03	1.17E+03	-	9.12E+02	9.12E+02
60-57-1	Dieldrin	mg/kg	3.32E+00	2.84E+01	3.32E+00	1.66E+01	2.84E+01	1.66E+01	5.08E+00	4.35E+01	5.08E+00
1746-01-6	Dioxins/Furans, Total (as TCDD) ^f	mg/kg	5.76E-04	5.67E-04	5.67E-04	2.88E-03	5.67E-04	5.67E-04	1.57E-03	1.57E-03	1.57E-03
37871-00-4	~HpCDD	mg/kg	5.77E-02	5.67E-02	5.67E-02	2.89E-01	5.67E-02	5.67E-02	1.58E-01	1.57E-01	1.57E-01
38998-75-3	~HpCDF, 2,3,7,8-	mg/kg	5.79E-02	5.67E-02	5.67E-02	2.90E-01	5.67E-02	5.67E-02	1.60E-01	1.57E-01	1.57E-01
34465-46-8	~HxCDD, 2,3,7,8-	mg/kg	5.80E-03	5.67E-03	5.67E-03	2.90E-02	5.67E-03	5.67E-03	1.61E-02	1.58E-02	1.58E-02
55684-94-1	~HxCDF, 2,3,7,8-	mg/kg	5.80E-03	5.67E-03	5.67E-03	2.90E-02	5.67E-03	5.67E-03	1.61E-02	1.58E-02	1.58E-02
3268-87-9	~OCDD	mg/kg	1.93E+00	1.89E+00	1.89E+00	9.67E+00	1.89E+00	1.89E+00	5.38E+00	5.25E+00	5.25E+00
39001-02-0	~OCDF	mg/kg	1.93E+00	1.89E+00	1.89E+00	9.67E+00	1.89E+00	1.89E+00	5.38E+00	5.25E+00	5.25E+00
36088-22-9	~PeCDD, 2,3,7,8-	mg/kg	5.80E-04	5.67E-04	5.67E-04	2.90E-03	5.67E-04	5.67E-04	1.61E-03	1.58E-03	1.58E-03
57117-41-6	~PeCDF, 1,2,3,7,8-	mg/kg	1.93E-02	1.89E-02	1.89E-02	9.67E-02	1.89E-02	1.89E-02	5.38E-02	5.25E-02	5.25E-02
57117-31-4	~PeCDF, 2,3,4,7,8-	mg/kg	1.93E-03	1.89E-03	1.89E-03	9.67E-03	1.89E-03	1.89E-03	5.38E-03	5.25E-03	5.25E-03
1746-01-6	~TCDD, 2,3,7,8-	mg/kg	5.76E-04	5.67E-04	5.67E-04	2.88E-03	5.67E-04	5.67E-04	1.57E-03	1.57E-03	1.57E-03
51207-31-9	~TCDF, 2,3,7,8-	mg/kg	5.77E-03	5.67E-03	5.67E-03	2.89E-02	5.67E-03	5.67E-03	1.58E-02	1.57E-02	1.57E-02
100-41-4	Ethylbenzene	mg/kg	2.59E+03	3.31E+04	2.59E+03	1.30E+04	3.31E+04	1.30E+04	2.66E+03	6.14E+04	2.66E+03
206-44-0	Fluoranthene	mg/kg	-	2.02E+04	2.02E+04	-	2.02E+04	2.02E+04	-	2.76E+04	2.76E+04
86-73-7	Fluorene	mg/kg	-	2.02E+04	2.02E+04	-	2.02E+04	2.02E+04	-	2.76E+04	2.76E+04
118-74-1	Hexachlorobenzene	mg/kg	4.66E+01	9.86E+00	9.86E+00	2.33E+02	9.86E+00	9.86E+00	1.26E+02	7.01E+01	7.01E+01
91-20-3	Naphthalene	mg/kg	3.34E+02	1.98E+03	3.34E+02	1.67E+03	1.98E+03	1.67E+03	4.06E+02	1.61E+03	4.06E+02
88-74-4	Nitroaniline, 2-	mg/kg	-	5.67E+03	5.67E+03	-	5.67E+03	5.67E+03	-	8.61E+03	8.61E+03
621-64-7	Nitroso-di-N-propylamine, N-	mg/kg	7.58E+00	-	7.58E+00	3.79E+01	-	3.79E+01	1.16E+01	-	1.16E+01
87-86-5	Pentachlorophenol	mg/kg	8.11E+01	1.74E+03	8.11E+01	4.06E+02	1.74E+03	4.06E+02	8.77E+01	1.88E+03	8.77E+01
85-01-8	Phenanthrene ^d	mg/kg	-	3.03E+04	3.03E+04	-	3.03E+04	3.03E+04	-	4.14E+04	4.14E+04
1336-36-3	Polychlorinated Biphenyls, Total	mg/kg	2.24E+01	-	2.24E+01	1.12E+02	-	1.12E+02	2.93E+01	-	2.93E+01
12674-11-2	~Aroclor 1016	mg/kg	6.40E+02	3.39E+01	3.39E+01	3.20E+03	3.39E+01	3.39E+01	8.39E+02	4.50E+01	4.50E+01
11104-28-2	~Aroclor 1221	mg/kg	2.19E+01	-	2.19E+01	1.09E+02	-	1.09E+02	2.81E+01	-	2.81E+01
11141-16-5	~Aroclor 1232	mg/kg	2.12E+01	-	2.12E+01	1.06E+02	-	1.06E+02	2.67E+01	-	2.67E+01
53469-21-9	~Aroclor 1242	mg/kg	2.24E+01	-	2.24E+01	1.12E+02	-	1.12E+02	2.94E+01	-	2.94E+01
12672-29-6	~Aroclor 1248	mg/kg	2.24E+01	-	2.24E+01	1.12E+02	-	1.12E+02	2.93E+01	-	2.93E+01
11097-69-1	~Aroclor 1254	mg/kg	2.25E+01	9.72E+00	9.72E+00	1.12E+02	9.72E+00	9.72E+00	2.96E+01	1.29E+01	1.29E+01
11096-82-5	~Aroclor 1260	mg/kg	2.26E+01	-	2.26E+01	1.13E+02	-	1.13E+02	2.98E+01	-	2.98E+01

CAS			Ou	tdoor Work	ker	Exc	avation Wo	rker	Industrial Worker			
Number	Analyte	Units	Cancer	Hazard	Action	Cancer	Hazard	Action	Cancer	Hazard	Action	
50-32-8	Polycyclic aromatic hydrocarbons, Total Carcinogenic ^g	mg/kg	4.71E+01	1.51E+02	4.71E+01	2.35E+02	1.51E+02	1.51E+02	6.43E+01	2.06E+02	6.43E+01	
56-55-3	~Benz[a]anthracene	mg/kg	4.69E+02	-	4.69E+02	2.35E+03	1	2.35E+03	6.39E+02	-	6.39E+02	
50-32-8	~Benzo[a]pyrene	mg/kg	4.71E+01	1.51E+02	4.71E+01	2.35E+02	1.51E+02	1.51E+02	6.43E+01	2.06E+02	6.43E+01	
205-99-2	~Benzo[b]fluoranthene	mg/kg	4.71E+02	-	4.71E+02	2.35E+03	1	2.35E+03	6.43E+02	-	6.43E+02	
207-08-9	~Benzo[k]fluoranthene	mg/kg	4.71E+03	-	4.71E+03	2.35E+04	1	2.35E+04	6.43E+03	-	6.43E+03	
218-01-9	~Chrysene	mg/kg	4.71E+04	-	4.71E+04	1.00E+05	-	1.00E+05	6.43E+04	-	6.43E+04	
53-70-3	~Dibenz[a,h]anthracene	mg/kg	4.71E+01	-	4.71E+01	2.35E+02	-	2.35E+02	6.43E+01	-	6.43E+01	
193-39-5	~Indeno[1,2,3-cd]pyrene	mg/kg	4.71E+02	-	4.71E+02	2.35E+03	-	2.35E+03	6.43E+02	-	6.43E+02	
129-00-0	Pyrene	mg/kg	-	1.52E+04	1.52E+04	-	1.52E+04	1.52E+04	-	2.07E+04	2.07E+04	
127-18-4	Tetrachloroethylene	mg/kg	1.12E+04	1.30E+03	1.30E+03	5.58E+04	1.30E+03	1.30E+03	1.07E+04	1.20E+03	1.20E+03	
108-88-3	Toluene ^e	mg/kg	-	6.54E+04	6.54E+04	-	6.54E+04	6.54E+04	-	1.00E+05	1.00E+05	
76-13-1	Trichloro-1,2,2-trifluoroethane, 1,1,2- (Freon-113) ^e	mg/kg	-	1.00E+05	1.00E+05	-	1.00E+05	1.00E+05	-	8.43E+04	8.43E+04	
71-55-6	Trichloroethane, 1,1,1-	mg/kg	-	1.00E+05	1.00E+05	-	1.00E+05	1.00E+05	-	1.00E+05	1.00E+05	
79-00-5	Trichloroethane, 1,1,2-	mg/kg	5.11E+02	2.55E+01	2.55E+01	2.56E+03	2.55E+01	2.55E+01	5.28E+02	1.90E+01	1.90E+01	
79-01-6	Trichloroethylene	mg/kg	6.17E+02	6.78E+01	6.78E+01	3.09E+03	6.78E+01	6.78E+01	6.31E+02	5.70E+01	5.70E+01	
75-01-4	Vinyl Chloride	mg/kg	9.44E+01	9.31E+02	9.44E+01	4.72E+02	9.31E+02	4.72E+02	2.06E+02	9.59E+02	2.06E+02	
108-38-3	Xylene, m-	mg/kg	-	9.25E+03	9.25E+03	-	9.25E+03	9.25E+03	-	7.14E+03	7.14E+03	
95-47-6	Xylene, o-	mg/kg	-	1.08E+04	1.08E+04	-	1.08E+04	1.08E+04	-	8.43E+03	8.43E+03	
106-42-3	Xylene, p-	mg/kg	-	9.43E+03	9.43E+03	-	9.43E+03	9.43E+03	-	7.29E+03	7.29E+03	
1330-20-7	Xylene, Mixture	mg/kg	-	9.69E+03	9.69E+03	-	9.69E+03	9.69E+03	-	7.50E+03	7.50E+03	

CAS			Recreational User	Adult Recre	eational User	Child Recre	ational User	Teen Recrea	ational User
Number	Analyte	Units	Cancer ^h	Hazard	Action	Hazard	Action	Hazard	Action
7429-90-5	Aluminum	mg/kg	-	1.00E+05	1.00E+05	1.00E+05	1.00E+05	1.00E+05	1.00E+05
7440-36-0	Antimony (metallic)	mg/kg	-	3.37E+03	3.37E+03	2.35E+02	2.35E+02	1.38E+03	1.38E+03
7440-38-2	Arsenic, Inorganic	mg/kg	8.09E+01	1.05E+03	8.09E+01	1.84E+02	8.09E+01	3.62E+02	8.09E+01
7440-39-3	Barium	mg/kg	-	1.00E+05	1.00E+05	1.00E+05	1.00E+05	1.00E+05	1.00E+05
7440-41-7	Beryllium and compounds	mg/kg	1.00E+05	1.66E+04	1.66E+04	1.17E+03	1.17E+03	6.84E+03	6.84E+03
7440-42-8	Boron And Borates Only	mg/kg	-	1.00E+05	1.00E+05	1.00E+05	1.00E+05	1.00E+05	1.00E+05
7440-43-9	Cadmium (Diet)	mg/kg	1.00E+05	2.47E+02	2.47E+02	3.98E+01	3.98E+01	8.60E+01	8.60E+01
16065-83-1	Chromium(III), Insoluble Salts	mg/kg	-	1.00E+05	1.00E+05	1.00E+05	1.00E+05	1.00E+05	1.00E+05
18540-29-9	Chromium(VI)	mg/kg	7.47E+01	2.52E+04	7.47E+01	1.76E+03	7.47E+01	1.03E+04	7.47E+01
7440-47-3	Chromium (Total) ^a	-	-	-	-	-	-	-	-
7440-48-4	Cobalt	mg/kg	1.00E+05	2.51E+03	2.51E+03	1.76E+02	1.76E+02	1.03E+03	1.03E+03
7440-50-8	Copper	mg/kg	-	1.00E+05	1.00E+05	2.35E+04	2.35E+04	1.00E+05	1.00E+05
16984-48-8	Fluoride	mg/kg	-	1.00E+05	1.00E+05	2.35E+04	2.35E+04	1.00E+05	1.00E+05
7439-89-6	Iron	mg/kg	-	1.00E+05	1.00E+05	1.00E+05	1.00E+05	1.00E+05	1.00E+05
7439-92-1	Lead ^b	mg/kg	-	-	4.00E+02	-	4.00E+02	-	4.00E+02
7439-96-5	Manganese (Non-Diet)	mg/kg	-	1.00E+05	1.00E+05	1.40E+04	1.40E+04	8.00E+04	8.00E+04
Various	Mercury, Inorganic Salts	mg/kg	-	2.53E+03	2.53E+03	1.76E+02	1.76E+02	1.03E+03	1.03E+03
7439-98-7	Molybdenum	mg/kg	-	4.21E+04	4.21E+04	2.93E+03	2.93E+03	1.72E+04	1.72E+04
7440-02-0	Nickel Soluble Salts	mg/kg	1.00E+05	1.00E+05	1.00E+05	1.17E+04	1.17E+04	6.78E+04	6.78E+04
7782-49-2	Selenium	mg/kg	-	4.21E+04	4.21E+04	2.93E+03	2.93E+03	1.72E+04	1.72E+04
7440-22-4	Silver	mg/kg	-	4.21E+04	4.21E+04	2.93E+03	2.93E+03	1.72E+04	1.72E+04
7440-28-0	Thallium (Soluble Salts)	mg/kg	-	8.42E+01	8.42E+01	5.87E+00	5.87E+00	3.44E+01	3.44E+01
N/A	Uranium (Insoluble Compounds) ^c	mg/kg	-	2.51E+04	2.51E+04	1.76E+03	1.76E+03	1.03E+04	1.03E+04
7440-61-1	Uranium (Soluble Salts) ^c	mg/kg	-	1.68E+03	1.68E+03	1.17E+02	1.17E+02	6.88E+02	6.88E+02
7440-62-2	Vanadium and Compounds	mg/kg	-	4.22E+04	4.22E+04	2.95E+03	2.95E+03	1.73E+04	1.73E+04
7440-66-6	Zinc and Compounds	mg/kg	-	1.00E+05	1.00E+05	1.00E+05	1.00E+05	1.00E+05	1.00E+05
83-32-9	Acenaphthene	mg/kg	-	5.72E+04	5.72E+04	1.38E+04	1.38E+04	1.92E+04	1.92E+04
208-96-8	Acenaphthylene ^d	mg/kg	-	5.72E+04	5.72E+04	1.38E+04	1.38E+04	1.92E+04	1.92E+04
107-13-1	Acrylonitrile	mg/kg	1.80E+02	7.70E+02	1.80E+02	5.25E+02	1.80E+02	5.68E+02	1.80E+02
120-12-7	Anthracene	mg/kg	-	1.00E+05	1.00E+05	6.92E+04	6.92E+04	9.60E+04	9.60E+04
71-43-2	Benzene	mg/kg	1.09E+03	4.63E+03	1.09E+03	1.48E+03	1.09E+03	3.09E+03	1.09E+03
117-81-7	Bis(2-ethylhexyl)phthalate ^e	mg/kg	3.32E+03	2.40E+04	3.32E+03	5.37E+03	3.32E+03	8.10E+03	3.32E+03
75-27-4	Bromodichloromethane	mg/kg	3.49E+02	6.74E+04	3.49E+02	4.69E+03	3.49E+02	2.75E+04	3.49E+02
86-74-8	Carbazole	mg/kg	2.32E+03	-	2.32E+03	-	2.32E+03	-	2.32E+03
56-23-5	Carbon Tetrachloride	mg/kg	6.72E+02	6.17E+03	6.72E+02	1.65E+03	6.72E+02	3.99E+03	6.72E+02
67-66-3	Chloroform	mg/kg	3.96E+02	1.12E+04	3.96E+02	3.65E+03	3.96E+02	7.53E+03	3.96E+02
75-71-8	Dichlorodifluoromethane (Freon-12) ^e	mg/kg	-	4.24E+03	4.23E+03	3.07E+03	3.06E+03	3.14E+03	3.15E+03

CAS			Recreational User	Adult Recre	eational User	Child Recre	ational User	Teen Recrea	tional User
Number	Analyte	Units	Cancer ^h	Hazard	Action	Hazard	Action	Hazard	Action
75-34-3	Dichloroethane, 1,1-e	mg/kg	4.18E+03	5.11E+04	4.18E+03	2.93E+04	4.18E+03	3.70E+04	4.18E+03
107-06-2	Dichloroethane, 1,2-	mg/kg	4.86E+02	1.57E+03	4.86E+02	8.96E+02	4.86E+02	1.14E+03	4.86E+02
75-35-4	Dichloroethylene, 1,1-	mg/kg	-	1.14E+04	1.14E+04	6.70E+03	6.69E+03	8.26E+03	8.25E+03
540-59-0	Dichloroethylene, 1,2- (Mixed Isomers)	mg/kg	-	7.58E+04	7.58E+04	5.28E+03	5.28E+03	3.10E+04	3.09E+04
156-59-2	Dichloroethylene, cis-1,2-	mg/kg	-	3.89E+03	3.89E+03	8.94E+02	8.94E+02	2.43E+03	2.43E+03
156-60-5	Dichloroethylene, trans-1,2-	mg/kg	-	3.46E+03	3.46E+03	2.14E+03	2.14E+03	2.53E+03	2.53E+03
60-57-1	Dieldrin	mg/kg	2.90E+00	5.99E+01	2.90E+00	1.34E+01	2.90E+00	2.02E+01	2.90E+00
1746-01-6	Dioxins/Furans, Total (as TCDD) ^f	mg/kg	7.22E-04	2.10E-03	7.22E-04	3.03E-04	3.03E-04	7.41E-04	7.22E-04
37871-00-4	~HpCDD	mg/kg	7.23E-02	2.10E-01	7.23E-02	3.03E-02	3.03E-02	7.41E-02	7.23E-02
38998-75-3	~HpCDF, 2,3,7,8-	mg/kg	7.24E-02	2.10E-01	7.24E-02	3.03E-02	3.03E-02	7.41E-02	7.24E-02
34465-46-8	~HxCDD, 2,3,7,8-	mg/kg	7.25E-03	1.94E-02	7.25E-03	3.03E-03	3.03E-03	7.41E-03	7.25E-03
55684-94-1	~HxCDF, 2,3,7,8-	mg/kg	7.25E-03	2.10E-02	7.25E-03	3.03E-03	3.03E-03	7.41E-03	7.25E-03
3268-87-9	~OCDD	mg/kg	2.42E+00	7.00E+00	2.42E+00	1.01E+00	1.01E+00	2.47E+00	2.42E+00
39001-02-0	~OCDF	mg/kg	2.42E+00	7.00E+00	2.42E+00	1.01E+00	1.01E+00	2.47E+00	2.42E+00
36088-22-9	~PeCDD, 2,3,7,8-	mg/kg	7.25E-04	2.10E-03	7.25E-04	3.03E-04	3.03E-04	7.41E-04	7.25E-04
57117-41-6	~PeCDF, 1,2,3,7,8-	mg/kg	2.42E-02	7.00E-02	2.42E-02	1.01E-02	1.01E-02	2.47E-02	2.42E-02
57117-31-4	~PeCDF, 2,3,4,7,8-	mg/kg	2.42E-03	7.00E-03	2.42E-03	1.01E-03	1.01E-03	2.47E-03	2.42E-03
1746-01-6	~TCDD, 2,3,7,8-	mg/kg	7.22E-04	2.10E-03	7.22E-04	3.03E-04	3.03E-04	7.41E-04	7.22E-04
51207-31-9	~TCDF, 2,3,7,8-	mg/kg	7.23E-03	2.10E-02	7.23E-03	3.03E-03	3.03E-03	7.41E-03	7.23E-03
100-41-4	Ethylbenzene	mg/kg	5.46E+03	1.00E+05	5.46E+03	2.58E+04	5.46E+03	1.00E+05	5.46E+03
206-44-0	Fluoranthene	mg/kg	-	3.81E+04	3.81E+04	9.23E+03	9.23E+03	1.28E+04	1.28E+04
86-73-7	Fluorene	mg/kg	-	3.81E+04	3.81E+04	9.23E+03	9.23E+03	1.28E+04	1.28E+04
118-74-1	Hexachlorobenzene	mg/kg	8.86E+01	8.42E+01	8.42E+01	5.87E+00	5.87E+00	3.44E+01	3.44E+01
91-20-3	Naphthalene	mg/kg	2.99E+02	5.13E+03	2.99E+02	2.45E+03	2.99E+02	2.87E+03	2.99E+02
88-74-4	Nitroaniline, 2-	mg/kg	-	1.19E+04	1.19E+04	2.68E+03	2.68E+03	4.04E+03	4.04E+03
621-64-7	Nitroso-di-N-propylamine, N-	mg/kg	6.63E+00	-	6.63E+00	-	6.63E+00	ı	6.63E+00
87-86-5	Pentachlorophenol	mg/kg	5.56E+01	2.62E+03	5.56E+01	7.39E+02	5.56E+01	8.71E+02	5.56E+01
85-01-8	Phenanthrene ^d	mg/kg	-	5.72E+04	5.72E+04	1.38E+04	1.38E+04	1.92E+04	1.92E+04
1336-36-3	Polychlorinated Biphenyls, Total	mg/kg	1.79E+01	-	1.79E+01	-	1.79E+01	-	1.79E+01
12674-11-2	~Aroclor 1016	mg/kg	5.12E+02	6.24E+01	6.24E+01	1.54E+01	1.54E+01	2.09E+01	2.09E+01
11104-28-2	~Aroclor 1221	mg/kg	1.77E+01	-	1.77E+01	-	1.77E+01	1	1.77E+01
11141-16-5	~Aroclor 1232	mg/kg	1.76E+01	-	1.76E+01	-	1.76E+01	1	1.76E+01
53469-21-9	~Aroclor 1242	mg/kg	1.79E+01	-	1.79E+01	-	1.79E+01	-	1.79E+01
12672-29-6	~Aroclor 1248	mg/kg	1.79E+01	-	1.79E+01	-	1.79E+01		1.79E+01
11097-69-1	~Aroclor 1254	mg/kg	1.79E+01	1.78E+01	1.78E+01	4.41E+00	4.41E+00	5.99E+00	5.99E+00
11096-82-5	~Aroclor 1260	mg/kg	1.80E+01	-	1.80E+01	-	1.80E+01	-	1.80E+01

CAS			Recreational User	Adult Recre	eational User	Child Recre	ational User	Teen Recreational User	
Number	Analyte	Units	Cancer ^h	Hazard	Action	Hazard	Action	Hazard	Action
50-32-8	Polycyclic aromatic hydrocarbons, Total Carcinogenic ^g	mg/kg	1.09E+01	2.85E+02	1.09E+01	6.92E+01	1.09E+01	9.59E+01	1.09E+01
56-55-3	~Benz[a]anthracene	mg/kg	1.09E+02	-	1.09E+02	-	1.09E+02	-	1.09E+02
50-32-8	~Benzo[a]pyrene	mg/kg	1.09E+01	2.85E+02	1.09E+01	6.92E+01	1.09E+01	9.59E+01	1.09E+01
205-99-2	~Benzo[b]fluoranthene	mg/kg	1.09E+02	-	1.09E+02	-	1.09E+02	-	1.09E+02
207-08-9	~Benzo[k]fluoranthene	mg/kg	1.09E+03	-	1.09E+03	-	1.09E+03	-	1.09E+03
218-01-9	~Chrysene	mg/kg	1.09E+04	-	1.09E+04	-	1.09E+04	-	1.09E+04
53-70-3	~Dibenz[a,h]anthracene	mg/kg	1.09E+01	-	1.09E+01	-	1.09E+01	-	1.09E+01
193-39-5	~Indeno[1,2,3-cd]pyrene	mg/kg	1.09E+02	-	1.09E+02	-	1.09E+02	-	1.09E+02
129-00-0	Pyrene	mg/kg	-	2.86E+04	2.86E+04	6.92E+03	6.92E+03	9.60E+03	9.60E+03
127-18-4	Tetrachloroethylene	mg/kg	2.38E+04	4.34E+03	4.34E+03	1.76E+03	1.76E+03	3.01E+03	3.01E+03
108-88-3	Toluene ^e	mg/kg	-	1.00E+05	1.00E+05	4.43E+04	4.43E+04	1.00E+05	1.00E+05
76-13-1	Trichloro-1,2,2-trifluoroethane, 1,1,2- (Freon-113) ^e	mg/kg	-	1.00E+05	1.00E+05	1.00E+05	1.00E+05	1.00E+05	1.00E+05
71-55-6	Trichloroethane, 1,1,1-	mg/kg	-	1.00E+05	1.00E+05	1.00E+05	1.00E+05	1.00E+05	1.00E+05
79-00-5	Trichloroethane, 1,1,2-	mg/kg	1.07E+03	7.28E+01	7.28E+01	5.30E+01	5.30E+01	5.40E+01	5.40E+01
79-01-6	Trichloroethylene	mg/kg	8.32E+02	2.12E+02	2.12E+02	1.06E+02	1.06E+02	1.51E+02	1.51E+02
75-01-4	Vinyl Chloride	mg/kg	6.78E+00	3.35E+03	6.78E+00	1.09E+03	6.78E+00	2.25E+03	6.78E+00
108-38-3	Xylene, m-	mg/kg	-	2.72E+04	2.72E+04	1.75E+04	1.75E+04	1.99E+04	1.99E+04
95-47-6	Xylene, o-	mg/kg	-	3.21E+04	3.21E+04	2.01E+04	2.01E+04	2.34E+04	2.34E+04
106-42-3	Xylene, p-	mg/kg	-	3.20E+04	3.20E+04	1.78E+04	1.78E+04	2.03E+04	2.03E+04
1330-20-7	Xylene, Mixture	mg/kg	-	2.77E+04	2.77E+04	1.82E+04	1.82E+04	2.09E+04	2.09E+04

CAS			Resident	Adult F	Resident	Child R	tesident
Number	Analyte	Units	Cancer ^h	Hazard	Action	Hazard	Action
7429-90-5	Aluminum	mg/kg	-	1.00E+05	1.00E+05	1.00E+05	1.00E+05
7440-36-0	Antimony (metallic)	mg/kg	-	9.99E+02	9.99E+02	9.39E+01	9.39E+01
7440-38-2	Arsenic, Inorganic	mg/kg	3.56E+01	3.09E+02	3.56E+01	7.35E+01	3.56E+01
7440-39-3	Barium	mg/kg	-	1.00E+05	1.00E+05	4.59E+04	4.59E+04
7440-41-7	Beryllium and compounds	mg/kg	1.00E+05	4.74E+03	4.74E+03	4.68E+02	4.68E+02
7440-42-8	Boron And Borates Only	mg/kg	-	1.00E+05	1.00E+05	4.68E+04	4.68E+04
7440-43-9	Cadmium (Diet)	mg/kg	1.00E+05	7.32E+01	7.32E+01	1.59E+01	1.59E+01
16065-83-1	Chromium(III), Insoluble Salts	mg/kg	-	1.00E+05	1.00E+05	3.51E+05	3.51E+05
18540-29-9	Chromium(VI)	mg/kg	3.01E+01	7.38E+03	3.01E+01	7.02E+02	3.01E+01
7440-47-3	Chromium (Total) ^a	-	-	-	-	-	-
7440-48-4	Cobalt	mg/kg	4.24E+04	7.29E+02	7.29E+02	7.02E+01	7.02E+01
7440-50-8	Copper	mg/kg	-	1.00E+05	1.00E+05	9.39E+03	9.39E+03
16984-48-8	Fluoride	mg/kg	-	9.99E+04	9.99E+04	9.39E+03	9.39E+03
7439-89-6	Iron	mg/kg	-	1.00E+05	1.00E+05	1.00E+05	1.00E+05
7439-92-1	Lead ^b	mg/kg	-	-	4.00E+02	-	4.00E+02
7439-96-5	Manganese (Non-Diet)	mg/kg	-	4.68E+04	4.68E+04	5.49E+03	5.49E+03
Various	Mercury, Inorganic Salts	mg/kg	-	7.50E+02	7.50E+02	7.05E+01	7.05E+01
7439-98-7	Molybdenum	mg/kg	-	1.25E+04	1.25E+04	1.17E+03	1.17E+03
7440-02-0	Nickel Soluble Salts	mg/kg	1.00E+05	4.44E+04	4.44E+04	4.65E+03	4.65E+03
7782-49-2	Selenium	mg/kg	-	1.25E+04	1.25E+04	1.17E+03	1.17E+03
7440-22-4	Silver	mg/kg	-	1.25E+04	1.25E+04	1.17E+03	1.17E+03
7440-28-0	Thallium (Soluble Salts)	mg/kg	-	2.50E+01	2.50E+01	2.35E+00	2.35E+00
N/A	Uranium (Insoluble Compounds) ^c	mg/kg	-	7.20E+03	7.20E+03	7.02E+02	7.02E+02
7440-61-1	Uranium (Soluble Salts) ^c	mg/kg	-	4.98E+02	4.98E+02	4.68E+01	4.68E+01
7440-62-2	Vanadium and Compounds	mg/kg	-	1.22E+04	1.22E+04	1.18E+03	1.18E+03
7440-66-6	Zinc and Compounds	mg/kg	-	1.00E+05	1.00E+05	7.05E+04	7.05E+04
83-32-9	Acenaphthene	mg/kg	-	1.70E+04	1.70E+04	5.55E+03	5.55E+03
208-96-8	Acenaphthylene ^d	mg/kg	-	1.70E+04	1.70E+04	5.55E+03	5.55E+03
107-13-1	Acrylonitrile	mg/kg	2.55E+01	4.80E+01	2.55E+01	4.71E+01	2.55E+01
120-12-7	Anthracene	mg/kg	-	8.49E+04	8.49E+04	2.77E+04	2.77E+04
71-43-2	Benzene	mg/kg	1.16E+02	3.21E+02	1.16E+02	2.45E+02	1.16E+02
117-81-7	Bis(2-ethylhexyl)phthalate ^e	mg/kg	1.49E+03	7.11E+03	1.49E+03	2.15E+03	1.49E+03
75-27-4	Bromodichloromethane	mg/kg	2.93E+01	2.00E+04	2.93E+01	1.88E+03	2.93E+01
86-74-8	Carbazole	mg/kg	1.04E+03	-	1.04E+03	-	1.04E+03
56-23-5	Carbon Tetrachloride	mg/kg	6.53E+01	4.47E+02	6.53E+01	3.12E+02	6.53E+01
67-66-3	Chloroform	mg/kg	3.16E+01	7.77E+02	3.16E+01	5.97E+02	3.16E+01
75-71-8	Dichlorodifluoromethane (Freon-12) ^e	mg/kg	-	2.63E+02	2.63E+02	2.62E+02	2.62E+02

Table A.1a. Soil/Sediment Action Levels for Significant COPCs at PGDP (Continued) (Values were calculated in October 2022 and are based on best available information.)

CAS			Resident	Adult R	Resident	Child R	Resident
Number	Analyte	Units	Cancer ^h	Hazard	Action	Hazard	Action
75-34-3	Dichloroethane, 1,1-e	mg/kg	3.55E+02	3.24E+03	3.55E+02	3.06E+03	3.55E+02
107-06-2	Dichloroethane, 1,2-	mg/kg	4.64E+01	9.96E+01	4.64E+01	9.36E+01	4.64E+01
75-35-4	Dichloroethylene, 1,1-	mg/kg	-	7.20E+02	7.20E+02	6.81E+02	6.81E+02
540-59-0	Dichloroethylene, 1,2- (Mixed Isomers)	mg/kg	-	2.25E+04	2.25E+04	2.11E+03	2.11E+03
156-59-2	Dichloroethylene, cis-1,2-	mg/kg	-	2.94E+02	5.01E+03	4.68E+02	4.68E+02
156-60-5	Dichloroethylene, trans-1,2-	mg/kg	-	2.18E+02	2.18E+02	2.09E+02	2.09E+02
60-57-1	Dieldrin	mg/kg	1.30E+00	1.78E+01	1.30E+00	5.37E+00	1.30E+00
1746-01-6	Dioxins/Furans, Total (as TCDD) ^f	mg/kg	3.08E-04	6.21E-04	3.08E-04	1.21E-04	1.21E-04
37871-00-4	~HpCDD	mg/kg	3.09E-02	6.21E-02	3.09E-02	1.21E-02	1.21E-02
38998-75-3	~HpCDF, 2,3,7,8-	mg/kg	3.12E-02	6.24E-02	3.12E-02	1.21E-02	1.21E-02
34465-46-8	~HxCDD, 2,3,7,8-	mg/kg	3.14E-03	6.24E-03	3.14E-03	1.21E-03	1.21E-03
55684-94-1	~HxCDF, 2,3,7,8-	mg/kg	3.14E-03	6.24E-03	3.14E-03	1.21E-03	1.21E-03
3268-87-9	~OCDD	mg/kg	1.05E+00	2.08E+00	1.05E+00	4.05E-01	4.05E-01
39001-02-0	~OCDF	mg/kg	1.05E+00	2.08E+00	1.05E+00	4.05E-01	4.05E-01
36088-22-9	~PeCDD, 2,3,7,8-	mg/kg	3.14E-04	6.24E-04	3.14E-04	1.21E-04	1.21E-04
57117-41-6	~PeCDF, 1,2,3,7,8-	mg/kg	1.05E-02	2.08E-02	1.05E-02	4.05E-03	4.05E-03
57117-31-4	~PeCDF, 2,3,4,7,8-	mg/kg	1.05E-03	2.08E-03	1.05E-03	4.05E-04	4.05E-04
1746-01-6	~TCDD, 2,3,7,8-	mg/kg	3.08E-04	6.21E-04	3.08E-04	1.21E-04	1.21E-04
51207-31-9	~TCDF, 2,3,7,8-	mg/kg	3.09E-03	6.21E-03	3.09E-03	1.21E-03	1.21E-03
100-41-4	Ethylbenzene	mg/kg	5.78E+02	1.55E+04	5.78E+02	7.05E+03	5.78E+02
206-44-0	Fluoranthene	mg/kg	-	1.13E+04	1.13E+04	3.69E+03	3.69E+03
86-73-7	Fluorene	mg/kg	-	1.13E+04	1.13E+04	3.69E+03	3.69E+03
118-74-1	Hexachlorobenzene	mg/kg	2.12E+01	2.50E+01	2.12E+01	2.35E+00	2.12E+01
91-20-3	Naphthalene	mg/kg	1.04E+02	4.05E+02	1.04E+02	3.51E+02	1.04E+02
88-74-4	Nitroaniline, 2-	mg/kg	-	3.51E+03	3.51E+03	1.07E+03	1.07E+03
621-64-7	Nitroso-di-N-propylamine, N-	mg/kg	2.97E+00	-	2.97E+00	-	2.97E+00
87-86-5	Pentachlorophenol	mg/kg	2.54E+01	7.77E+02	2.54E+01	2.96E+02	2.54E+01
85-01-8	Phenanthrene ^d	mg/kg	-	1.70E+04	1.70E+04	5.55E+03	5.55E+03
1336-36-3	Polychlorinated Biphenyls, Total	mg/kg	7.88E+00	-	7.88E+00	-	7.88E+00
12674-11-2	~Aroclor 1016	mg/kg	2.26E+02	1.85E+01	1.85E+01	6.18E+00	6.18E+00
11104-28-2	~Aroclor 1221	mg/kg	7.52E+00	-	7.52E+00	-	7.52E+00
11141-16-5	~Aroclor 1232	mg/kg	7.08E+00	-	7.08E+00	-	7.08E+00
53469-21-9	~Aroclor 1242	mg/kg	7.91E+00	-	7.91E+00	-	7.91E+00
12672-29-6	~Aroclor 1248	mg/kg	7.88E+00	-	7.88E+00	-	7.88E+00
11097-69-1	~Aroclor 1254	mg/kg	7.97E+00	5.31E+00	5.31E+00	1.76E+00	1.76E+00
11096-82-5	~Aroclor 1260	mg/kg	8.03E+00	-	8.03E+00	-	8.03E+00

Table A.1a. Soil/Sediment Action Levels for Significant COPCs at PGDP (Continued)

(Values were calculated in October 2022 and are based on best available information.)

CAS			Resident	Adult F	Resident	Child R	Resident
Number	Analyte	Units	Cancer ^h	Hazard	Action	Hazard	Action
50-32-8	Polycyclic aromatic hydrocarbons, Total Carcinogenic ^g	mg/kg	4.78E+00	8.40E+01	4.78E+00	2.76E+01	4.78E+00
56-55-3	~Benz[a]anthracene	mg/kg	4.75E+01	-	4.75E+01	1	4.75E+01
50-32-8	~Benzo[a]pyrene	mg/kg	4.78E+00	8.40E+01	4.78E+00	2.76E+01	4.78E+00
205-99-2	~Benzo[b]fluoranthene	mg/kg	4.78E+01	-	4.78E+01	•	4.78E+01
207-08-9	~Benzo[k]fluoranthene	mg/kg	4.78E+02	-	4.78E+02	•	4.78E+02
218-01-9	~Chrysene	mg/kg	4.78E+03	-	4.78E+03	•	4.78E+03
53-70-3	~Dibenz[a,h]anthracene	mg/kg	4.78E+00	-	4.78E+00	•	4.78E+00
193-39-5	~Indeno[1,2,3-cd]pyrene	mg/kg	4.78E+01	-	4.78E+01	•	4.78E+01
129-00-0	Pyrene	mg/kg	-	8.49E+03	8.49E+03	2.77E+03	2.77E+03
127-18-4	Tetrachloroethylene	mg/kg	2.36E+03	2.88E+02	2.88E+02	2.43E+02	2.43E+02
108-88-3	Toluene ^e	mg/kg	-	5.01E+04	5.01E+04	1.47E+04	1.47E+04
76-13-1	Trichloro-1,2,2-trifluoroethane, 1,1,2- (Freon-113) ^e	mg/kg	-	2.01E+04	2.01E+04	2.01E+04	2.01E+04
71-55-6	Trichloroethane, 1,1,1-	mg/kg	-	2.57E+04	2.57E+04	2.45E+04	2.45E+04
79-00-5	Trichloroethane, 1,1,2-	mg/kg	1.15E+02	4.50E+00	4.50E+00	4.50E+00	4.50E+00
79-01-6	Trichloroethylene	mg/kg	9.43E+01	1.37E+01	1.37E+01	1.24E+01	1.24E+01
75-01-4	Vinyl Chloride	mg/kg	5.92E+00	2.30E+02	5.92E+00	1.79E+02	5.92E+00
108-38-3	Xylene, m-	mg/kg	-	1.70E+03	1.70E+03	1.65E+03	1.65E+03
95-47-6	Xylene, o-	mg/kg	-	2.01E+03	2.01E+03	1.94E+03	1.94E+03
106-42-3	Xylene, p-	mg/kg	-	1.74E+03	1.74E+03	1.68E+03	1.68E+03
1330-20-7	Xylene, Mixture	mg/kg	-	1.79E+03	1.79E+03	1.73E+03	1.73E+03

NOTES: An HI of 3 is used for the AL derivation because the range of values for HI (based on RGO tables) are 0.1, 1, and 3.

Values are provided in these tables for significant COPCs for PGDP. Values for other COPCs can be obtained using the RAIS Chemical PRG online calculator, as modified using PGDP-specific inputs.

^a Chromium (Total) AL should utilize Chromium III or Chromium VI, as appropriate.

^b Lead values should be checked prior to use to ensure they are still current.

^c Based on recommendation from the EPA, ALs for uranium (soluble salts) now use the oral reference dose (RfD) and the reference concentration (RfC) for soluble compounds of uranium derived from the Agency for Toxic Substances and Disease Registry (ATSDR) (EPA 2016). ALs for uranium (insoluble compounds) use the RfD for uranium (soluble salts), which is available in Integrated Risk Information System (IRIS); the RfC for insoluble compounds of uranium are derived from ATSDR.

d Acenaphthylene and phenanthrene use values for acenaphthene as a surrogate.

^e Analytes are not PGDP-significant COPCs (Table 2.1), but are provided for project support.

f Total dioxins/furans uses values for 2,3,7,8-TCDD, see screening note 9f in the Appendix A introduction, "Screening Levels," on pages A-3-A-5.

g Total carcinogenic PAHs uses values for BaP, see screening note 9d in the Appendix A introduction, "Screening Levels," on pages A-3-A-5.

h For the recreational user and the resident, ELCRs (i.e., cancer ALs) were calculated using the child/teen/adult or child/adult age-adjusted lifetime scenario, respectively.

			(Outdoor Work	er	Ex	xcavation Wor	ker	Iı	ndustrial Wor	ker
CAS Number	Analyte	Units	Peak Risk Infinite ^a	Peak Risk 1,000 Years	Secular Equilibrium	Peak Risk Infinite ^a	Peak Risk 1,000 Years	Secular Equilibrium	Peak Risk Infinite ^a	Peak Risk 1,000 Years	Secular Equilibrium
14596-10-2	Am-241	pCi/g	4.47E+02	4.47E+02	3.38E+01	2.20E+03	2.20E+03	1.69E+02	1.61E+03	1.61E+03	3.99E+01
10045-97-3	Cs-137	pCi/g	5.94E+01	5.94E+01	4.52E+01	2.39E+02	2.39E+02	2.26E+02	4.53E+01	4.53E+01	3.44E+01
13994-20-2	Np-237	pCi/g	4.49E+01	1.20E+02	3.66E+01	2.25E+02	5.99E+02	1.83E+02	5.00E+01	1.01E+02	4.09E+01
13981-16-3	Pu-238	pCi/g	4.20E+02	4.20E+02	7.81E+00	1.94E+03	1.94E+03	3.90E+01	2.67E+03	2.67E+03	9.63E+00
15117-48-3	Pu-239	pCi/g	3.67E+02	3.67E+02	2.87E+01	1.83E+03	1.83E+03	1.44E+02	2.32E+03	2.32E+03	3.44E+01
14119-33-6	Pu-240	pCi/g	3.68E+02	3.68E+02	8.30E+00	1.84E+03	1.84E+03	4.15E+01	2.33E+03	2.33E+03	7.49E+00
14133-76-7	Tc-99	pCi/g	3.32E+04	3.32E+04	3.32E+04	1.66E+05	1.66E+05	1.66E+05	1.95E+05	1.95E+05	1.95E+05
14269-63-7	Th-230	pCi/g	8.74E+00	2.29E+01	8.04E+00	4.37E+01	1.13E+02	4.02E+01	1.05E+01	2.79E+01	9.69E+00
7440-29-1	Th-232 ^b	pCi/g	8.57E+00	8.57E+00	8.57E+00	4.29E+01	4.29E+01	4.29E+01	7.52E+00	7.52E+00	7.52E+00
13966-29-5	U-234	pCi/g	1.35E+01	7.30E+02	7.97E+00	6.73E+01	3.64E+03	3.98E+01	1.63E+01	2.76E+03	9.67E+00
15117-96-1	U-235	pCi/g	3.12E+01	1.53E+02	3.11E+01	1.56E+02	7.66E+02	1.56E+02	3.49E+01	1.39E+02	3.49E+01
7440-61-1	U-238	pCi/g	7.83E+00	4.40E+02	7.83E+00	3.92E+01	2.20E+03	3.91E+01	9.53E+00	6.42E+02	9.53E+00

Table A.1b. Soil/Sediment Action Levels for Significant Radionuclide COPCs at PGDP (Continued)

(Values were calculated in October 2022 and are based on best available information.)

			R	ecreational Us	er ^b		Resident ^c	
CAS Number	Analyte	Units	Peak Risk Infinite ^a	Peak Risk 1,000 Years	Secular Equilibrium	Peak Risk Infinite ^a	Peak Risk 1,000 Years	Secular Equilibrium
14596-10-2	Am-241	pCi/g	8.82E+02	8.82E+02	2.52E+01	2.10E+02	2.10E+02	4.55E+00
10045-97-3	Cs-137	pCi/g	3.13E+01	3.13E+01	2.36E+01	5.24E+00	5.24E+00	3.95E+00
13994-20-2	Np-237	pCi/g	3.18E+01	6.76E+01	2.60E+01	5.69E+00	1.15E+01	4.66E+00
13981-16-3	Pu-238	pCi/g	1.30E+03	1.30E+03	6.13E+00	4.28E+02	4.28E+02	1.10E+00
15117-48-3	Pu-239	pCi/g	1.16E+03	1.16E+03	2.21E+01	3.79E+02	3.79E+02	3.97E+00
14119-33-6	Pu-240	pCi/g	1.16E+03	1.16E+03	4.95E+00	3.81E+02	3.81E+02	8.54E-01
14133-76-7	Tc-99	pCi/g	3.49E+05	3.49E+05	3.49E+05	1.12E+04	1.12E+04	1.12E+04
14269-63-7	Th-230	pCi/g	6.73E+00	1.78E+01	6.19E+00	1.21E+00	3.20E+00	1.11E+00
7440-29-1	Th-232 ^b	pCi/g	4.99E+00	4.99E+00	4.99E+00	8.57E-01	8.57E-01	8.57E-01
13966-29-5	U-234	pCi/g	1.04E+01	1.19E+03	6.17E+00	1.87E+00	3.09E+02	1.11E+00
15117-96-1	U-235	pCi/g	2.25E+01	9.20E+01	2.25E+01	4.01E+00	1.59E+01	4.01E+00
7440-61-1	U-238	pCi/g	6.07E+00	3.67E+02	6.07E+00	1.09E+00	7.09E+02	1.09E+00

NOTES: Values are provided in these tables for significant radionuclide COPCs for PGDP. Values for other radionuclides can be obtained using the EPA Radionuclide PRG calculator, as modified using PGDP-specific inputs. The resident ALs do not include the consumption of produce pathway that is included in the EPA Radionuclide PRG calculator.

Radionuclide ALs are based on the cancer endpoint using a target ELCR of 1.0E-4. The soil and sediment radionuclide ALs consider ingestion, inhalation, and external exposure pathways.

^a The time frame for calculating peak risk over an infinite period in the EPA Radionuclide PRG calculator is 1 trillion years. The National Aeronautics and Space Administration (NASA) most recently has estimated the age of the universe as being 13.7 billion years, with an uncertainty of only 200 million years (https://imagine.gsfc.nasa.gov).

^b For the recreational user, the radionuclide ALs were calculated using the child/adult age-adjusted scenario (i.e., 26-year exposure duration, with 6 years as a child and 20 years as an adult). The EPA Radionuclide PRG calculator only allows exposure parameter entries for two life stages (i.e., child and adult), rather than the four life stage groups provided in the EPA and RAIS chemical online calculators; therefore, the EPA Radionuclide PRG calculator does not allow entry of the teen exposure parameters shown in Table B.5.

^c For the resident, the radionuclide ALs were calculated using the child/adult age-adjusted lifetime scenario (i.e., lifetime exposure).

^d Analyte is not a PGDP-significant COPC (Table 2.1), but it is provided for project support.

Table A.2a. Groundwater Action Levels and Primary MCLs for Significant COPCs at PGDP (Values were calculated in October 2022 and are based on best available information.)

CAS			Resident	Adult F	Resident	Child R	Resident	Primary
Number	Analyte	Units	Cancer ^a	Hazard	Action	Hazard	Action	MCLs ^j
7429-90-5	Aluminum	μg/L	-	9.96E+04	9.96E+04	6.00E+04	6.00E+04	-
7440-36-0	Antimony (metallic)	μg/L	-	3.87E+01	3.87E+01	2.34E+01	2.34E+01	6.00E+00
7440-38-2	Arsenic, Inorganic	μg/L	5.17E+00	2.99E+01	5.17E+00	1.80E+01	5.17E+00	1.00E+01
7440-39-3	Barium	μg/L	-	1.85E+04	1.85E+04	1.13E+04	1.13E+04	2.00E+03
7440-41-7	Beryllium and compounds	μg/L	-	1.11E+02	1.11E+02	7.38E+01	7.38E+01	4.00E+00
7440-42-8	Boron And Borates Only	μg/L	-	1.99E+04	1.99E+04	1.20E+04	1.20E+04	-
7440-43-9	Cadmium (Water)	μg/L	-	9.00E+00	9.00E+00	5.52E+00	5.52E+00	5.00E+00
16065-83-1	Chromium(III), Insoluble Salts	μg/L	-	1.05E+05	1.05E+05	6.75E+04	6.75E+04	-
18540-29-9	Chromium(VI)	μg/L	3.50E+00	2.08E+02	3.50E+00	1.34E+02	3.50E+00	-
7440-47-3	Chromium (Total) ^b	μg/L	-	-	-	-	-	1.00E+02
7440-48-4	Cobalt	μg/L	-	3.00E+01	3.00E+01	1.80E+01	1.80E+01	-
7440-50-8	Copper	μg/L	-	3.99E+03	3.99E+03	2.40E+03	2.40E+03	1.30E+03
16984-48-8	Fluoride	μg/L	-	3.99E+03	3.99E+03	2.40E+03	2.40E+03	4.00E+03
7439-89-6	Iron	μg/L	-	6.96E+04	6.96E+04	4.20E+04	4.20E+04	-
7439-92-1	Lead ^c	μg/L	-	-	3.00E+01	-	3.00E+01	1.50E+01
7439-96-5	Manganese (Non-Diet)	μg/L	-	2.11E+03	2.11E+03	1.30E+03	1.30E+03	-
Various	Mercury, Inorganic Salts	μg/L	-	2.78E+01	2.78E+01	1.70E+01	1.70E+01	2.00E+00 ^d
7439-98-7	Molybdenum	μg/L	-	4.98E+02	4.98E+02	2.99E+02	2.99E+02	-
7440-02-0	Nickel Soluble Salts	μg/L	-	1.95E+03	1.95E+03	1.18E+03	1.18E+03	-
7782-49-2	Selenium	μg/L	-	4.98E+02	4.98E+02	2.99E+02	2.99E+02	5.00E+01
7440-22-4	Silver	μg/L	-	4.62E+02	4.62E+02	2.82E+02	2.82E+02	-
7440-28-0	Thallium (Soluble Salts)	μg/L	-	9.96E-01	9.96E-01	6.00E-01	6.00E-01	2.00E+00
N/A	Uranium (Insoluble Compounds) ^e	μg/L	-	2.99E+02	2.99E+02	1.80E+02	1.80E+02	3.00E+01
7440-61-1	Uranium (Soluble Salts) ^e	μg/L	-	1.99E+01	1.99E+01	1.20E+01	1.20E+01	3.00E+01
7440-62-2	Vanadium and Compounds	μg/L	-	4.14E+02	4.14E+02	2.59E+02	2.59E+02	-
7440-66-6	Zinc and Compounds	μg/L	-	2.99E+04	2.99E+04	1.80E+04	1.80E+04	-
83-32-9	Acenaphthene	μg/L	-	2.52E+03	2.52E+03	1.61E+03	1.61E+03	-
208-96-8	Acenaphthylene ^f	μg/L	-	2.52E+03	2.52E+03	1.61E+03	1.61E+03	-
107-13-1	Acrylonitrile	μg/L	5.23E+00	1.24E+01	5.23E+00	1.23E+01	5.23E+00	-
120-12-7	Anthracene	μg/L	-	8.22E+03	8.22E+03	5.31E+03	5.31E+03	-
71-43-2	Benzene	μg/L	4.55E+01	1.22E+02	4.55E+01	9.96E+01	4.55E+01	5.00E+00
117-81-7	Bis(2-ethylhexyl)phthalateg	μg/L	5.56E+02	2.00E+03	5.56E+02	1.20E+03	5.56E+02	6.00E+00
75-27-4	Bromodichloromethane	μg/L	1.34E+01	7.50E+02	1.34E+01	4.53E+02	1.34E+01	8.00E+01k
86-74-8	Carbazole	μg/L	2.03E+02	-	2.03E+02	-	2.03E+02	-
56-23-5	Carbon Tetrachloride	μg/L	4.55E+01	2.11E+02	4.55E+01	1.49E+02	4.55E+01	5.00E+00
67-66-3	Chloroform	μg/L	2.21E+01	3.66E+02	2.21E+01	2.92E+02	2.21E+01	8.00E+01k
75-71-8	Dichlorodifluoromethane (Freon-12) ^g	μg/L	-	6.06E+02	6.06E+02	5.91E+02	5.91E+02	-

Table A.2a. Groundwater Action Levels and Primary MCLs for Significant COPCs at PGDP (Continued) (Values were calculated in October 2022 and are based on best available information.)

CAS			Resident		Resident		Resident	Primary
Number	Analyte	Units	Cancer ^a	Hazard	Action	Hazard	Action	MCLs ^j
75-34-3	Dichloroethane, 1,1-g	μg/L	2.75E+02		2.75E+02	2.45E+03	2.75E+02	-
107-06-2	Dichloroethane, 1,2-	μg/L	1.71E+01		1.71E+01	3.90E+01	1.71E+01	5.00E+00
75-35-4	Dichloroethylene, 1,1-	μg/L	-		9.75E+02	8.55E+02	8.55E+02	7.00E+00
540-59-0	Dichloroethylene, 1,2- (Mixed Isomers)	μg/L	-	8.04E+02	8.04E+02	4.89E+02	4.89E+02	-
156-59-2	Dichloroethylene, cis- 1,2-	μg/L	-	1.04E+02	1.04E+02	7.56E+01	7.56E+01	7.00E+01
156-60-5	Dichloroethylene, trans-1,2-	μg/L	-	2.20E+02		2.03E+02	2.03E+02	1.00E+02
60-57-1	Dieldrin	μg/L	1.75E-01	1.79E+00		1.14E+00	1.75E-01	-
1746-01-6	Dioxins/Furans, Total (as TCDD)h	μg/L	1.19E-05	5.49E-05	1.19E-05	3.60E-05	1.19E-05	3.00E-05
37871-00-4	~HpCDD, 2,3,7,8-	μg/L	1.19E-03	5.49E-03	1.19E-03	3.60E-03	1.19E-03	-
38998-75-3	~HpCDF, 2,3,7,8-	μg/L	1.19E-03	5.49E-03	1.19E-03	3.60E-03	1.19E-03	-
34465-46-8	~HxCDD	μg/L	5.99E-04	7.02E-04	5.99E-04	4.20E-04	4.20E-04	-
55684-94-1	~HxCDF, 2,3,7,8-	μg/L	5.99E-04	7.02E-04	5.99E-04	4.20E-04	4.20E-04	-
3268-87-9	~OCDD	μg/L	2.00E-01	2.34E-01	2.00E-01	1.40E-01	1.40E-01	-
39001-02-0	~OCDF	μg/L	2.00E-01	2.34E-01	2.00E-01	1.40E-01	1.40E-01	-
36088-22-9	~PeCDD, 2,3,7,8-	μg/L	5.99E-05	7.02E-05	5.99E-05	4.20E-05	4.20E-05	-
57117-41-6	~PeCDF, 1,2,3,7,8-	μg/L	2.00E-03	2.34E-03	2.00E-03	1.40E-03	1.40E-03	-
57117-31-4	~PeCDF, 2,3,4,7,8-	μg/L	2.00E-04	2.34E-04	2.00E-04	1.40E-04	1.40E-04	-
1746-01-6	~TCDD, 2,3,7,8-	μg/L	1.19E-05	5.49E-05	1.19E-05	3.60E-05	1.19E-05	3.00E-05
51207-31-9	~TCDF, 2,3,7,8-	μg/L	1.19E-04	5.49E-04	1.19E-04	3.60E-04	1.19E-04	-
100-41-4	Ethylbenzene	μg/L	1.50E+02	2.10E+03	1.50E+02	1.50E+03	1.50E+02	7.00E+02
206-44-0	Fluoranthene	μg/L	-	3.99E+03	3.99E+03	2.41E+03	2.41E+03	-
86-73-7	Fluorene	μg/L	-	1.38E+03	1.38E+03	8.82E+02	8.82E+02	-
118-74-1	Hexachlorobenzene	μg/L	9.76E-01	1.00E+00	9.76E-01	6.02E-01	6.02E-01	1.00E+00
91-20-3	Naphthalene	μg/L	1.17E+01	1.85E+01	1.17E+01	1.83E+01	1.17E+01	-
88-74-4	Nitroaniline, 2-	μg/L	-	9.39E+02	9.39E+02	5.67E+02	5.67E+02	-
621-64-7	Nitroso-di-N-propylamine, N-	μg/L	1.08E+00	-	1.08E+00	-	1.08E+00	-
87-86-5	Pentachlorophenol	μg/L	4.13E+00	1.05E+02	4.13E+00	6.81E+01	4.13E+00	1.00E+00
85-01-8	Phenanthrene ^f	μg/L	-	2.52E+03	2.52E+03	1.61E+03	1.61E+03	-
1336-36-3	Polychlorinated Biphenyls, Total	μg/L	4.36E+00	-	4.36E+00	-	4.36E+00	5.00E-01
12674-11-2	~Aroclor 1016	μg/L	2.24E+01	7.02E+00	7.02E+00	4.20E+00	4.20E+00	-
11104-28-2	~Aroclor 1221	μg/L	4.71E-01	-	4.71E-01	-	4.71E-01	-
11141-16-5	~Aroclor 1232	μg/L	4.71E-01	-	4.71E-01	-	4.71E-01	-
53469-21-9	~Aroclor 1242	μg/L	7.85E-01	-	7.85E-01	-	7.85E-01	-
12672-29-6	~Aroclor 1248	μg/L	7.85E-01	-	7.85E-01	-	7.85E-01	-
11097-69-1	~Aroclor 1254	μg/L	7.85E-01	2.00E+00	7.85E-01	1.20E+00	7.85E-01	-
11096-82-5	~Aroclor 1260	μg/L	7.85E-01	-	7.85E-01	-	7.85E-01	-

Table A.2a. Groundwater Action Levels and Primary MCLs for Significant COPCs at PGDP (Continued)

(Values were calculated in October 2022 and are based on best available information.)

CAS			Resident	Adult F	Resident	Child R	lesident	Primary
Number	Analyte	Units	Cancera	Hazard	Action	Hazard	Action	MCLs ^j
50-32-8	Polycyclic aromatic hydrocarbons, Total Carcinogenic ⁱ	μg/L	2.51E+00	3.00E+01	2.51E+00	1.81E+01	2.51E+00	2.00E-01
56-55-3	~Benz[a]anthracene	μg/L	2.98E+00	-	2.98E+00	-	2.98E+00	-
50-32-8	~Benzo[a]pyrene	μg/L	2.51E+00	3.00E+01	2.51E+00	1.81E+01	2.51E+00	2.00E-01
205-99-2	~Benzo[b]fluoranthene	μg/L	2.51E+01	-	2.51E+01	-	2.51E+01	-
207-08-9	~Benzo[k]fluoranthene	μg/L	2.51E+02	-	2.51E+02	-	2.51E+02	-
218-01-9	~Chrysene	μg/L	2.51E+03	-	2.51E+03	-	2.51E+03	-
53-70-3	~Dibenz[a,h]anthracene	μg/L	2.51E+00	-	2.51E+00	-	2.51E+00	-
193-39-5	~Indeno[1,2,3-cd]pyrene	μg/L	2.51E+01	-	2.51E+01	-	2.51E+01	-
129-00-0	Pyrene	μg/L	-	5.58E+02	5.58E+02	3.63E+02	3.63E+02	-
127-18-4	Tetrachloroethylene	μg/L	1.13E+03	1.51E+02	1.51E+02	1.22E+02	1.22E+02	5.00E+00
108-88-3	Tolueneg	μg/L	-	5.04E+03	5.04E+03	3.30E+03	3.30E+03	1.00E+03
76-13-1	Trichloro-1,2,2-trifluoroethane, 1,1,2- (Freon-113) ^g	μg/L	-	3.09E+04	3.09E+04	3.06E+04	3.06E+04	-
71-55-6	Trichloroethane, 1,1,1-	μg/L	-	2.64E+04	2.64E+04	2.40E+04	2.40E+04	2.00E+02
79-00-5	Trichloroethane, 1,1,2-	μg/L	2.75E+01	1.25E+00	1.25E+00	1.25E+00	1.25E+00	5.00E+00
79-01-6	Trichloroethylene	μg/L	4.94E+01	9.69E+00	9.69E+00	8.49E+00	8.49E+00	5.00E+00
75-01-4	Vinyl Chloride	μg/L	1.88E+00	1.79E+02	1.88E+00	1.26E+02	1.88E+00	2.00E+00
108-38-3	Xylene, m-	μg/L	-	5.94E+02	5.94E+02	5.79E+02	5.79E+02	-
95-47-6	Xylene, o-	μg/L	-	5.97E+02	5.97E+02	5.79E+02	5.79E+02	-
106-42-3	Xylene, p-	μg/L	-	5.97E+02	5.97E+02	5.79E+02	5.79E+02	-
1330-20-7	Xylene, Mixture	μg/L	-	5.97E+02	5.97E+02	5.79E+02	5.79E+02	1.00E+04

NOTES: An HI of 3 is used for the AL derivation because the range of values for HI (based on RGO tables) are 0.1, 1, and 3. Values are provided in these tables for significant COPCs for PGDP. Values for other COPCs can be obtained using the RAIS Chemical PRG online calculator, as modified using PGDP-specific inputs. ALs are not adjusted for solubility limits.

^a For the resident, ELCRs (i.e., cancer ALs) were calculated using the child/adult age-adjusted lifetime scenario (i.e., lifetime exposure).

^b Chromium (Total) AL should utilize Chromium III or Chromium VI, as appropriate.

^c Lead values should be checked prior to use to ensure they are still current.

d MCL is for mercury (elemental).

^e Based on recommendation from EPA, ALs for uranium (soluble salts) now use the RfD and the RfC for soluble compounds of uranium derived from ATSDR (EPA 2016). ALs for uranium (insoluble compounds) use the RfD for uranium (soluble salts), which is available in IRIS; the RfC for insoluble compounds of uranium are derived from ATSDR.

f Acenaphthylene and phenanthrene use values for acenaphthene as a surrogate.

g Analytes are not PGDP-significant COPCs (Table 2.1) but are provided for project support.

h Total dioxins/furans uses values for 2,3,7,8-TCDD, see screening note 9f in the Appendix A introduction, "Screening Levels," on pages A-3-A-5.

¹ Total carcinogenic PAHs uses values for BaP, see screening note 9d in the Appendix A introduction, "Screening Levels," on pages A-3-A-5.

^j Accessed at https://www.epa.gov/sites/default/files/2016-06/documents/npwdr complete table.pdf on October 27, 2022.

^kMCL is for the sum of the concentrations for trihalomethanes.

Table A.2b. Groundwater Action Levels and Primary MCLs for Significant Radionuclide COPCs at PGDP

(Values were calculated in October 2022 and are based on best available information.)

CAS				Residenta		D:
Number Number	Analyte	Units	Peak Risk Infinite ^b	Peak Risk 1,000 Years	Secular Equilibrium	Primary MCLs
14596-10-2	Am-241	pCi/L	5.15E+01	5.15E+01	6.77E+00	1.50E+01°
10045-97-3	Cs-137	pCi/L	2.28E+02	2.28E+02	1.71E+02	d
13994-20-2	Np-237	pCi/L	9.64E+00	7.61E+01	7.83E+00	1.50E+01e
13981-16-3	Pu-238	pCi/L	4.40E+01	4.40E+01	1.56E+00	1.50E+01 ^f
15117-48-3	Pu-239	pCi/L	3.87E+01	3.87E+01	6.03E+00	1.50E+01 ^f
14119-33-6	Pu-240	pCi/L	3.87E+01	3.87E+01	3.18E+00	1.50E+01 ^f
14133-76-7	Tc-99	pCi/L	1.90E+03	1.90E+03	1.90E+03	g
14269-63-7	Th-230	pCi/L	1.81E+00	4.65E+00	1.66E+00	1.50E+01 ^h
7440-29-1	Th-232i	pCi/L	3.63E+00	3.63E+00	3.63E+00	1.50E+01 ^h
13966-29-5	U-234	pCi/L	2.74E+00	6.86E+01	1.62E+00	j
15117-96-1	U-235	pCi/L	7.15E+00	6.14E+01	7.14E+00	j
7440-61-1	U-238	pCi/L	1.58E+00	5.99E+01	1.58E+00	j

NOTES: Values are provided in these tables for significant radionuclide COPCs for PGDP. Values for other radionuclides can be obtained using the EPA Radionuclide PRG calculator, as modified using PGDP-specific inputs.

Radionuclide ALs are based on the cancer endpoint using a target ELCR of 1.0E-4. The groundwater radionuclide ALs are based on the ingestion pathway only.

- U-234/U-235 ranges 21-22 obtained from conversion approximately 21.9.
- U-235/U-238 ranges 0.04-0.05 obtained from conversion approximately 0.045.

^a For the resident, cancer ALs were calculated using the child/adult age-adjusted lifetime scenario (i.e., lifetime exposure).

^b The time frame for calculating peak risk over an infinite period in the EPA Radionuclide PRG calculator is 1 trillion years. NASA most recently has estimated the age of the universe as being 13.7 billion years, with an uncertainty of only 200 million years (https://imagine.gsfc.nasa.gov).

^c Additional information regarding Am-241 can be found in "EPA Facts about Americium-241," dated July 2002, at the following link: https://semspub.epa.gov/work/HQ/176297.pdf; accessed October 21, 2022.

^d The EPA MCL for Cs-137 is 4 mrem/year. The value derived by the EPA from the 4 mrem/year MCL for Cs-137 is 200 pCi/L (see "Limits for Beta Particles and Photon Emitters at 4 millrems/year" found on https://www.epa.gov/sites/production/files/2015-09/documents/guide radionuclides table-betaphotonemitters.pdf, accessed October 5, 2022).

^e "Maximum Contaminant Level's in EPA's PRG and Dose Compliance Concentration Calculators," revised September 2015, found on https://epa-prgs.ornl.gov/radionuclides/MCLs 2015.pdf; accessed October 5, 2022).

Additional information regarding plutonium can be found at the following link: http://www.epa.gov/radiation/radionuclides.

E The value derived by the EPA from the 4 mrem/year MCL for Tc-99 is 900 pCi/L, (see https://www.epa.gov/sites/default/files/2015-09/documents/guide radionuclides table-betaphotonemitters.pdf; accessed October 5, 2022). An alternate value derived by EPA from the 4 mrem/year MCL is 3,790 pCi/L and was proposed in the July 18, 1991, Federal Register. See Table A.9 for Tc-99 dose-based groundwater screening levels resulting in a 4 mrem/year dose based upon more recent dosimetry.

h Additional information regarding thorium can be found at the following link: http://www.epa.gov/radiation/radionuclides.

ⁱ Analyte is not a PGDP-significant COPC (Table 2.1), but it is provided for project support.

^j The uranium MCL is 30 μg/L and can be assumed to be at a 1:1 ratio for pCi/L (or 30 pCi/L). The MCL also can be converted to 20 pCi/L for total uranium using a uranium activity expected at PGDP. Isotopic uranium values derived from this conversion are 10.24 pCi/L for U-234, 0.466 pCi/L for U-235, and 9.99 pCi/L for U-238, assuming natural occurring uranium at 0.725% U-235 and the following ratios:

CAS			Ou	tdoor Work	er ^a	Exca	avation Wo	kera	Indu	ıstrial Worl	ker ^a
Number	Analyte	Units	Cancer	Hazard	Action	Cancer	Hazard	Action	Cancer	Hazard	Action
7429-90-5	Aluminum	μg/L	-	1.55E+08	1.55E+08	-	1.55E+08	1.55E+08	-	3.81E+07	3.81E+07
7440-36-0	Antimony (metallic)	μg/L	-	9.30E+03	9.30E+03	-	9.30E+03	9.30E+03	-	2.29E+03	2.29E+03
7440-38-2	Arsenic, Inorganic	μg/L	9.66E+03	4.65E+04	9.66E+03	4.83E+04	4.65E+04	4.65E+04	2.38E+03	1.15E+04	2.38E+03
7440-39-3	Barium	μg/L	-	2.17E+06	2.17E+06	-	2.17E+06	2.17E+06	-	5.34E+05	5.34E+05
7440-41-7	Beryllium and compounds	μg/L	-	2.17E+03	2.17E+03	-	2.17E+03	2.17E+03	-	5.34E+02	5.34E+02
7440-42-8	Boron And Borates Only	μg/L	-	3.09E+07	3.09E+07	-	3.09E+07	3.09E+07	-	7.65E+06	7.65E+06
7440-43-9	Cadmium (Water)	μg/L	-	7.76E+02	7.76E+02	-	7.76E+02	7.76E+02	-	1.91E+02	1.91E+02
16065-83-1	Chromium(III), Insoluble Salts	μg/L	-	3.03E+06	3.03E+06	-	3.03E+06	3.03E+06	-	7.44E+05	7.44E+05
18540-29-9	Chromium(VI)	μg/L	3.62E+02	5.82E+03	3.62E+02	1.81E+03	5.82E+03	1.81E+03	8.92E+01	1.43E+03	8.92E+01
7440-47-3	Chromium (Total) ^b	-	-	-	-	-	-	-	-	-	-
7440-48-4	Cobalt	μg/L	-	1.16E+05	1.16E+05	-	1.16E+05	1.16E+05	-	2.87E+04	2.87E+04
7440-50-8	Copper	μg/L	-	6.21E+06	6.21E+06	-	6.21E+06	6.21E+06	-	1.53E+06	1.53E+06
16984-48-8	Fluoride	μg/L	-	6.21E+06	6.21E+06	-	6.21E+06	6.21E+06	-	1.53E+06	1.53E+06
7439-89-6	Iron	μg/L	-	1.09E+08	1.09E+08	-	1.09E+08	1.09E+08	-	2.68E+07	2.68E+07
7439-92-1	Lead ^c	μg/L	-	-	3.00E+01	-	-	3.00E+01	-	-	3.00E+01
7439-96-5	Manganese (Non-Diet)	μg/L	-	1.49E+05	1.49E+05	-	1.49E+05	1.49E+05	-	3.66E+04	3.66E+04
Various	Mercury, Inorganic Salts	μg/L	-	3.27E+03	3.27E+03	-	3.27E+03	3.27E+03	-	8.01E+02	8.01E+02
7439-98-7	Molybdenum	μg/L	-	7.77E+05	7.77E+05	-	7.77E+05	7.77E+05	-	1.91E+05	1.91E+05
7440-02-0	Nickel Soluble Salts	μg/L	-	6.21E+05	6.21E+05	-	6.21E+05	6.21E+05	-	1.53E+05	1.53E+05
7782-49-2	Selenium	μg/L	-	7.77E+05	7.77E+05	-	7.77E+05	7.77E+05	-	1.91E+05	1.91E+05
7440-22-4	Silver	μg/L	-	5.16E+04	5.16E+04	-	5.16E+04	5.16E+04	-	1.28E+04	1.28E+04
7440-28-0	Thallium (Soluble Salts)	μg/L	-	1.55E+03	1.55E+03	-	1.55E+03	1.55E+03	-	3.81E+02	3.81E+02
N/A	Uranium (Insoluble Compounds) ^d	μg/L	-	4.65E+05	4.65E+05	-	4.65E+05	4.65E+05	-	1.15E+05	1.15E+05
7440-61-1	Uranium (Soluble Salts) ^d	μg/L	-	3.09E+04	3.09E+04	-	3.09E+04	3.09E+04	-	7.65E+03	7.65E+03
7440-62-2	Vanadium and Compounds	μg/L	-	2.03E+04	2.03E+04	-	2.03E+04	2.03E+04	-	5.01E+03	5.01E+03
7440-66-6	Zinc and Compounds	μg/L	-	7.77E+07	7.77E+07	-	7.77E+07	7.77E+07	-	1.91E+07	1.91E+07
83-32-9	Acenaphthene	μg/L	-	1.11E+05	1.11E+05	-	1.11E+05	1.11E+05	-	1.75E+04	1.75E+04
208-96-8	Acenaphthylenee	μg/L	-	1.11E+05	1.11E+05	-	1.11E+05	1.11E+05	-	1.75E+04	1.75E+04
107-13-1	Acrylonitrile	μg/L	2.20E+04	1.28E+06	2.20E+04	1.10E+05	1.28E+06	1.10E+05	4.92E+03	2.85E+05	4.92E+03
120-12-7	Anthracene	μg/L	-	3.30E+05	3.30E+05	-	3.30E+05	3.30E+05	-	4.59E+04	4.59E+04
71-43-2	Benzene	μg/L	1.72E+04	4.05E+04	1.72E+04	8.60E+04	4.05E+04	4.05E+04	3.67E+03	8.67E+03	3.67E+03
117-81-7	Bis(2-ethylhexyl)phthalatef	μg/L	-	-	-	-	-	-	-	-	-
75-27-4	Bromodichloromethane	μg/L	4.83E+04	2.57E+05	4.83E+04	2.42E+05	2.57E+05	2.42E+05	8.61E+03	4.57E+04	8.61E+03
86-74-8	Carbazole	μg/L	1.26E+04	-	1.26E+04	6.29E+04	-	6.29E+04	2.00E+03	-	2.00E+03
56-23-5	Carbon Tetrachloride	μg/L	1.12E+04	3.36E+04	1.12E+04	5.60E+04	3.36E+04	3.36E+04	2.00E+03	6.00E+03	2.00E+03
67-66-3	Chloroform	μg/L	6.23E+04	2.07E+05	6.23E+04	3.12E+05	2.07E+05	2.07E+05	1.24E+04	4.11E+04	1.24E+04
75-71-8	Dichlorodifluoromethane (Freon-12) ^f	μg/L	-	3.18E+06	3.18E+06	-	3.18E+06		-	6.27E+05	6.27E+05

CAS			Ou	tdoor Work	er ^a	Exca	avation Wor	ker ^a	Ind	ustrial Wor	ker ^a
Number	Analyte	Units	Cancer	Hazard	Action	Cancer	Hazard	Action	Cancer	Hazard	Action
75-34-3	Dichloroethane, 1,1-f	μg/L	3.51E+05	4.29E+06	3.51E+05	1.76E+06	4.29E+06	1.76E+06	7.29E+04	8.91E+05	7.29E+04
107-06-2	Dichloroethane, 1,2-	μg/L	3.51E+04	2.06E+05	3.51E+04	1.76E+05	2.06E+05	1.76E+05	7.30E+03	4.26E+04	7.30E+03
75-35-4	Dichloroethylene, 1,1-	μg/L	-	6.30E+05	6.30E+05	-	6.30E+05	6.30E+05	-	1.31E+05	1.31E+05
540-59-0	Dichloroethylene, 1,2- (Mixed Isomers)	μg/L	-	1.20E+05	1.20E+05	-	1.20E+05	1.20E+05	-	2.49E+04	2.49E+04
156-59-2	Dichloroethylene, cis-1,2-	μg/L	-	2.67E+04	2.67E+04	-	2.67E+04	2.67E+04	-	5.55E+03	5.55E+03
156-60-5	Dichloroethylene, trans-1,2-	μg/L	-	2.67E+05	2.67E+05	-	2.67E+05	2.67E+05	-	5.55E+04	5.55E+04
60-57-1	Dieldrin	μg/L	9.40E+00	8.07E+01	9.40E+00	4.70E+01	8.07E+01	4.70E+01	1.32E+00	1.13E+01	1.32E+00
1746-01-6	Dioxins/Furans, Total (as TCDD) ^g	μg/L	-	-	-	-	-	-	-	-	-
37871-00-4	~HpCDD	μg/L	-	-	-	-	-	-	-	-	-
38998-75-3	~HpCDF, 2,3,7,8-	μg/L	-	-	-	-	-	-	-	-	-
34465-46-8	~HxCDD, 2,3,7,8-	μg/L	-	-	-	-	-	-	-	-	-
55684-94-1	~HxCDF, 2,3,7,8-	μg/L	-	-	-	-	-	-	-	-	-
3268-87-9	~OCDD	μg/L	-	-	-	-	-	-	-	-	-
39001-02-0	~OCDF	μg/L	-	-	-	-	-	-	-	-	-
36088-22-9	~PeCDD, 2,3,7,8-	μg/L	-	-	-	-	-	-	-	-	-
57117-41-6	~PeCDF, 1,2,3,7,8-	μg/L	-	-	-	-	-	-	-	-	-
57117-31-4	~PeCDF, 2,3,4,7,8-	μg/L	-	-	-	-	-	-	-	-	-
1746-01-6	~TCDD, 2,3,7,8-	μg/L	-	-	-	-	-	-	-	-	-
51207-31-9	~TCDF, 2,3,7,8-	μg/L	-	-	-	-	-	-	-	-	-
100-41-4	Ethylbenzene	μg/L	2.78E+04	1.64E+05	2.78E+04	1.39E+05	1.64E+05	1.39E+05	5.41E+03	3.19E+04	5.41E+03
206-44-0	Fluoranthene	μg/L	-	-	-	-	-	-	-	-	-
86-73-7	Fluorene	μg/L	-	5.76E+04	5.76E+04	-	5.76E+04	5.76E+04	-	8.34E+03	8.34E+03
118-74-1	Hexachlorobenzene	μg/L	-	-	-	-	-	-	-	-	-
91-20-3	Naphthalene	μg/L	2.60E+03	6.69E+04	2.60E+03	9.56E+03	6.69E+04	9.56E+03	4.77E+02	1.23E+04	4.77E+02
88-74-4	Nitroaniline, 2-	μg/L	-	3.06E+05	3.06E+05	-	3.06E+05	3.06E+05	-	5.82E+04	5.82E+04
621-64-7	Nitroso-di-N-propylamine, N-	μg/L	7.85E+02	-	7.85E+02	3.92E+03	-	3.92E+03	1.53E+02	-	1.53E+02
87-86-5	Pentachlorophenol	μg/L	1.80E+02	3.84E+03	1.80E+02	8.98E+02	3.84E+03	8.98E+02	2.52E+01	5.40E+02	2.52E+01
85-01-8	Phenanthrene ^e	μg/L	-	1.11E+05	1.11E+05	-	1.11E+05	1.11E+05	-	1.75E+04	1.75E+04
1336-36-3	Polychlorinated Biphenyls, Total	μg/L	-	-	-	-	-	-	-	-	-
12674-11-2	~Aroclor 1016	μg/L	-	-	-	-	-	-	-	-	-
11104-28-2	~Aroclor 1221	μg/L	4.16E+01	-	4.16E+01	2.08E+02	-	2.08E+02	5.66E+00	-	5.66E+00
11141-16-5	~Aroclor 1232	μg/L	4.16E+01	-	4.16E+01	2.08E+02	-	2.08E+02	5.66E+00	-	5.66E+00
53469-21-9	~Aroclor 1242	μg/L	-	-	-	-	-	-	-	-	-
12672-29-6	~Aroclor 1248	μg/L	-	-	-	-	-	-	-	-	-
11097-69-1	~Aroclor 1254	μg/L	-	-	-	-	-	-	-	-	-
11096-82-5	~Aroclor 1260	μg/L	-	-	-	-	-	-	-	-	-

CAS			Ou	tdoor Work	era	Exca	vation Wor	ker ^a	Industrial Workera		
Number	Analyte	Units	Cancer	Hazard	Action	Cancer	Hazard	Action	Cancer	Hazard	Action
50-32-8	Polycyclic aromatic hydrocarbons, Total Carcinogenich	μg/L	-	-	-	-	-	-	-	-	-
56-55-3	~Benz[a]anthracene	μg/L	-	-	-	-	-	-	-	-	-
50-32-8	~Benzo[a]pyrene	μg/L	-	-	-	-	-	-	-	-	-
205-99-2	~Benzo[b]fluoranthene	μg/L	1	1	1	-	1	-	-	-	-
207-08-9	~Benzo[k]fluoranthene	μg/L	1	1	1	-	1	-	-	-	-
218-01-9	~Chrysene	μg/L	1	1	1	-	1	-	-	-	-
53-70-3	~Dibenz[a,h]anthracene	μg/L	1	1	1	-	1	-	-	-	-
193-39-5	~Indeno[1,2,3-cd]pyrene	μg/L	1	1	1	-	1	-	-	-	-
129-00-0	Pyrene	μg/L	1	2.07E+04	2.07E+04	-	2.07E+04	2.07E+04	-	2.78E+03	2.78E+03
127-18-4	Tetrachloroethylene	μg/L	1.85E+05	2.50E+04	2.50E+04	9.24E+05	2.50E+04	2.50E+04	3.07E+04	4.14E+03	4.14E+03
108-88-3	Toluenef	μg/L	-	4.02E+05	4.02E+05	-	4.02E+05	4.02E+05	-	8.25E+04	8.25E+04
76-13-1	Trichloro-1,2,2-trifluoroethane, 1,1,2- (Freon-113) ^f	μg/L	ı	2.15E+08	2.15E+08	-	2.15E+08	2.15E+08	-	3.51E+07	3.51E+07
71-55-6	Trichloroethane, 1,1,1-	μg/L	ı	2.24E+07	2.24E+07	-	2.24E+07	2.24E+07	-	4.26E+06	4.26E+06
79-00-5	Trichloroethane, 1,1,2-	μg/L	4.47E+04	1.09E+05	4.47E+04	2.24E+05	1.09E+05	1.09E+05	8.62E+03	2.11E+04	8.62E+03
79-01-6	Trichloroethylene	μg/L	2.45E+04	6.06E+03	6.06E+03	1.22E+05	6.06E+03	6.06E+03	4.70E+03	1.16E+03	1.16E+03
75-01-4	Vinyl Chloride	μg/L	2.32E+03	5.37E+04	2.32E+03	1.16E+04	5.37E+04	1.16E+04	5.09E+02	1.18E+04	5.09E+02
108-38-3	Xylene, m-	μg/L	1	6.15E+05	6.15E+05	-	6.15E+05	6.15E+05	-	1.19E+05	1.19E+05
95-47-6	Xylene, o-	μg/L	-	6.84E+05	6.84E+05	-	6.84E+05	6.84E+05	-	1.33E+05	1.33E+05
106-42-3	Xylene, p-	μg/L	-	6.57E+05	6.57E+05	-	6.57E+05	6.57E+05	-	1.28E+05	1.28E+05
1330-20-7	Xylene, Mixture	μg/L	-	6.48E+05	6.48E+05	-	6.48E+05	6.48E+05	-	1.26E+05	1.26E+05

		Recreational User Swimming	Adult Recre Swim			eational User nming	Teen Recrea	
Analyte	Units	Cancer ⁱ	Hazard	Action	Hazard	Action	Hazard	Action
Aluminum	μg/L	-	6.71E+06	6.71E+06	1.37E+06	1.37E+06	2.52E+06	2.52E+06
Antimony (metallic)	μg/L	-	1.34E+03	1.34E+03	4.06E+02	4.06E+02	6.94E+02	6.94E+02
Arsenic, Inorganic	μg/L	1.71E+02	2.01E+03	1.71E+02	4.11E+02	1.71E+02	7.57E+02	1.71E+02
Barium	μg/L	-	4.02E+05	4.02E+05	1.50E+05	1.50E+05	2.44E+05	2.44E+05
Beryllium and compounds	μg/L	-	5.16E+02	5.16E+02	2.79E+02	2.79E+02	4.07E+02	4.07E+02
Boron And Borates Only	μg/L	-	1.34E+06	1.34E+06	2.74E+05	2.74E+05	5.05E+05	5.05E+05
Cadmium (Water)	μg/L	-	1.55E+02	1.55E+02	6.30E+01	6.30E+01	9.99E+01	9.99E+01
Chromium(III), Insoluble Salts	μg/L	-	7.00E+05	7.00E+05	3.60E+05	3.60E+05	5.34E+05	5.34E+05
Chromium(VI)	μg/L	1.89E+01	1.35E+03	1.89E+01	6.96E+02	1.89E+01	1.03E+03	1.89E+01
Chromium (Total) ^b	-	-	-	-	-	-	-	-
Cobalt	μg/L	-	2.25E+03	2.25E+03	4.27E+02	4.27E+02	7.96E+02	7.96E+02
Copper	μg/L	-	2.68E+05	2.68E+05	5.49E+04	5.49E+04	1.01E+05	1.01E+05
Fluoride	μg/L	-	2.68E+05	2.68E+05	5.49E+04	5.49E+04	1.01E+05	1.01E+05
Iron	μg/L	-	4.69E+06	4.69E+06	9.60E+05	9.60E+05	1.77E+06	1.77E+06
Lead ^c	μg/L	-	-	3.00E+01	-	3.00E+01	-	3.00E+01
Manganese (Non-Diet)	μg/L	-	3.08E+04	3.08E+04	1.32E+04	1.32E+04	2.07E+04	2.07E+04
Mercury, Inorganic Salts	μg/L	-	6.03E+02	6.03E+02	2.25E+02	2.25E+02	3.66E+02	3.66E+02
Molybdenum	μg/L	-	3.35E+04	3.35E+04	6.86E+03	6.86E+03	1.26E+04	1.26E+04
Nickel Soluble Salts	μg/L	-	7.87E+04	7.87E+04	2.20E+04	2.20E+04	3.82E+04	3.82E+04
Selenium	μg/L	-	3.35E+04	3.35E+04	6.86E+03	6.86E+03	1.26E+04	1.26E+04
Silver	μg/L	-	9.68E+03	9.68E+03	3.67E+03	3.67E+03	5.94E+03	5.94E+03
Thallium (Soluble Salts)	μg/L	-	6.71E+01	6.71E+01	1.37E+01	1.37E+01	2.52E+01	2.52E+01
Uranium (Insoluble Compounds) ^d	μg/L	-	2.01E+04	2.01E+04	4.11E+03	4.11E+03	7.57E+03	7.57E+03
Uranium (Soluble Salts) ^d	μg/L	-	1.34E+03	1.34E+03	2.74E+02	2.74E+02	5.05E+02	5.05E+02
Vanadium and Compounds	μg/L	-	4.45E+03	4.45E+03	2.08E+03	2.08E+03	3.17E+03	3.17E+03
Zinc and Compounds	μg/L	-	2.16E+06	2.16E+06	4.22E+05	4.22E+05	7.83E+05	7.83E+05
Acenaphthene	μg/L	-	1.69E+04	1.69E+04	9.07E+03	9.07E+03	1.33E+04	1.33E+04
Acenaphthylene ^e	μg/L	-	1.69E+04	1.69E+04	9.07E+03	9.07E+03	1.33E+04	1.33E+04
Acrylonitrile	μg/L	4.60E+02	6.32E+04	4.60E+02	1.34E+04	4.60E+02	2.46E+04	4.60E+02
Anthracene	μg/L	-	4.50E+04	4.50E+04	2.50E+04	2.50E+04	3.62E+04	3.62E+04
Benzene	μg/L	1.84E+03	6.82E+03	1.84E+03	2.70E+03	1.84E+03	4.32E+03	1.84E+03
Bis(2-ethylhexyl)phthalate ^f	μg/L	2.01E+04	1.63E+05	2.01E+04	2.92E+04	2.01E+04	5.49E+04	2.01E+04
Bromodichloromethane	μg/L	2.71E+03	2.68E+04	2.71E+03	8.11E+03	2.71E+03	1.39E+04	2.71E+03
Carbazole	μg/L	1.41E+03	-	1.41E+03	-	1.41E+03	-	1.41E+03
Carbon Tetrachloride	μg/L	1.13E+03	5.06E+03	1.13E+03	2.18E+03	1.13E+03	3.41E+03	1.13E+03
Chloroform	μg/L	4.69E+03	2.73E+04	4.69E+03	9.05E+03	4.69E+03	1.52E+04	4.69E+03
Dichlorodifluoromethane (Freon-12) ^f	μg/L	-	4.51E+05	4.51E+05	1.62E+05	1.62E+05	2.66E+05	2.66E+05

		Recreational User		ational User		ational User	Teen Recre	
		Swimming		ming	Swim		Swim	
Analyte	Units	Canceri	Hazard	Action	Hazard	Action	Hazard	Action
Dichloroethane, 1,1-f	μg/L	2.65E+04	5.74E+05	2.65E+04	1.86E+05	2.65E+04	3.14E+05	2.65E+04
Dichloroethane, 1,2-	μg/L	2.01E+03	2.27E+04	2.01E+03	6.47E+03	2.01E+03	1.12E+04	2.01E+03
Dichloroethylene, 1,1-	μg/L	-	9.85E+04	9.85E+04	3.71E+04	3.71E+04	6.02E+04	6.02E+04
Dichloroethylene, 1,2- (Mixed Isomers)	μg/L	-	1.86E+04	1.86E+04	6.87E+03	6.87E+03	1.12E+04	1.12E+04
Dichloroethylene, <i>cis</i> -1,2-	μg/L	-	4.12E+03	4.12E+03	1.53E+03	1.53E+03	2.49E+03	2.49E+03
Dichloroethylene, <i>trans</i> -1,2-	μg/L	-	4.12E+04	4.12E+04	1.53E+04	1.53E+04	2.49E+04	2.49E+04
Dieldrin	μg/L	9.75E-01	1.10E+01	9.75E-01	5.99E+00	9.75E-01	8.71E+00	9.75E-01
Dioxins/Furans, Total (as TCDD) ^g	μg/L	2.17E-03	5.70E-03	2.17E-03	1.02E-03	1.02E-03	1.92E-03	1.92E-03
~HpCDD	μg/L	2.17E-01	5.70E-01	2.17E-01	1.02E-01	1.02E-01	1.92E-01	1.92E-01
~HpCDF, 2,3,7,8-	μg/L	2.17E-01	5.70E-01	2.17E-01	1.02E-01	1.02E-01	1.92E-01	1.92E-01
~HxCDD, 2,3,7,8-	μg/L	2.17E-02	5.70E-02	2.17E-02	1.02E-02	1.02E-02	1.92E-02	1.92E-02
~HxCDF, 2,3,7,8-	μg/L	2.17E-02	5.70E-02	2.17E-02	1.02E-02	1.02E-02	1.92E-02	1.92E-02
~OCDD	μg/L	7.23E+00	1.90E+01	7.23E+00	3.41E+00	3.41E+00	6.41E+00	6.41E+00
~OCDF	μg/L	7.23E+00	1.90E+01	7.23E+00	3.41E+00	3.41E+00	6.41E+00	6.41E+00
~PeCDD, 2,3,7,8-	μg/L	2.17E-03	5.70E-03	2.17E-03	1.02E-03	1.02E-03	1.92E-03	1.92E-03
~PeCDF, 1,2,3,7,8-	μg/L	7.23E-02	1.90E-01	7.23E-02	3.41E-02	3.41E-02	6.41E-02	6.41E-02
~PeCDF, 2,3,4,7,8-	μg/L	7.23E-03	1.90E-02	7.23E-03	3.41E-03	3.41E-03	6.41E-03	6.41E-03
~TCDD, 2,3,7,8-	μg/L	2.17E-03	5.70E-03	2.17E-03	1.02E-03	1.02E-03	1.92E-03	1.92E-03
~TCDF, 2,3,7,8-	μg/L	2.17E-02	5.70E-02	2.17E-02	1.02E-02	1.02E-02	1.92E-02	1.92E-02
Ethylbenzene	μg/L	3.63E+03	2.95E+04	3.63E+03	1.47E+04	3.63E+03	2.20E+04	3.63E+03
Fluoranthene	μg/L	-	3.26E+05	3.26E+05	5.85E+04	5.85E+04	1.10E+05	1.10E+05
Fluorene	μg/L	-	8.11E+03	8.11E+03	4.45E+03	4.45E+03	6.46E+03	6.46E+03
Hexachlorobenzene	μg/L	1.76E+02	8.13E+01	8.13E+01	1.46E+01	1.46E+01	2.75E+01	2.75E+01
Naphthalene	μg/L	3.22E+02	1.14E+04	3.22E+02	5.70E+03	3.22E+02	8.52E+03	8.52E+03
Nitroaniline, 2-	μg/L	-	3.39E+04	3.39E+04	1.02E+04	1.02E+04	1.75E+04	1.75E+04
Nitroso-di-N-propylamine, N-	μg/L	3.01E+01	-	3.01E+01	-	3.01E+01	-	3.01E+01
Pentachlorophenol	μg/L	1.92E+01	5.31E+02	1.92E+01	2.99E+02	1.92E+01	4.30E+02	1.92E+01
Phenanthrene ^e	μg/L	-	1.69E+04	1.69E+04	9.07E+03	9.07E+03	1.33E+04	1.33E+04
Polychlorinated Biphenyls, Total	μg/L	7.05E+02	-	7.05E+02	-	7.05E+02	-	7.05E+02
~Aroclor 1016	μg/L	4.03E+03	5.70E+02	5.70E+02	1.02E+02	1.02E+02	1.92E+02	1.92E+02
~Aroclor 1221	μg/L	4.29E+00	-	4.29E+00	-	4.29E+00	-	4.29E+00
~Aroclor 1232	μg/L	4.29E+00	-	4.29E+00	-	4.29E+00	-	4.29E+00
~Aroclor 1242	μg/L	1.41E+02	-	1.41E+02	-	1.41E+02	-	1.41E+02
~Aroclor 1248	μg/L	1.41E+02	-	1.41E+02	-	1.41E+02	-	1.41E+02
~Aroclor 1254	μg/L	1.41E+02	1.63E+02	1.41E+02	2.92E+01	2.92E+01	5.49E+01	5.49E+01
~Aroclor 1260	μg/L	1.41E+02	-	1.41E+02	-	1.41E+02	-	1.41E+02

		Recreational User Swimming	Adult Recreational User Swimming		Child Recre Swim	ational User ming		ational User ming
Analyte	Units	Canceri	Hazard	Action	Hazard	Action	Hazard	Action
Polycyclic aromatic hydrocarbons, Total Carcinogenich	μg/L	6.85E+01	2.44E+03	6.85E+01	4.39E+02	6.85E+01	8.24E+02	6.85E+01
~Benz[a]anthracene	μg/L	6.85E+02	-	6.85E+02	-	6.85E+02	-	6.85E+02
~Benzo[a]pyrene	μg/L	6.85E+01	2.44E+03	6.85E+01	4.39E+02	6.85E+01	8.24E+02	6.85E+01
~Benzo[b]fluoranthene	μg/L	6.85E+02	-	6.85E+02	-	6.85E+02	-	6.85E+02
~Benzo[k]fluoranthene	μg/L	6.85E+03	-	6.85E+03	-	6.85E+03	-	6.85E+03
~Chrysene	μg/L	6.85E+04	-	6.85E+04	-	6.85E+04	-	6.85E+04
~Dibenz[a,h]anthracene	μg/L	6.85E+01	-	6.85E+01	-	6.85E+01	-	6.85E+01
~Indeno[1,2,3-cd]pyrene	μg/L	6.85E+02	-	6.85E+02	-	6.85E+02	-	6.85E+02
Pyrene	μg/L	-	2.75E+03	2.75E+03	1.55E+03	1.55E+03	2.23E+03	2.23E+03
Tetrachloroethylene	μg/L	2.04E+04	3.81E+03	3.81E+03	1.88E+03	1.88E+03	2.83E+03	2.83E+03
Toluenef	μg/L	-	7.30E+04	7.30E+04	3.38E+04	3.38E+04	5.18E+04	5.18E+04
Trichloro-1,2,2-trifluoroethane, 1,1,2- (Freon-113) ^f	μg/L	-	3.07E+07	3.07E+07	1.39E+07	1.39E+07	2.14E+07	2.14E+07
Trichloroethane, 1,1,1-	μg/L	-	3.37E+06	3.37E+06	1.34E+06	1.34E+06	2.14E+06	2.14E+06
Trichloroethane, 1,1,2-	μg/L	2.85E+03	1.28E+04	2.85E+03	3.95E+03	2.85E+03	6.72E+03	2.85E+03
Trichloroethylene	μg/L	1.52E+03	9.03E+02	9.03E+02	3.50E+02	3.50E+02	5.64E+02	5.64E+02
Vinyl Chloride	μg/L	4.95E+00	7.93E+03	4.95E+00	2.67E+03	4.95E+00	4.45E+03	4.95E+00
Xylene, m-	μg/L	-	1.10E+05	1.10E+05	5.56E+04	5.56E+04	8.30E+04	8.30E+04
Xylene, o-	μg/L	-	1.23E+05	1.23E+05	6.09E+04	6.09E+04	9.13E+04	9.13E+04
Xylene, p-	μg/L	-	1.18E+05	1.18E+05	5.88E+04	5.88E+04	8.81E+04	8.81E+04
Xylene, Mixture	μg/L	-	1.17E+05	1.17E+05	5.82E+04	5.82E+04	8.71E+04	8.71E+04

		Recreational User Wading ^a	Wadinga			ational User ling ^a	Teen Recre Wad	
Analyte	Units	Cancer ⁱ	Hazard	Action	Hazard	Action	Hazard	Action
Aluminum	μg/L	-	6.11E+07	6.11E+07	1.37E+07	1.37E+07	1.76E+07	1.76E+07
Antimony (metallic)	μg/L	-	3.67E+03	3.67E+03	8.20E+02	8.20E+02	1.06E+03	1.06E+03
Arsenic, Inorganic	μg/L	1.41E+03	1.83E+04	1.41E+03	4.10E+03	1.41E+03	5.29E+03	1.41E+03
Barium	μg/L	-	8.56E+05	8.56E+05	1.91E+05	1.91E+05	2.47E+05	2.47E+05
Beryllium and compounds	μg/L	-	8.56E+02	8.56E+02	1.91E+02	1.91E+02	2.47E+02	2.47E+02
Boron And Borates Only	μg/L	-	1.22E+07	1.22E+07	2.73E+06	2.73E+06	3.53E+06	3.53E+06
Cadmium (Water)	μg/L	-	3.06E+02	3.06E+02	6.84E+01	6.84E+01	8.82E+01	8.82E+01
Chromium(III), Insoluble Salts	μg/L	-	1.19E+06	1.19E+06	2.67E+05	2.67E+05	3.44E+05	3.44E+05
Chromium(VI)	μg/L	1.39E+01	2.29E+03	1.39E+01	5.13E+02	1.39E+01	6.62E+02	1.39E+01
Chromium (Total) ^b	μg/L	-	-	-	-	-	-	-
Cobalt	μg/L	-	4.58E+04	4.58E+04	1.03E+04	1.03E+04	1.32E+04	1.32E+04
Copper	μg/L	-	2.45E+06	2.45E+06	5.47E+05	5.47E+05	7.06E+05	7.06E+05
Fluoride	μg/L	-	2.45E+06	2.45E+06	5.47E+05	5.47E+05	7.06E+05	7.06E+05
Iron	μg/L	-	4.28E+07	4.28E+07	9.57E+06	9.57E+06	1.24E+07	1.24E+07
Lead ^c	μg/L	-	-	3.00E+01	-	3.00E+01	-	3.00E+01
Manganese (Non-Diet)	μg/L	-	5.88E+04	5.88E+04	1.31E+04	1.31E+04	1.69E+04	1.69E+04
Mercury, Inorganic Salts	μg/L	-	1.28E+03	1.28E+03	2.87E+02	2.87E+02	3.71E+02	3.71E+02
Molybdenum	μg/L	-	3.06E+05	3.06E+05	6.84E+04	6.84E+04	8.82E+04	8.82E+04
Nickel Soluble Salts	μg/L	-	2.45E+05	2.45E+05	5.47E+04	5.47E+04	7.06E+04	7.06E+04
Selenium	μg/L	-	3.06E+05	3.06E+05	6.84E+04	6.84E+04	8.82E+04	8.82E+04
Silver	μg/L	-	2.04E+04	2.04E+04	4.56E+03	4.56E+03	5.88E+03	5.88E+03
Thallium (Soluble Salts)	μg/L	-	6.11E+02	6.11E+02	1.37E+02	1.37E+02	1.76E+02	1.76E+02
Uranium (Insoluble Compounds) ^d	μg/L	-	1.83E+05	1.83E+05	4.11E+04	4.11E+04	5.28E+04	5.28E+04
Uranium (Soluble Salts) ^d	μg/L	-	1.22E+04	1.22E+04	2.73E+03	2.73E+03	3.53E+03	3.53E+03
Vanadium and Compounds	μg/L	-	8.01E+03	8.01E+03	1.79E+03	1.79E+03	2.31E+03	2.31E+03
Zinc and Compounds	μg/L	-	3.06E+07	3.06E+07	6.84E+06	6.84E+06	8.82E+06	8.82E+06
Acenaphthene	μg/L	-	2.80E+04	2.80E+04	6.27E+03	6.27E+03	8.09E+03	8.09E+03
Acenaphthylene ^e	μg/L	-	2.80E+04	2.80E+04	6.27E+03	6.27E+03	8.09E+03	8.09E+03
Acrylonitrile	μg/L	2.91E+03	4.55E+05	2.91E+03	1.02E+05	2.91E+03	1.31E+05	2.91E+03
Anthracene	μg/L	-	7.36E+04	7.36E+04	1.65E+04	1.65E+04	2.13E+04	2.13E+04
Benzene	μg/L	2.18E+03	1.39E+04	2.18E+03	3.10E+03	2.18E+03	4.00E+03	2.18E+03
Bis(2-ethylhexyl)phthalate ^f	μg/L	-	-	-	-	-	-	-
Bromodichloromethane	μg/L	5.10E+03	7.32E+04	5.10E+03	1.64E+04	5.10E+03	2.11E+04	5.10E+03
Carbazole	μg/L	1.18E+03	-	1.18E+03	-	1.18E+03	-	1.18E+03
Carbon Tetrachloride	μg/L	1.19E+03	9.61E+03	1.19E+03	2.15E+03	1.19E+03	2.77E+03	1.19E+03
Chloroform	μg/L	7.34E+03	6.58E+04	7.34E+03	1.47E+04	7.34E+03	1.90E+04	7.34E+03
Dichlorodifluoromethane (Freon-12) ^f	μg/L	-	1.00E+06	1.00E+06	2.24E+05	2.24E+05	2.89E+05	2.89E+05

		Recreational User Wading ^a		ational User ling ^a		ational User ling ^a	Teen Recre Wad	ational User
Analyte	Units	Canceri	Hazard	Action	Hazard	Action	Hazard	Action
Dichloroethane, 1,1-f	μg/L	4.32E+04	1.42E+06	4.32E+04	3.18E+05	4.32E+04	4.11E+05	4.32E+04
Dichloroethane, 1,2-	μg/L	4.32E+03	6.83E+04	4.32E+03	1.53E+04	4.32E+03	1.97E+04	4.32E+03
Dichloroethylene, 1,1-	μg/L	-	2.09E+05	2.09E+05	4.68E+04	4.68E+04	6.02E+04	6.02E+04
Dichloroethylene, 1,2- (Mixed Isomers)	μg/L	-	3.99E+04	3.99E+04	8.92E+03	8.92E+03	1.15E+04	1.15E+04
Dichloroethylene, cis-1,2-	μg/L	-	8.86E+03	8.86E+03	1.98E+03	1.98E+03	2.56E+03	2.56E+03
Dichloroethylene, trans-1,2-	μg/L	-	8.86E+04	8.86E+04	1.98E+04	1.98E+04	2.56E+04	2.56E+04
Dieldrin	μg/L	7.82E-01	1.81E+01	7.82E-01	4.05E+00	7.82E-01	5.22E+00	7.82E-01
Dioxins/Furans, Total (as TCDD)g	μg/L	-	-	-	-	-	-	-
~HpCDD	μg/L	-	-	-	-	-	-	-
~HpCDF, 2,3,7,8-	μg/L	-	-	-	-	-	-	-
~HxCDD, 2,3,7,8-	μg/L	-	-	-	-	-	-	-
~HxCDF, 2,3,7,8-	μg/L	-	-	-	-	-	-	-
~OCDD	μg/L	-	-	-	-	-	-	-
~OCDF	μg/L	-	-	-	-	-	-	-
~PeCDD, 2,3,7,8-	μg/L	-	-	-	-	-	-	-
~PeCDF, 1,2,3,7,8-	μg/L	-	-	-	-	-	-	-
~PeCDF, 2,3,4,7,8-	μg/L	-	-	-	-	-	-	-
~TCDD, 2,3,7,8-	μg/L	-	-	-	-	-	-	-
~TCDF, 2,3,7,8-	μg/L	-	-	-	-	-	-	-
Ethylbenzene	μg/L	3.21E+03	5.10E+04	3.21E+03	1.14E+04	3.21E+03	1.47E+04	3.21E+03
Fluoranthene	μg/L	-	-	-	-	-	-	-
Fluorene	μg/L	-	1.33E+04	1.33E+04	2.98E+03	2.98E+03	3.85E+03	3.85E+03
Hexachlorobenzene	μg/L	-	-	-	-	-	-	-
Naphthalene	μg/L	2.83E+02	1.96E+04	2.83E+02	4.39E+03	2.83E+02	5.67E+03	5.67E+03
Nitroaniline, 2-	μg/L	-	9.32E+04	9.32E+04	2.09E+04	2.09E+04	2.69E+04	2.69E+04
Nitroso-di-N-propylamine, N-	μg/L	9.07E+01	-	9.07E+01	-	9.07E+01	-	9.07E+01
Pentachlorophenol	μg/L	1.49E+01	8.64E+02	1.49E+01	1.93E+02	1.49E+01	2.49E+02	1.49E+01
Phenanthrene ^e	μg/L	-	2.80E+04	2.80E+04	6.27E+03	6.27E+03	8.09E+03	8.09E+03
Polychlorinated Biphenyls, Total	μg/L	-	-	-	-	-	-	-
~Aroclor 1016	μg/L	-	-	-	-	-	-	-
~Aroclor 1221	μg/L	3.35E+00	-	3.35E+00	-	3.35E+00	-	3.35E+00
~Aroclor 1232	μg/L	3.35E+00	-	3.35E+00	-	3.35E+00	-	3.35E+00
~Aroclor 1242	μg/L	-	-	-	-	-	-	-
~Aroclor 1248	μg/L	-	-	-	-	-	-	-
~Aroclor 1254	μg/L	-	-	-	-	-	-	-
~Aroclor 1260	μg/L	-	-	-	-	-	-	-

Table A.3a. Surface Water Action Levels for Significant COPCs at PGDP (Continued)

(Values were calculated in October 2022 and are based on best available information.)

		Recreational User Wading ^a	Wading ^a			ational User ling ^a		ational User ling ^a
Analyte	Units	Canceri	Hazard	Action	Hazard	Action	Hazard	Action
Polycyclic aromatic hydrocarbons, Total Carcinogenich	μg/L	-	-	-	-	-	-	-
~Benz[a]anthracene	μg/L	-	-	-	-	-	-	-
~Benzo[a]pyrene	μg/L	-	-	1	-	-	-	-
~Benzo[b]fluoranthene	μg/L	-	-	1	-	-	-	-
~Benzo[k]fluoranthene	μg/L	-	-	1	-	-	-	-
~Chrysene	μg/L	-	-	ı	-	-	-	-
~Dibenz[a,h]anthracene	μg/L	-	-	-	-	-	-	-
~Indeno[1,2,3-cd]pyrene	μg/L	-	-	ı	-	-	-	-
Pyrene	μg/L	-	4.46E+03	4.46E+03	9.97E+02	9.97E+02	1.29E+03	1.29E+03
Tetrachloroethylene	μg/L	1.82E+04	6.63E+03	6.63E+03	1.48E+03	1.48E+03	1.91E+03	1.91E+03
Toluene ^f	μg/L	-	1.32E+05	1.32E+05	2.95E+04	2.95E+04	3.81E+04	3.81E+04
Trichloro-1,2,2-trifluoroethane, 1,1,2- (Freon-113) ^f	μg/L	-	5.63E+07	5.63E+07	1.26E+07	1.26E+07	1.63E+07	1.63E+07
Trichloroethane, 1,1,1-	μg/L	-	6.81E+06	6.81E+06	1.52E+06	1.52E+06	1.97E+06	1.97E+06
Trichloroethane, 1,1,2-	μg/L	5.11E+03	3.37E+04	5.11E+03	7.54E+03	5.11E+03	9.73E+03	5.11E+03
Trichloroethylene	μg/L	1.78E+03	1.86E+03	1.78E+03	4.17E+02	4.17E+02	5.38E+02	5.38E+02
Vinyl Chloride	μg/L	2.31E+01	1.89E+04	2.31E+01	4.22E+03	2.31E+01	5.44E+03	2.31E+01
Xylene, m-	μg/L	-	1.90E+05	1.90E+05	4.25E+04	4.25E+04	5.49E+04	5.49E+04
Xylene, o-	μg/L	-	2.13E+05	2.13E+05	4.76E+04	4.76E+04	6.15E+04	6.15E+04
Xylene, p-	μg/L	-	2.04E+05	2.04E+05	4.57E+04	4.57E+04	5.89E+04	5.89E+04
Xylene, Mixture	μg/L	-	2.01E+05	2.01E+05	4.51E+04	4.51E+04	5.82E+04	5.82E+04

NOTES: An HI of 3 is used for the AL derivation because the range of values for HI (based on RGO tables) are 0.1, 1, and 3. Values are provided in these tables for significant COPCs for PGDP. Values for other COPCs can be obtained using the RAIS Chemical PRG online calculator, as modified using PGDP-specific inputs. ALs are not adjusted for solubility limits.

^a Recreational User Wading and all Worker scenarios consider dermal contact only.

^b Chromium (Total) AL should utilize Chromium III or Chromium VI, as appropriate.

^c Lead values should be checked prior to use to ensure they are still current.

d Based on recommendation from EPA, ALs for uranium (soluble salts) now use the RfD and the RfC for soluble compounds of uranium derived from ATSDR (EPA 2016). ALs for uranium (insoluble compounds) use the RfD for uranium (soluble salts), which is available in IRIS; the RfC for insoluble compounds of uranium are derived from ATSDR.

^e Acenaphthylene and Phenanthrene use values for Acenaphthene as a surrogate.

^f Analytes are not PGDP-significant COPCs (Table 2.1), but are provided for project support.

g Total dioxins/furans uses values for 2,3,7,8-TCDD, see screening note 9f in the Appendix A introduction, "Screening Levels," on pages A-3-A-5.

h Total carcinogenic PAHs uses values for BaP, see screening note 9d in the Appendix A introduction, "Screening Levels," on pages A-3-A-5.

For the recreational user, ELCRs (i.e., cancer ALs) were calculated using the child/teen/adult age-adjusted lifetime scenario.

CAS Number	Analyte	Units	Outdoor Worker Wading	Excavation Worker Wading	Industrial Worker Wading
14596-10-2	Am-241	pCi/L			
10045-97-3	Cs-137	pCi/L			
13994-20-2	Np-237	pCi/L			
13981-16-3	Pu-238	pCi/L			
15117-48-3	Pu-239	pCi/L			
14119-33-6	Pu-240	pCi/L	No Son	eening Values for the Wading Sce	marial
14133-76-7	Tc-99	pCi/L	NO SCI	eening values for the wading see	chario
14269-63-7	Th-230	pCi/L			
7440-29-1	Th-232°	pCi/L			
13966-29-5	U-234	pCi/L			
15117-96-1	U-235	pCi/L			
7440-61-1	U-238	pCi/L			

Table A.3b. Surface Water Action Levels for Significant Radionuclide COPCs at PGDP (Continued)

(Values were calculated in October 2022 and are based on best available information.)

CAS		TT - *4	I	Recreational User Swimming	.b	Recreational User Wading						
Number	Analyte	Units	Peak Risk Infinite ^c	Peak Risk 1,000 Years	Secular Equilibrium	Peak Risk Infinite	Peak Risk 1,000 Years	Secular Equilibrium				
14596-10-2	Am-241	pCi/L	3.49E+03	3.49E+03	4.59E+02							
10045-97-3	Cs-137	pCi/L	1.54E+04	1.54E+04	1.16E+04	1						
13994-20-2	Np-237	pCi/L	6.53E+02	5.15E+03	5.30E+02	1						
13981-16-3	Pu-238	pCi/L	2.98E+03	2.98E+03	1.06E+02							
15117-48-3	Pu-239	pCi/L	2.62E+03	2.62E+03	4.08E+02	1						
14119-33-6	Pu-240	pCi/L	2.62E+03	2.62E+03	2.16E+02	No Concenino	Values for the V	Vadina Caanania				
14133-76-7	Tc-99	pCi/L	1.29E+05	1.29E+05	1.29E+05	No Screening	values for the v	Vading Scenario ^a				
14269-63-7	Th-230	pCi/L	1.22E+02	3.15E+02	1.12E+02							
7440-29-1	Th-232 ^d	pCi/L	2.46E+02	2.46E+02	2.46E+02							
13966-29-5	U-234	pCi/L	1.86E+02	4.64E+03	1.10E+02							
15117-96-1	U-235	pCi/L	4.84E+02	4.16E+03	4.84E+02							
7440-61-1	U-238	pCi/L	1.07E+02	4.05E+03	1.07E+02							

NOTES: Values are provided in these tables for significant radionuclide COPCs for PGDP. Values for other radionuclides can be obtained using the EPA Radionuclide PRG calculator, as modified using PGDP-specific inputs.

Radionuclide ALs are based on the cancer endpoint using a target ELCR of 1.0E-4. The surface water radionuclide ALs are based on the ingestion pathway only.

^a The wading exposure scenarios shown in this table only consider dermal contact with surface water, not incidental surface water ingestion (see Figure 8.1). Dermal absorption is not evaluated for radionuclides so no screening values are derived for the wading exposure scenarios.

^b For the recreational user, the radionuclide ALs were calculated using the child/adult age-adjusted scenario (i.e., 26-year exposure duration, with 6 years as a child and 20 years as an adult). The EPA Radiological PRG calculator only allows exposure parameter entries for two life stages (i.e., child and adult), rather than the four life stage groups provided in the EPA and RAIS chemical online calculators. Thus, the EPA Radiological PRG calculator does not allow entry of the teen exposure parameters shown in Table B.5.

^c The time frame for calculating peak risk over an infinite period in the EPA Radionuclide PRG calculator is 1 trillion years. NASA most recently has estimated the age of the universe as being 13.7 billion years, with an uncertainty of only 200 million years (https://imagine.gsfc.nasa.gov).

^d Analyte is not a PGDP significant COPC (Table 2.1), but it is provided for project support.

CAS			Oı	ıtdoor Work	xer	Exc	avation Wo	rker	Ind	ustrial Wor	·ker
Number	Analyte	Units	Cancer	Hazard	No Action	Cancer	Hazard	No Action	Cancer	Hazard	No Action
7429-90-5	Aluminum	mg/kg	-	3.26E+04	3.26E+04	-	3.26E+04	3.26E+04	-	1.00E+05	1.00E+05
7440-36-0	Antimony (metallic)	mg/kg	-	1.32E+01	1.32E+01	-	1.32E+01	1.32E+01	-	9.34E+01	9.34E+01
7440-38-2	Arsenic, Inorganic	mg/kg	7.48E-01	1.20E+01	7.48E-01	3.74E+00	1.20E+01	3.74E+00	1.60E+00	2.57E+01	1.60E+00
7440-39-3	Barium	mg/kg	-	6.47E+03	6.47E+03	-	6.47E+03	6.47E+03	-	4.04E+04	4.04E+04
7440-41-7	Beryllium and compounds	mg/kg	9.39E+03	6.55E+01	6.55E+01	4.69E+04	6.55E+01	6.55E+01	6.95E+03	4.50E+02	4.50E+02
7440-42-8	Boron And Borates Only	mg/kg	-	6.57E+03	6.57E+03	-	6.57E+03	6.57E+03	-	4.65E+04	4.65E+04
7440-43-9	Cadmium (Diet)	mg/kg	1.25E+04	2.54E+00	2.54E+00	6.26E+04	2.54E+00	2.54E+00	9.26E+03	6.11E+00	6.11E+00
16065-83-1	Chromium(III), Insoluble Salts	mg/kg	-	4.93E+04	4.93E+04	-	4.93E+04	4.93E+04	-	1.00E+05	1.00E+05
18540-29-9	Chromium(VI)	mg/kg	1.83E+00	9.85E+01	1.83E+00	9.14E+00	9.85E+01	9.14E+00	1.23E+01	6.93E+02	1.23E+01
7440-47-3	Chromium (Total) ^a	mg/kg	-	-	-	-	-	-	-	-	-
7440-48-4	Cobalt	mg/kg	2.50E+03	9.84E+00	9.84E+00	1.25E+04	9.84E+00	9.84E+00	1.85E+03	6.87E+01	6.87E+01
7440-50-8	Copper	mg/kg	-	1.32E+03	1.32E+03	-	1.32E+03	1.32E+03	-	9.34E+03	9.34E+03
16984-48-8	Fluoride	mg/kg	-	1.32E+03	1.32E+03	-	1.32E+03	1.32E+03	-	9.33E+03	9.33E+03
7439-89-6	Iron	mg/kg	-	2.30E+04	2.30E+04	-	2.30E+04	2.30E+04	-	1.00E+05	1.00E+05
7439-92-1	Lead ^b	mg/kg	-	-	8.00E+02	-	-	8.00E+02	-	-	8.00E+02
7439-96-5	Manganese (Non-Diet)	mg/kg	-	7.74E+02	7.74E+02	-	7.74E+02	7.74E+02	-	4.72E+03	4.72E+03
Various	Mercury, Inorganic Salts	mg/kg	-	9.86E+00	9.86E+00	-	9.86E+00	9.86E+00	-	7.01E+01	7.01E+01
7439-98-7	Molybdenum	mg/kg	-	1.64E+02	1.64E+02	-	1.64E+02	1.64E+02	-	1.17E+03	1.17E+03
7440-02-0	Nickel Soluble Salts	mg/kg	8.66E+04	6.52E+02	6.52E+02	1.00E+05	6.52E+02	6.52E+02	6.41E+04	4.30E+03	4.30E+03
7782-49-2	Selenium	mg/kg	1	1.64E+02	1.64E+02	-	1.64E+02	1.64E+02	-	1.17E+03	1.17E+03
7440-22-4	Silver	mg/kg	1	1.64E+02	1.64E+02	-	1.64E+02	1.64E+02	-	1.17E+03	1.17E+03
7440-28-0	Thallium (Soluble Salts)	mg/kg	1	3.29E-01	3.29E-01	-	3.29E-01	3.29E-01	-	2.34E+00	2.34E+00
N/A	Uranium (Insoluble Compounds) ^c	mg/kg	1	9.83E+01	9.83E+01	-	9.83E+01	9.83E+01	-	6.81E+02	6.81E+02
7440-61-1	Uranium (Soluble Salts) ^c	mg/kg	1	6.58E+00	6.58E+00	-	6.58E+00	6.58E+00	-	4.66E+01	4.66E+01
7440-62-2	Vanadium and Compounds	mg/kg	-	1.65E+02	1.65E+02	-	1.65E+02	1.65E+02	-	1.15E+03	1.15E+03
7440-66-6	Zinc and Compounds	mg/kg	ı	9.86E+03	9.86E+03	-	9.86E+03	9.86E+03	-	7.01E+04	7.01E+04
83-32-9	Acenaphthene	mg/kg	ı	1.01E+03	1.01E+03	-	1.01E+03	1.01E+03	-	1.38E+03	1.38E+03
208-96-8	Acenaphthylene ^d	mg/kg	ı	1.01E+03	1.01E+03	-	1.01E+03	1.01E+03	-	1.38E+03	1.38E+03
107-13-1	Acrylonitrile	mg/kg	8.93E-01	8.85E+00	8.93E-01	4.46E+00	8.85E+00	4.46E+00	1.24E+00	6.71E+00	1.24E+00
120-12-7	Anthracene	mg/kg	ı	5.05E+03	5.05E+03	-	5.05E+03	5.05E+03	-	6.89E+03	6.89E+03
71-43-2	Benzene	mg/kg	5.19E+00	4.25E+01	5.19E+00	2.59E+01	4.25E+01	2.59E+01	5.31E+00	4.43E+01	5.31E+00
117-81-7	Bis(2-ethylhexyl)phthalate ^e	mg/kg	3.79E+01	3.79E+02	3.79E+01	1.90E+02	3.79E+02	1.90E+02	5.80E+01	5.80E+02	5.80E+01
75-27-4	Bromodichloromethane	mg/kg	1.59E+00	2.63E+02	1.59E+00	7.93E+00	2.63E+02	7.93E+00	1.30E+00	1.87E+03	1.30E+00
86-74-8	Carbazole	mg/kg	2.65E+01	-	2.65E+01	1.33E+02	-	1.33E+02	4.06E+01	-	4.06E+01
56-23-5	Carbon Tetrachloride	mg/kg	3.14E+00	5.29E+01	3.14E+00	1.57E+01	5.29E+01	1.57E+01	2.96E+00	6.12E+01	2.96E+00
67-66-3	Chloroform	mg/kg	1.78E+00	1.04E+02	1.78E+00	8.90E+00	1.04E+02	8.90E+00	1.39E+00	1.07E+02	1.39E+00
75-71-8	Dichlorodifluoromethane (Freon-12) ^e	mg/kg	1	4.94E+01	4.94E+01	-	4.94E+01	4.94E+01	-	3.68E+01	3.68E+01

Table A.4a. Soil/Sediment No Action Levels for Significant COPCs at PGDP (Continued) (Values were calculated in October 2022 and are based on best available information.)

CAS			Oı	ıtdoor Work	ker	Exc	avation Wo	rker	Ind	ustrial Wor	rker
Number	Analyte	Units	Cancer	Hazard	No Action	Cancer	Hazard	No Action	Cancer	Hazard	No Action
75-34-3	Dichloroethane, 1,1-e	mg/kg	1.90E+01	5.64E+02	1.90E+01	9.52E+01	5.64E+02	9.52E+01	1.58E+01	4.52E+02	1.58E+01
107-06-2	Dichloroethane, 1,2-	mg/kg	2.26E+00	1.73E+01	2.26E+00	1.13E+01	1.73E+01	1.13E+01	2.09E+00	1.39E+01	2.09E+00
75-35-4	Dichloroethylene, 1,1-	mg/kg	-	1.26E+02	1.26E+02	-	1.26E+02	1.26E+02	-	1.00E+02	1.00E+02
540-59-0	Dichloroethylene, 1,2- (Mixed Isomers)	mg/kg	ı	2.96E+02	2.96E+02	-	2.96E+02	2.96E+02	-	2.10E+03	2.10E+03
156-59-2	Dichloroethylene, cis-1,2-	mg/kg	ı	3.11E+01	3.11E+01	-	3.12E+01	3.12E+01	-	4.0E+01	4.0E+01
156-60-5	Dichloroethylene, trans-1,2-	mg/kg	ı	3.89E+01	3.89E+01	-	3.89E+01	3.89E+01	-	3.04E+01	3.04E+01
60-57-1	Dieldrin	mg/kg	3.32E-02	9.48E-01	3.32E-02	1.66E-01	9.48E-01	1.66E-01	5.08E-02	1.45E+00	5.08E-02
1746-01-6	Dioxins/Furans, Total (as TCDD) ^f	mg/kg	5.76E-06	1.89E-05	5.76E-06	2.88E-05	1.89E-05	1.89E-05	1.57E-05	5.24E-05	1.57E-05
37871-00-4	~HpCDD, 2,3,7,8-	mg/kg	5.77E-04	1.89E-03	5.77E-04	2.89E-03	1.89E-03	1.89E-03	1.58E-03	5.24E-03	1.58E-03
38998-75-3	~HpCDF, 2,3,7,8-	mg/kg	5.79E-04	1.89E-03	5.79E-04	2.90E-03	1.89E-03	1.89E-03	1.60E-03	5.24E-03	1.60E-03
34465-46-8	~HxCDD	mg/kg	5.80E-05	1.89E-04	5.80E-05	2.90E-04	1.89E-04	1.89E-04	1.61E-04	5.25E-04	1.61E-04
55684-94-1	~HxCDF, 2,3,7,8-	mg/kg	5.80E-05	1.89E-04	5.80E-05	2.90E-04	1.89E-04	1.89E-04	1.61E-04	5.25E-04	1.61E-04
3268-87-9	~OCDD	mg/kg	1.93E-02	6.29E-02	1.93E-02	9.67E-02	6.29E-02	6.29E-02	5.38E-02	1.75E-01	5.38E-02
39001-02-0	~OCDF	mg/kg	1.93E-02	6.29E-02	1.93E-02	9.67E-02	6.29E-02	6.29E-02	5.38E-02	1.75E-01	5.38E-02
36088-22-9	~PeCDD, 2,3,7,8-	mg/kg	5.80E-06	1.89E-05	5.80E-06	2.90E-05	1.89E-05	1.89E-05	1.61E-05	5.25E-05	1.61E-05
57117-41-6	~PeCDF, 1,2,3,7,8-	mg/kg	1.93E-04	6.29E-04	1.93E-04	9.67E-04	6.29E-04	6.29E-04	5.38E-04	1.75E-03	5.38E-04
57117-31-4	~PeCDF, 2,3,4,7,8-	mg/kg	1.93E-05	6.29E-05	1.93E-05	9.67E-05	6.29E-05	6.29E-05	5.38E-05	1.75E-04	5.38E-05
1746-01-6	~TCDD, 2,3,7,8-	mg/kg	5.76E-06	1.89E-05	5.76E-06	2.88E-05	1.89E-05	1.89E-05	1.57E-05	5.24E-05	1.57E-05
51207-31-9	~TCDF, 2,3,7,8-	mg/kg	5.77E-05	1.89E-04	5.77E-05	2.89E-04	1.89E-04	1.89E-04	1.58E-04	5.24E-04	1.58E-04
100-41-4	Ethylbenzene	mg/kg	2.59E+01	1.10E+03	2.59E+01	1.30E+02	1.10E+03	1.30E+02	2.66E+01	2.05E+03	2.66E+01
206-44-0	Fluoranthene	mg/kg	-	6.73E+02	6.73E+02	-	6.73E+02	6.73E+02	-	9.19E+02	9.19E+02
86-73-7	Fluorene	mg/kg	-	6.73E+02	6.73E+02	-	6.73E+02	6.73E+02	-	9.19E+02	9.19E+02
118-74-1	Hexachlorobenzene	mg/kg	4.66E-01	3.29E-01	3.29E-01	2.33E+00	3.29E-01	3.29E-01	1.26E+00	2.34E+00	1.26E+00
91-20-3	Naphthalene	mg/kg	3.34E+00	6.61E+01	3.34E+00	1.67E+01	6.61E+01	1.67E+01	4.06E+00	5.38E+01	4.06E+00
88-74-4	Nitroaniline, 2-	mg/kg	-	1.89E+02	1.89E+02	-	1.89E+02	1.89E+02	-	2.87E+02	2.87E+02
621-64-7	Nitroso-di-N-propylamine, N-	mg/kg	7.58E-02	-	7.58E-02	3.79E-01	-	3.79E-01	1.16E-01	-	1.16E-01
87-86-5	Pentachlorophenol	mg/kg	8.11E-01	5.80E+01	8.11E-01	4.06E+00	5.80E+01	4.06E+00	8.77E-01	6.27E+01	8.77E-01
85-01-8	Phenanthrene ^d	mg/kg	-	1.01E+03	1.01E+03	-	1.01E+03	1.01E+03	-	1.38E+03	1.38E+03
1336-36-3	Polychlorinated Biphenyls, Total	mg/kg	2.24E-01	-	2.24E-01	1.12E+00	-	1.12E+00	2.93E-01	-	2.93E-01
12674-11-2	~Aroclor 1016	mg/kg	6.40E+00	1.13E+00	1.13E+00	3.20E+01	1.13E+00	1.13E+00	8.39E+00	1.50E+00	1.50E+00
11104-28-2	~Aroclor 1221	mg/kg	2.19E-01	-	2.19E-01	1.09E+00	-	1.09E+00	2.81E-01	-	2.81E-01
11141-16-5	~Aroclor 1232	mg/kg	2.12E-01	-	2.12E-01	1.06E+00	-	1.06E+00	2.67E-01	-	2.67E-01
53469-21-9	~Aroclor 1242	mg/kg	2.24E-01	-	2.24E-01	1.12E+00	-	1.12E+00	2.94E-01	-	2.94E-01
12672-29-6	~Aroclor 1248	mg/kg	2.24E-01	-	2.24E-01	1.12E+00	-	1.12E+00	2.93E-01	-	2.93E-01
11097-69-1	~Aroclor 1254	mg/kg	2.25E-01	3.24E-01	2.25E-01	1.12E+00	3.24E-01	3.24E-01	2.96E-01	4.30E-01	2.96E-01
11096-82-5	~Aroclor 1260	mg/kg	2.26E-01	-	2.26E-01	1.13E+00	-	1.13E+00	2.98E-01	-	2.98E-01

CAS			Oı	ıtdoor Work	er	Exc	avation Wo	rker	Ind	ustrial Wor	ker
Number	Analyte	Units	Cancer	Hazard	No Action	Cancer	Hazard	No Action	Cancer	Hazard	No Action
50-32-8	Polycyclic aromatic hydrocarbons, Total Carcinogenic ^g	mg/kg	4.71E-01	5.03E+00	4.71E-01	2.35E+00	5.03E+00	2.35E+00	6.43E-01	6.85E+00	6.43E-01
56-55-3	~Benz[a]anthracene	mg/kg	4.69E+00	-	4.69E+00	2.35E+01	-	2.35E+01	6.39E+00	-	6.39E+00
50-32-8	~Benzo[a]pyrene	mg/kg	4.71E-01	5.03E+00	4.71E-01	2.35E+00	5.03E+00	2.35E+00	6.43E-01	6.85E+00	6.43E-01
205-99-2	~Benzo[b]fluoranthene	mg/kg	4.71E+00	-	4.71E+00	2.35E+01	-	2.35E+01	6.43E+00	-	6.43E+00
207-08-9	~Benzo[k]fluoranthene	mg/kg	4.71E+01	-	4.71E+01	2.35E+02	-	2.35E+02	6.43E+01	-	6.43E+01
218-01-9	~Chrysene	mg/kg	4.71E+02	-	4.71E+02	2.35E+03	-	2.35E+03	6.43E+02	-	6.43E+02
53-70-3	~Dibenz[a,h]anthracene	mg/kg	4.71E-01	-	4.71E-01	2.35E+00	-	2.35E+00	6.43E-01	-	6.43E-01
193-39-5	~Indeno[1,2,3-cd]pyrene	mg/kg	4.71E+00	-	4.71E+00	2.35E+01	-	2.35E+01	6.43E+00	-	6.43E+00
129-00-0	Pyrene	mg/kg	-	5.05E+02	5.05E+02	-	5.05E+02	5.05E+02	-	6.89E+02	6.89E+02
127-18-4	Tetrachloroethylene	mg/kg	1.12E+02	4.34E+01	4.34E+01	5.58E+02	4.34E+01	4.34E+01	1.07E+02	4.00E+01	4.00E+01
108-88-3	Toluene ^e	mg/kg	-	2.18E+03	2.18E+03	-	2.18E+03	2.18E+03	-	6.25E+03	6.25E+03
76-13-1	Trichloro-1,2,2-trifluoroethane, 1,1,2- (Freon-113) ^e	mg/kg	-	3.79E+03	3.79E+03	-	3.79E+03	3.79E+03	-	2.81E+03	2.81E+03
71-55-6	Trichloroethane, 1,1,1-	mg/kg	-	4.54E+03	4.54E+03	-	4.54E+03	4.54E+03	-	3.58E+03	3.58E+03
79-00-5	Trichloroethane, 1,1,2-	mg/kg	5.11E+00	8.49E-01	8.49E-01	2.56E+01	8.49E-01	8.49E-01	5.28E+00	6.32E-01	6.32E-01
79-01-6	Trichloroethylene	mg/kg	6.17E+00	2.26E+00	2.26E+00	3.09E+01	2.26E+00	2.26E+00	6.31E+00	1.90E+00	1.90E+00
75-01-4	Vinyl Chloride	mg/kg	9.44E-01	3.10E+01	9.44E-01	4.72E+00	3.10E+01	4.72E+00	2.06E+00	3.20E+01	2.06E+00
108-38-3	Xylene, m-	mg/kg	ı	3.08E+02	3.08E+02	-	3.08E+02	3.08E+02	-	2.38E+02	2.38E+02
95-47-6	Xylene, o-	mg/kg	ı	3.61E+02	3.61E+02	-	3.61E+02	3.61E+02	-	2.81E+02	2.81E+02
106-42-3	Xylene, p-	mg/kg	1	3.14E+02	3.14E+02	-	3.14E+02	3.14E+02	-	2.43E+02	2.43E+02
1330-20-7	Xylene, Mixture	mg/kg	-	3.23E+02	3.23E+02	-	3.23E+02	3.23E+02	-	2.50E+02	2.50E+02

CAS			Recreational User	Adult Recre	ational User	Child Recre	ational User	Teen Recrea	ational User
Number	Analyte	Units	Cancer ^h	Hazard	No Action	Hazard	No Action	Hazard	No Action
7429-90-5	Aluminum	mg/kg	-	1.00E+05	1.00E+05	1.95E+04	1.95E+04	1.00E+05	1.00E+05
7440-36-0	Antimony (metallic)	mg/kg	-	1.12E+02	1.12E+02	7.82E+00	7.82E+00	4.59E+01	4.59E+01
7440-38-2	Arsenic, Inorganic	mg/kg	8.09E-01	3.49E+01	8.09E-01	6.13E+00	8.09E-01	1.21E+01	8.09E-01
7440-39-3	Barium	mg/kg	-	5.35E+04	5.35E+04	3.89E+03	3.89E+03	2.23E+04	2.23E+04
7440-41-7	Beryllium and compounds	mg/kg	2.12E+04	5.55E+02	5.55E+02	3.91E+01	3.91E+01	2.28E+02	2.28E+02
7440-42-8	Boron And Borates Only	mg/kg	-	5.61E+04	5.61E+04	3.91E+03	3.91E+03	2.29E+04	2.29E+04
7440-43-9	Cadmium (Diet)	mg/kg	2.82E+04	8.22E+00	8.22E+00	1.33E+00	1.33E+00	2.87E+00	2.87E+00
16065-83-1	Chromium(III), Insoluble Salts	mg/kg	-	1.00E+05	1.00E+05	2.93E+04	2.93E+04	1.00E+05	1.00E+05
18540-29-9	Chromium(VI)	mg/kg	7.47E-01	8.39E+02	7.47E-01	5.86E+01	7.47E-01	3.43E+02	7.47E-01
7440-47-3	Chromium (Total) ^a	mg/kg	-	-	-	-	-	-	-
7440-48-4	Cobalt	mg/kg	5.65E+03	8.37E+01	8.37E+01	5.86E+00	5.86E+00	3.43E+01	3.43E+01
7440-50-8	Copper	mg/kg	-	1.12E+04	1.12E+04	7.82E+02	7.82E+02	4.59E+03	4.59E+03
16984-48-8	Fluoride	mg/kg	-	1.12E+04	1.12E+04	7.82E+02	7.82E+02	4.59E+03	4.59E+03
7439-89-6	Iron	mg/kg	-	1.00E+05	1.00E+05	1.37E+04	1.37E+04	8.03E+04	8.03E+04
7439-92-1	Lead ^b	mg/kg	-	-	4.00E+02	-	4.00E+02	-	4.00E+02
7439-96-5	Manganese (Non-Diet)	mg/kg	-	6.36E+03	6.36E+03	4.67E+02	4.67E+02	2.67E+03	2.67E+03
Various	Mercury, Inorganic Salts	mg/kg	-	8.42E+01	8.42E+01	5.87E+00	5.87E+00	3.44E+01	3.44E+01
7439-98-7	Molybdenum	mg/kg	-	1.40E+03	1.40E+03	9.78E+01	9.78E+01	5.73E+02	5.73E+02
7440-02-0	Nickel Soluble Salts	mg/kg	1.00E+05	5.47E+03	5.47E+03	3.90E+02	3.90E+02	2.26E+03	2.26E+03
7782-49-2	Selenium	mg/kg	-	1.40E+03	1.40E+03	9.78E+01	9.78E+01	5.74E+02	5.74E+02
7440-22-4	Silver	mg/kg	-	1.40E+03	1.40E+03	9.78E+01	9.78E+01	5.74E+02	5.74E+02
7440-28-0	Thallium (Soluble Salts)	mg/kg	-	2.81E+00	2.81E+00	1.96E-01	1.96E-01	1.15E+00	1.15E+00
N/A	Uranium (Insoluble Compounds) ^c	mg/kg	-	8.35E+02	8.35E+02	5.86E+01	5.86E+01	3.42E+02	3.42E+02
7440-61-1	Uranium (Soluble Salts) ^c	mg/kg	-	5.61E+01	5.61E+01	3.91E+00	3.91E+00	2.29E+01	2.29E+01
7440-62-2	Vanadium and Compounds	mg/kg	-	1.41E+03	1.41E+03	9.85E+01	9.85E+01	5.76E+02	5.76E+02
7440-66-6	Zinc and Compounds	mg/kg	-	8.42E+04	8.42E+04	5.87E+03	5.87E+03	3.44E+04	3.44E+04
83-32-9	Acenaphthene	mg/kg	-	1.91E+03	1.91E+03	4.61E+02	4.61E+02	6.40E+02	6.40E+02
208-96-8	Acenaphthylene ^d	mg/kg	-	1.91E+03	1.91E+03	4.61E+02	4.61E+02	6.40E+02	6.40E+02
107-13-1	Acrylonitrile	mg/kg	1.80E+00	2.57E+01	1.80E+00	1.75E+01	1.80E+00	1.89E+01	1.80E+00
120-12-7	Anthracene	mg/kg	-	9.53E+03	9.53E+03	2.31E+03	2.31E+03	3.20E+03	3.20E+03
71-43-2	Benzene	mg/kg	1.09E+01	1.54E+02	1.09E+01	4.92E+01	1.09E+01	1.03E+02	1.09E+01
117-81-7	Bis(2-ethylhexyl)phthalate ^e	mg/kg	3.32E+01	7.99E+02	3.32E+01	1.79E+02	3.32E+01	2.70E+02	3.32E+01
75-27-4	Bromodichloromethane	mg/kg	3.49E+00	2.25E+03	3.49E+00	1.56E+02	3.49E+00	9.18E+02	3.49E+00
86-74-8	Carbazole	mg/kg	2.32E+01	-	2.32E+01	-	2.32E+01	-	2.32E+01
56-23-5	Carbon Tetrachloride	mg/kg	6.72E+00	2.06E+02	6.72E+00	5.51E+01	6.72E+00	1.33E+02	6.72E+00
67-66-3	Chloroform	mg/kg	3.96E+00	3.75E+02	3.96E+00	1.22E+02	3.96E+00	2.51E+02	3.96E+00
75-71-8	Dichlorodifluoromethane (Freon-12) ^e	mg/kg	-	1.41E+02	1.41E+02	1.02E+02	1.02E+02	1.05E+02	1.05E+02

CAS			Recreational User	Adult Recre	ational User	Child Recre	ational User	Teen Recrea	tional User
Number	Analyte	Units	Cancerg	Hazard	No Action	Hazard	No Action	Hazard	No Action
75-34-3	Dichloroethane, 1,1-e	mg/kg	4.18E+01	1.70E+03	4.18E+01	9.78E+02	4.18E+01	1.23E+03	4.18E+01
107-06-2	Dichloroethane, 1,2-	mg/kg	4.86E+00	5.23E+01	4.86E+00	2.99E+01	4.86E+00	3.79E+01	4.86E+00
75-35-4	Dichloroethylene, 1,1-	mg/kg	-	3.79E+02	3.79E+02	2.23E+02	2.23E+02	2.75E+02	2.75E+02
540-59-0	Dichloroethylene, 1,2- (Mixed Isomers)	mg/kg	-	2.53E+03	2.53E+03	1.76E+02	1.76E+02	1.03E+03	1.03E+03
156-59-2	Dichloroethylene, cis-1,2-	mg/kg	-	1.30E+02	1.30E+02	2.98E+01	2.98E+01	8.10E+01	8.10E+01
156-60-5	Dichloroethylene, trans-1,2-	mg/kg	-	1.15E+02	1.15E+02	7.14E+01	7.14E+01	8.42E+01	8.42E+01
60-57-1	Dieldrin	mg/kg	2.90E-02	2.00E+00	2.90E-02	4.47E-01	2.90E-02	6.75E-01	2.90E-02
1746-01-6	Dioxins/Furans, Total (as TCDD) ^f	mg/kg	7.22E-06	6.99E+05	7.22E-06	1.01E-05	7.22E-06	2.47E-05	7.22E-06
37871-00-4	~HpCDD, 2,3,7,8-	mg/kg	7.23E-04	6.99E+03	7.23E-04	1.01E-03	7.23E-04	2.47E-03	7.23E-04
38998-75-3	~HpCDF, 2,3,7,8-	mg/kg	7.24E-04	6.99E+03	7.24E-04	1.01E-03	7.24E-04	2.47E-03	7.24E-04
34465-46-8	~HxCDD	mg/kg	7.25E-05	6.99E+03	7.25E-05	1.01E-04	7.25E-05	2.47E-04	7.25E-05
55684-94-1	~HxCDF, 2,3,7,8-	mg/kg	7.25E-05	7.00E+04	7.25E-05	1.01E-04	7.25E-05	2.47E-04	7.25E-05
3268-87-9	~OCDD	mg/kg	2.42E-02	2.33E-01	2.42E-02	3.36E-02	2.42E-02	8.24E-02	2.42E-02
39001-02-0	~OCDF	mg/kg	2.42E-02	2.33E-01	2.42E-02	3.36E-02	2.42E-02	8.24E-02	2.42E-02
36088-22-9	~PeCDD, 2,3,7,8-	mg/kg	7.25E-06	1.00E+05	7.25E-06	1.01E-05	7.25E-06	2.47E-05	7.25E-06
57117-41-6	~PeCDF, 1,2,3,7,8-	mg/kg	2.42E-04	2.33E-03	2.42E-04	3.36E-04	2.42E-04	8.24E-04	2.42E-04
57117-31-4	~PeCDF, 2,3,4,7,8-	mg/kg	2.42E-05	2.33E-04	2.42E-05	3.36E-05	2.42E-05	8.24E-05	2.42E-05
1746-01-6	~TCDD, 2,3,7,8-	mg/kg	7.22E-06	6.99E+05	7.22E-06	1.01E-05	7.22E-06	2.47E-05	7.22E-06
51207-31-9	~TCDF, 2,3,7,8-	mg/kg	7.23E-05	6.99E+04	7.23E-05	1.01E-04	7.23E-05	2.47E-04	7.23E-05
100-41-4	Ethylbenzene	mg/kg	5.46E+01	5.68E+03	5.46E+01	8.59E+02	5.46E+01	3.17E+03	5.46E+01
206-44-0	Fluoranthene	mg/kg	-	1.27E+03	1.27E+03	3.08E+02	3.08E+02	4.27E+02	4.27E+02
86-73-7	Fluorene	mg/kg	-	1.27E+03	1.27E+03	3.08E+02	3.08E+02	4.27E+02	4.27E+02
118-74-1	Hexachlorobenzene	mg/kg	8.86E-01	2.81E+00	8.86E-01	1.96E-01	1.96E-01	1.15E+00	8.86E-01
91-20-3	Naphthalene	mg/kg	2.99E+00	1.71E+02	2.99E+00	8.16E+01	2.99E+00	9.58E+01	2.99E+00
88-74-4	Nitroaniline, 2-	mg/kg	-	3.98E+02	3.98E+02	8.93E+01	8.93E+01	1.35E+02	1.35E+02
621-64-7	Nitroso-di-N-propylamine, N-	mg/kg	6.63E-02	-	6.63E-02	-	6.63E-02	-	6.63E-02
87-86-5	Pentachlorophenol	mg/kg	5.56E-01	8.73E+01	5.56E-01	2.46E+01	5.56E-01	2.90E+01	5.56E-01
85-01-8	Phenanthrene ^d	mg/kg	-	1.91E+03	1.91E+03	4.61E+02	4.61E+02	6.40E+02	6.40E+02
1336-36-3	Polychlorinated Biphenyls, Total	mg/kg	1.79E-01	-	1.79E-01	-	1.79E-01	-	1.79E-01
12674-11-2	~Aroclor 1016	mg/kg	5.12E+00	2.08E+00	2.08E+00	5.14E-01	5.14E-01	6.98E-01	6.98E-01
11104-28-2	~Aroclor 1221	mg/kg	1.77E-01	1	1.77E-01	1	1.77E-01	ı	1.77E-01
11141-16-5	~Aroclor 1232	mg/kg	1.76E-01	1	1.76E-01	1	1.76E-01	ı	1.76E-01
53469-21-9	~Aroclor 1242	mg/kg	1.79E-01	1	1.79E-01	-	1.79E-01	-	1.79E-01
12672-29-6	~Aroclor 1248	mg/kg	1.79E-01	ı	1.79E-01	ı	1.79E-01	ı	1.79E-01
11097-69-1	~Aroclor 1254	mg/kg	1.79E-01	5.95E-01	1.79E-01	1.47E-01	1.47E-01	2.00E-01	1.79E-01
11096-82-5	~Aroclor 1260	mg/kg	1.80E-01	-	1.80E-01	-	1.80E-01	-	1.80E-01

CAS			Recreational User	Adult Recre	ational User	Child Recrea	ational User	Teen Recrea	ational User
Number	Analyte	Units	Cancer ^h	Hazard	No Action	Hazard	No Action	Hazard	No Action
50-32-8	Polycyclic aromatic hydrocarbons, Total Carcinogenic ^g	mg/kg	1.09E-01	9.51E+00	1.09E-01	2.31E+00	1.09E-01	3.20E+00	1.09E-01
56-55-3	~Benz[a]anthracene	mg/kg	1.09E+00	-	1.09E+00	-	1.09E+00	-	1.09E+00
50-32-8	~Benzo[a]pyrene	mg/kg	1.09E-01	9.51E+00	1.09E-01	2.31E+00	1.09E-01	3.20E+00	1.09E-01
205-99-2	~Benzo[b]fluoranthene	mg/kg	1.09E+00	-	1.09E+00	-	1.09E+00	-	1.09E+00
207-08-9	~Benzo[k]fluoranthene	mg/kg	1.09E+01	-	1.09E+01	-	1.09E+01	-	1.09E+01
218-01-9	~Chrysene	mg/kg	1.09E+02	-	1.09E+02	-	1.09E+02	-	1.09E+02
53-70-3	~Dibenz[a,h]anthracene	mg/kg	1.09E-01	-	1.09E-01	-	1.09E-01	-	1.09E-01
193-39-5	~Indeno[1,2,3-cd]pyrene	mg/kg	1.09E+00	-	1.09E+00	-	1.09E+00	-	1.09E+00
129-00-0	Pyrene	mg/kg	-	9.53E+02	9.53E+02	2.31E+02	2.31E+02	3.20E+02	3.20E+02
127-18-4	Tetrachloroethylene	mg/kg	2.38E+02	1.45E+02	1.45E+02	5.87E+01	5.87E+01	1.00E+02	1.00E+02
108-88-3	Toluene ^e	mg/kg	-	1.38E+04	1.38E+04	1.48E+03	1.48E+03	6.84E+03	6.84E+03
76-13-1	Trichloro-1,2,2-trifluoroethane, 1,1,2- (Freon-113) ^e	mg/kg	-	1.08E+04	1.08E+04	7.94E+03	7.94E+03	8.03E+03	8.03E+03
71-55-6	Trichloroethane, 1,1,1-	mg/kg	-	1.36E+04	1.36E+04	8.16E+03	8.16E+03	9.87E+03	9.87E+03
79-00-5	Trichloroethane, 1,1,2-	mg/kg	1.07E+01	2.43E+00	2.43E+00	1.77E+00	1.77E+00	1.80E+00	1.80E+00
79-01-6	Trichloroethylene	mg/kg	8.32E+00	7.07E+00	7.07E+00	3.53E+00	3.53E+00	5.04E+00	5.04E+00
75-01-4	Vinyl Chloride	mg/kg	6.78E-02	1.12E+02	6.78E-02	3.64E+01	6.78E-02	7.49E+01	6.78E-02
108-38-3	Xylene, m-	mg/kg	-	9.06E+02	9.06E+02	5.82E+02	5.82E+02	6.64E+02	6.64E+02
95-47-6	Xylene, o-	mg/kg	-	1.07E+03	1.07E+03	6.69E+02	6.69E+02	7.80E+02	7.80E+02
106-42-3	Xylene, p-	mg/kg	-	9.24E+02	9.24E+02	5.92E+02	5.92E+02	6.77E+02	6.77E+02
1330-20-7	Xylene, Mixture	mg/kg	-	9.50E+02	9.50E+02	6.07E+02	6.07E+02	6.96E+02	6.96E+02

CAS			Resident	Adult 1	Resident	Child R	Resident
Number	Analyte	Units	Cancer ^h	Hazard	No Action	Hazard	No Action
7429-90-5	Aluminum	mg/kg	-	7.46E+04	7.46E+04	7.74E+03	7.74E+03
7440-36-0	Antimony (metallic)	mg/kg	-	3.33E+01	3.33E+01	3.13E+00	3.13E+00
7440-38-2	Arsenic, Inorganic	mg/kg	3.56E-01	1.03E+01	3.56E-01	2.45E+00	3.56E-01
7440-39-3	Barium	mg/kg	-	1.35E+04	1.35E+04	1.53E+03	1.53E+03
7440-41-7	Beryllium and compounds	mg/kg	1.59E+03	1.58E+02	1.58E+02	1.56E+01	1.56E+01
7440-42-8	Boron And Borates Only	mg/kg	-	1.66E+04	1.66E+04	1.56E+03	1.56E+03
7440-43-9	Cadmium (Diet)	mg/kg	2.12E+03	2.44E+00	2.44E+00	5.30E-01	5.30E-01
16065-83-1	Chromium(III), Insoluble Salts	mg/kg	-	1.00E+05	1.00E+05	1.17E+04	1.17E+04
18540-29-9	Chromium(VI)	mg/kg	3.01E-01	2.46E+02	3.01E-01	2.34E+01	3.01E-01
7440-47-3	Chromium (Total) ^a	mg/kg	-	-	-	-	-
7440-48-4	Cobalt	mg/kg	4.24E+02	2.43E+01	2.43E+01	2.34E+00	2.34E+00
7440-50-8	Copper	mg/kg	-	3.34E+03	3.34E+03	3.13E+02	3.13E+02
16984-48-8	Fluoride	mg/kg	-	3.33E+03	3.33E+03	3.13E+02	3.13E+02
7439-89-6	Iron	mg/kg	-	5.84E+04	5.84E+04	5.48E+03	5.48E+03
7439-92-1	Lead ^b	mg/kg	-	-	4.00E+02	-	4.00E+02
7439-96-5	Manganese (Non-Diet)	mg/kg	-	1.56E+03	1.56E+03	1.83E+02	1.83E+02
Various	Mercury, Inorganic Salts	mg/kg	-	2.50E+01	2.50E+01	2.35E+00	2.35E+00
7439-98-7	Molybdenum	mg/kg	-	4.17E+02	4.17E+02	3.91E+01	3.91E+01
7440-02-0	Nickel Soluble Salts	mg/kg	1.47E+04	1.48E+03	1.48E+03	1.55E+02	1.55E+02
7782-49-2	Selenium	mg/kg	-	4.17E+02	4.17E+02	3.91E+01	3.91E+01
7440-22-4	Silver	mg/kg	-	4.17E+02	4.17E+02	3.91E+01	3.91E+01
7440-28-0	Thallium (Soluble Salts)	mg/kg	-	8.34E-01	8.34E-01	7.82E-02	7.82E-02
N/A	Uranium (Insoluble Compounds) ^c	mg/kg	-	2.40E+02	2.40E+02	2.34E+01	2.34E+01
7440-61-1	Uranium (Soluble Salts) ^c	mg/kg	-	1.66E+01	1.66E+01	1.56E+00	1.56E+00
7440-62-2	Vanadium and Compounds	mg/kg	-	4.08E+02	4.08E+02	3.93E+01	3.93E+01
7440-66-6	Zinc and Compounds	mg/kg	-	2.50E+04	2.50E+04	2.35E+03	2.35E+03
83-32-9	Acenaphthene	mg/kg	-	5.66E+02	5.66E+02	1.85E+02	1.85E+02
208-96-8	Acenaphthylene ^d	mg/kg	-	5.66E+02	5.66E+02	1.85E+02	1.85E+02
107-13-1	Acrylonitrile	mg/kg	2.55E-01	1.60E+00	2.55E-01	1.57E+00	2.55E-01
120-12-7	Anthracene	mg/kg	-	2.83E+03	2.83E+03	9.23E+02	9.23E+02
71-43-2	Benzene	mg/kg	1.16E+00	1.07E+01	1.16E+00	8.17E+00	1.16E+00
117-81-7	Bis(2-ethylhexyl)phthalate ^e	mg/kg	1.49E+01	2.37E+02	1.49E+01	7.15E+01	1.49E+01
75-27-4	Bromodichloromethane	mg/kg	2.93E-01	6.67E+02	2.93E-01	6.26E+01	2.93E-01
86-74-8	Carbazole	mg/kg	1.04E+01	-	1.04E+01	-	1.04E+01
56-23-5	Carbon Tetrachloride	mg/kg	6.53E-01	1.49E+01	6.53E-01	1.04E+01	6.53E-01
67-66-3	Chloroform	mg/kg	3.16E-01	2.59E+01	3.16E-01	1.99E+01	3.16E-01
75-71-8	Dichlorodifluoromethane (Freon-12) ^e	mg/kg	-	8.76E+00	8.76E+00	8.72E+00	8.72E+00

CAS			Resident	Adult F	Resident	Child I	Resident
Number	Analyte	Units	Cancer ^h	Hazard	No Action	Hazard	No Action
75-34-3	Dichloroethane, 1,1-e	mg/kg	3.55E+00	1.08E+02	3.55E+00	1.02E+02	3.55E+00
107-06-2	Dichloroethane, 1,2-	mg/kg	4.64E-01	3.32E+00	4.64E-01	3.12E+00	4.64E-01
75-35-4	Dichloroethylene, 1,1-	mg/kg	-	2.40E+01	2.40E+01	2.27E+01	2.27E+01
540-59-0	Dichloroethylene, 1,2- (Mixed Isomers)	mg/kg	-	7.51E+02	7.51E+02	7.04E+01	7.04E+01
156-59-2	Dichloroethylene, cis-1,2-	mg/kg	-	1.04E+01	1.04E+01	6.26E+00	6.26E+00
156-60-5	Dichloroethylene, trans-1,2-	mg/kg	-	7.25E+00	7.25E+00	6.96E+00	6.96E+00
60-57-1	Dieldrin	mg/kg	1.30E-02	5.93E-01	1.30E-02	1.79E-01	1.30E-02
1746-01-6	Dioxins/Furans, Total (as TCDD) ^f	mg/kg	3.08E-06	2.07E-05	3.08E-06	4.04E-06	3.08E-06
37871-00-4	~HpCDD, 2,3,7,8-	mg/kg	3.09E-04	2.07E-03	3.09E-04	4.04E-04	3.09E-04
38998-75-3	~HpCDF, 2,3,7,8-	mg/kg	3.12E-04	2.08E-03	3.12E-04	4.04E-04	3.12E-04
34465-46-8	~HxCDD	mg/kg	3.14E-05	2.08E-04	3.14E-05	4.04E-05	3.14E-05
55684-94-1	~HxCDF, 2,3,7,8-	mg/kg	3.14E-05	2.08E-04	3.14E-05	4.04E-05	3.14E-05
3268-87-9	~OCDD	mg/kg	1.05E-02	6.93E-02	1.05E-02	1.35E-02	1.05E-02
39001-02-0	~OCDF	mg/kg	1.05E-02	6.93E-02	1.05E-02	1.35E-02	1.05E-02
36088-22-9	~PeCDD, 2,3,7,8-	mg/kg	3.14E-06	2.08E-05	3.14E-06	4.04E-06	3.14E-06
57117-41-6	~PeCDF, 1,2,3,7,8-	mg/kg	1.05E-04	6.93E-04	1.05E-04	1.35E-04	1.05E-04
57117-31-4	~PeCDF, 2,3,4,7,8-	mg/kg	1.05E-05	6.93E-05	1.05E-05	1.35E-05	1.05E-05
1746-01-6	~TCDD, 2,3,7,8-	mg/kg	3.08E-06	2.07E-05	3.08E-06	4.04E-06	3.08E-06
51207-31-9	~TCDF, 2,3,7,8-	mg/kg	3.09E-05	2.07E-04	3.09E-05	4.04E-05	3.09E-05
100-41-4	Ethylbenzene	mg/kg	5.78E+00	5.18E+02	5.78E+00	2.35E+02	5.78E+00
206-44-0	Fluoranthene	mg/kg	-	3.77E+02	3.77E+02	1.23E+02	1.23E+02
86-73-7	Fluorene	mg/kg	-	3.77E+02	3.77E+02	1.23E+02	1.23E+02
118-74-1	Hexachlorobenzene	mg/kg	2.12E-01	8.34E-01	2.12E-01	7.82E-02	2.12E-01
91-20-3	Naphthalene	mg/kg	1.04E+00	1.35E+01	1.04E+00	1.17E+01	1.04E+00
88-74-4	Nitroaniline, 2-	mg/kg	-	1.17E+02	1.17E+02	3.56E+01	3.56E+01
621-64-7	Nitroso-di-N-propylamine, N-	mg/kg	2.97E-02	-	2.97E-02	-	2.97E-02
87-86-5	Pentachlorophenol	mg/kg	2.54E-01	2.59E+01	2.54E-01	9.86E+00	2.54E-01
85-01-8	Phenanthrene ^d	mg/kg	-	5.66E+02	5.66E+02	1.85E+02	1.85E+02
1336-36-3	Polychlorinated Biphenyls, Total	mg/kg	7.88E-02	-	7.88E-02	-	7.88E-02
12674-11-2	~Aroclor 1016	mg/kg	2.26E+00	6.18E-01	6.18E-01	2.06E-01	2.06E-01
11104-28-2	~Aroclor 1221	mg/kg	7.52E-02	-	7.52E-02	-	7.52E-02
11141-16-5	~Aroclor 1232	mg/kg	7.08E-02	-	7.08E-02	-	7.08E-02
53469-21-9	~Aroclor 1242	mg/kg	7.91E-02	-	7.91E-02	-	7.91E-02
12672-29-6	~Aroclor 1248	mg/kg	7.88E-02	-	7.88E-02	-	7.88E-02
11097-69-1	~Aroclor 1254	mg/kg	7.97E-02	1.77E-01	7.97E-02	5.88E-02	5.88E-02
11096-82-5	~Aroclor 1260	mg/kg	8.03E-02	-	8.03E-02	-	8.03E-02

Table A.4a. Soil/Sediment No Action Levels for Significant COPCs at PGDP (Continued)

(Values were calculated in October 2022 and are based on best available information.)

CAS			Resident	Adult l	Resident	Child R	esident
Number	Analyte	Units	Cancer ^h	Hazard	No Action	Hazard	No Action
50-32-8	Polycyclic aromatic hydrocarbons, Total Carcinogenic ^g	mg/kg	4.78E-02	2.80E+00	4.78E-02	9.20E-01	4.78E-02
56-55-3	~Benz[a]anthracene	mg/kg	4.75E-01	-	4.75E-01	-	4.75E-01
50-32-8	~Benzo[a]pyrene	mg/kg	4.78E-02	2.80E+00	4.78E-02	9.20E-01	4.78E-02
205-99-2	~Benzo[b]fluoranthene	mg/kg	4.78E-01	-	4.78E-01	-	4.78E-01
207-08-9	~Benzo[k]fluoranthene	mg/kg	4.78E+00	-	4.78E+00	-	4.78E+00
218-01-9	~Chrysene	mg/kg	4.78E+01	-	4.78E+01	-	4.78E+01
53-70-3	~Dibenz[a,h]anthracene	mg/kg	4.78E-02	-	4.78E-02	-	4.78E-02
193-39-5	~Indeno[1,2,3-cd]pyrene	mg/kg	4.78E-01	-	4.78E-01	-	4.78E-01
129-00-0	Pyrene	mg/kg	-	2.83E+02	2.83E+02	9.23E+01	9.23E+01
127-18-4	Tetrachloroethylene	mg/kg	2.36E+01	9.61E+00	9.61E+00	8.10E+00	8.10E+00
108-88-3	Toluene ^e	mg/kg	-	1.67E+03	1.67E+03	4.89E+02	4.89E+02
76-13-1	Trichloro-1,2,2-trifluoroethane, 1,1,2- (Freon-113) ^e	mg/kg	-	6.70E+02	6.70E+02	6.69E+02	6.69E+02
71-55-6	Trichloroethane, 1,1,1-	mg/kg	-	8.55E+02	8.55E+02	8.15E+02	8.15E+02
79-00-5	Trichloroethane, 1,1,2-	mg/kg	1.15E+00	1.50E-01	1.50E-01	1.50E-01	1.50E-01
79-01-6	Trichloroethylene	mg/kg	9.43E-01	4.56E-01	4.56E-01	4.12E-01	4.12E-01
75-01-4	Vinyl Chloride	mg/kg	5.92E-02	7.73E+00	5.92E-02	5.95E+00	5.92E-02
108-38-3	Xylene, m-	mg/kg	-	5.68E+01	5.68E+01	5.50E+01	5.50E+01
95-47-6	Xylene, o-	mg/kg	-	6.70E+01	6.70E+01	6.45E+01	6.45E+01
106-42-3	Xylene, p-	mg/kg	-	5.80E+01	5.80E+01	5.61E+01	5.61E+01
1330-20-7	Xylene, Mixture	mg/kg	-	5.96E+01	5.96E+01	5.76E+01	5.76E+01

NOTES: Values are provided in these tables for significant COPCs for PGDP. Values for other COPCs can be obtained using the RAIS Chemical PRG online calculator, as modified using PGDP-specific inputs.

^a Chromium (Total) NAL should utilize Chromium III or Chromium VI Chromium VI, as appropriate.

^b Lead values should be checked prior to use to ensure they are still current.

^c Based on recommendation from EPA, NALs for uranium (soluble salts) now use the RfD and the RfC for soluble compounds of uranium derived from ATSDR (EPA 2016). NALs for uranium (insoluble compounds) use the RfD for uranium (soluble salts), which is available in IRIS; the RfC for insoluble compounds of uranium are derived from ATSDR.

^d Acenaphthylene and phenanthrene use values for Acenaphthene as a surrogate.

^e Analytes are not PGDP significant COPCs (Table 2.1), but are provided for project support.

^f Total dioxins/furans uses values for 2,3,7,8-TCDD, see screening note 9f in the Appendix A introduction, "Screening Levels," on pages A-3-A-5.

g Total carcinogenic PAHs uses values for BaP, see screening note 9d in the Appendix A introduction, "Screening Levels," on pages A-3-A-5.

h For the recreational user and the resident, ELCRs (i.e., cancer NALs) were calculated using the child/teen/adult or child/adult age-adjusted lifetime scenario, respectively.

Table A.4b. Soil/Sediment No Action Levels for Significant Radionuclide COPCs at PGDP (Values were calculated in October 2022 and are based on best available information.)

CAS				Outdoor Work	er	E	xcavation Wor	ker	I	ndustrial Work	er
Number	Analyte	Units	Peak Risk	Peak Risk	Secular	Peak Risk	Peak Risk	Secular	Peak Risk	Peak Risk	Secular
rumber			Infinite ^a	1,000 Years	Equilibrium	Infinite ^a	1,000 Years	Equilibrium	Infinite ^a	1,000 Years	Equilibrium
14596-10-2	Am-241	pCi/g	4.47E+00	4.47E+00	3.38E-01	2.20E+01	2.20E+01	1.69E+00	1.61E+01	1.61E+01	3.99E-01
10045-97-3	Cs-137	pCi/g	5.94E-01	5.94E-01	4.52E-01	2.39E+00	2.39E+00	2.26E+00	4.53E-01	4.53E-01	3.44E-01
13994-20-2	Np-237	pCi/g	4.49E-01	1.20E+00	3.66E-01	2.25E+00	5.99E+00	1.83E+00	5.00E-01	1.01E+00	4.09E-01
13981-16-3	Pu-238	pCi/g	4.20E+00	4.20E+00	7.81E-02	1.94E+01	1.94E+01	3.90E-01	2.67E+01	2.67E+01	9.63E-02
15117-48-3	Pu-239	pCi/g	3.67E+00	3.67E+00	2.87E-01	1.83E+01	1.83E+01	1.44E+00	2.32E+01	2.32E+01	3.44E-01
14119-33-6	Pu-240	pCi/g	3.68E+00	3.68E+00	8.30E-02	1.84E+01	1.84E+01	4.15E-01	2.33E+01	2.33E+01	7.49E-02
14133-76-7	Tc-99	pCi/g	3.32E+02	3.32E+02	3.32E+02	1.66E+03	1.66E+03	1.66E+03	1.95E+03	1.95E+03	1.95E+03
14269-63-7	Th-230	pCi/g	8.74E-02	2.29E-01	8.04E-02	4.37E-01	1.13E+00	4.02E-01	1.05E-01	2.79E-01	9.69E-02
7440-29-1	Th-232 ^b	pCi/g	8.57E-02	8.57E-02	8.57E-02	4.29E-01	4.29E-01	4.29E-01	7.52E-02	7.52E-02	7.52E-02
13966-29-5	U-234	pCi/g	1.35E-01	7.30E+00	7.97E-02	6.73E-01	3.64E+01	3.98E-01	1.63E-01	2.76E+01	9.67E-02
15117-96-1	U-235	pCi/g	3.12E-01	1.53E+00	3.11E-01	1.56E+00	7.66E+00	1.56E+00	3.49E-01	1.39E+00	3.49E-01
7440-61-1	U-238	pCi/g	7.83E-02	4.40E+00	7.83E-02	3.92E-01	2.20E+01	3.91E-01	9.53E-02	6.42E+00	9.53E-02

Table A.4b. Soil/Sediment No Action Levels for Significant Radionuclide COPCs at PGDP (Continued)

(Values were calculated in October 2022 and are based on best available information.)

CAS				Recreational User	a		Residentb	
Number	Analyte	Units	Peak Risk Infinite ^c	Peak Risk 1,000 Years	Secular Equilibrium	Peak Risk Infinite ^a	Peak Risk 1,000 Years	Secular Equilibrium
14596-10-2	Am-241	pCi/g	8.82E+00	8.82E+00	2.52E-01	2.10E+00	2.10E+00	4.55E-02
10045-97-3	Cs-137	pCi/g	3.13E-01	3.13E-01	2.36E-01	5.24E-02	5.24E-02	3.95E-02
13994-20-2	Np-237	pCi/g	3.18E-01	6.76E-01	2.60E-01	5.69E-02	1.15E-01	4.66E-02
13981-16-3	Pu-238	pCi/g	1.30E+01	1.30E+02	6.13E-02	4.28E+00	4.28E+00	1.10E-02
15117-48-3	Pu-239	pCi/g	1.16E+01	1.16E+02	2.21E-01	3.79E+00	3.79E+00	3.97E-02
14119-33-6	Pu-240	pCi/g	1.16E+01	1.16E+01	4.95E-02	3.81E+00	3.81E+00	8.54E-03
14133-76-7	Tc-99	pCi/g	3.49E+03	3.49E+02	3.49E+03	1.12E+02	1.12E+02	1.12E+02
14269-63-7	Th-230	pCi/g	6.73E-02	1.78E-01	6.19E-02	1.21E-02	3.20E-02	1.11E-02
7440-29-1	Th-232 ^d	pCi/g	4.99E-02	4.99E-02	4.99E-02	8.57E-03	8.57E-03	8.57E-03
13966-29-5	U-234	pCi/g	1.04E-01	1.19E+01	6.17E-02	1.87E-02	3.09E+00	1.11E-02
15117-96-1	U-235	pCi/g	2.25E-01	9.20E-01	2.25E-01	4.01E-02	1.59E-01	4.01E-02
7440-61-1	U-238	pCi/g	6.07E-02	3.67E+00	6.07E-02	1.09E-02	7.09E+00	1.09E-02

NOTES: Values are provided in these tables for significant COPCs for PGDP. Values for other radionuclides can be obtained using the EPA Radionuclide PRG calculator, as modified using PGDP-specific inputs. The resident NALs do not include the consumption of produce pathway included in the EPA Radionuclide PRG calculator.

Radionuclide NALs are based on the cancer endpoint using a target ELCR of 1E-6. The soil and sediment radionuclide NALs consider the ingestion, inhalation, and external exposure pathways.

^a For the recreational user, the radionuclide ALs were calculated using the child/adult age-adjusted scenario (i.e., 26-year exposure duration, with 6 years as a child and 20 years as an adult). The EPA Radiological PRG calculator only allows exposure parameter entries for two life stages (i.e., child and adult), rather than the four life stage groups provided in the EPA and RAIS chemical online calculators. Thus, the EPA Radiological PRG calculator does not allow entry of the teen exposure parameters shown in Table B.5.

^b For the resident, the radionuclide NALs were calculated using the child/adult age-adjusted lifetime scenario (i.e., lifetime exposure).

^c The time frame for calculating peak risk over an infinite period in the EPA Radionuclide PRG calculator is 1 trillion years. NASA most recently has estimated the age of the universe as being 13.7 billion years, with an uncertainty of only 200 million years (https://imagine.gsfc.nasa.gov).

^d Analyte is not a PGDP significant COPC (Table 2.1), but it is provided for project support.

Table A.5a. Groundwater No Action Levels and Primary MCLs for Significant COPCs at PGDP (Values were calculated in October 2022 and are based on best available information.)

CAS			Resident	Adult F	Resident	Child F	Resident	Primary
Number	Analyte	Units	Cancera		No Action		No Action	MCLs ^j
7429-90-5	Aluminum	μg/L	-	3.32E+03	3.32E+03	2.00E+03	2.00E+03	-
7440-36-0	Antimony (metallic)	μg/L	-	1.29E+00		7.79E-01	7.79E-01	6.00E+00
7440-38-2	Arsenic, Inorganic	μg/L	5.17E-02	9.96E-01	5.17E-02	5.99E-01	5.17E-02	1.00E+01
7440-39-3	Barium	μg/L	-	6.18E+02	6.18E+02	3.77E+02	3.77E+02	2.00E+03
7440-41-7	Beryllium and compounds	μg/L	ı	3.71E+00	3.71E+00	2.46E+00	2.46E+00	4.00E+00
7440-42-8	Boron And Borates Only	μg/L	ı	6.64E+02	6.64E+02	3.99E+02	3.99E+02	-
7440-43-9	Cadmium (Water)	μg/L	-	3.00E-01	3.00E-01	1.84E-01	1.84E-01	5.00E+00
16065-83-1	Chromium(III), Insoluble Salts	μg/L	-	3.50E+03			2.25E+03	-
18540-29-9	Chromium(VI)	μg/L	3.50E-02	6.92E+00	3.50E-02	4.45E+00	3.50E-02	-
7440-47-3	Chromium (Total) ^b	μg/L	-	-	-	-	-	1.00E+02
7440-48-4	Cobalt	μg/L	1	9.99E-01	9.99E-01	6.01E-01	6.01E-01	-
7440-50-8	Copper	μg/L	1	1.33E+02	1.33E+02	7.99E+01	7.99E+01	1.30E+03
16984-48-8	Fluoride	μg/L	-	1.33E+02	1.33E+02	7.99E+01	7.99E+01	4.00E+03
7439-89-6	Iron	μg/L	ı	2.32E+03	2.32E+03	1.40E+03	1.40E+03	-
7439-92-1	Lead ^c	μg/L	ı	-	1.50E+01	-	1.50E+01	1.50E+01
7439-96-5	Manganese (Non-Diet)	μg/L	ı	7.03E+01	7.03E+01	4.34E+01	4.34E+01	-
Various	Mercury, Inorganic Salts	μg/L	ı	9.27E-01	9.27E-01	5.66E-01	5.66E-01	2.00E+00d
7439-98-7	Molybdenum	μg/L	ı	1.66E+01	1.66E+01	9.98E+00	9.98E+00	-
7440-02-0	Nickel Soluble Salts	μg/L	ı	6.49E+01	6.49E+01	3.92E+01	3.92E+01	-
7782-49-2	Selenium	μg/L	ı	1.66E+01	1.66E+01	9.98E+00	9.98E+00	5.00E+01
7440-22-4	Silver	μg/L	-	1.54E+01	1.54E+01	9.41E+00	9.41E+00	-
7440-28-0	Thallium (Soluble Salts)	μg/L	-	3.32E-02	3.32E-02	2.00E-02	2.00E-02	2.00E+00
N/A	Uranium (Insoluble Compounds) ^e	μg/L	-	9.96E+00	9.96E+00	5.99E+00	5.99E+00	3.00E+01
7440-61-1	Uranium (Soluble Salts) ^e	μg/L	-	6.64E-01	6.64E-01	3.99E-01	3.99E-01	3.00E+01
7440-62-2	Vanadium and Compounds	μg/L	-	1.38E+01	1.38E+01	8.64E+00	8.64E+00	-
7440-66-6	Zinc and Compounds	μg/L	-	9.98E+02	9.98E+02	6.00E+02	6.00E+02	-
83-32-9	Acenaphthene	μg/L	-	8.41E+01	8.41E+01	5.35E+01	5.35E+01	-
208-96-8	Acenaphthylene ^f	μg/L	-	8.41E+01	8.41E+01	5.35E+01	5.35E+01	-
107-13-1	Acrylonitrile	μg/L	5.23E-02	4.12E-01	5.23E-02	4.09E-01	5.23E-02	-
120-12-7	Anthracene	μg/L	-	2.74E+02	2.74E+02	1.77E+02	1.77E+02	-
71-43-2	Benzene	μg/L	4.55E-01	4.07E+00	4.55E-01	3.32E+00	4.55E-01	5.00E+00
117-81-7	Bis(2-ethylhexyl)phthalateg	μg/L	5.56E+00	6.67E+01	5.56E+00	4.01E+01	5.56E+00	6.00E+00
75-27-4	Bromodichloromethane	μg/L	1.34E-01	2.50E+01	1.34E-01	1.51E+01		8.00E+01k
86-74-8	Carbazole	μg/L	2.03E+00	-	2.03E+00	-	2.03E+00	-
56-23-5	Carbon Tetrachloride	μg/L	4.55E-01	7.02E+00	4.55E-01	4.95E+00	4.55E-01	5.00E+00
67-66-3	Chloroform	μg/L	2.21E-01	1.22E+01	2.21E-01	9.72E+00	2.21E-01	8.00E+01k
75-71-8	Dichlorodifluoromethane (Freon-12) ^g	μg/L	-	2.02E+01	2.02E+01	1.97E+01	1.97E+01	-

Table A.5a. Groundwater No Action Levels and Primary MCLs for Significant COPCs at PGDP (Continued) (Values were calculated in October 2022 and are based on best available information.)

CAS			Resident		Resident		Resident	Primary
Number	Analyte	Units	Cancer ^a		No Action	Hazard	No Action	MCLs ^j
75-34-3	Dichloroethane, 1,1-g	μg/L	2.75E+00	8.93E+01	2.75E+00	8.16E+01	2.75E+00	-
107-06-2	Dichloroethane, 1,2-	μg/L	1.71E-01	1.36E+00		1.30E+00		5.00E+00
75-35-4	Dichloroethylene, 1,1-	μg/L	ı	3.25E+01	3.25E+01	2.85E+01	2.85E+01	7.00E+00
540-59-0	Dichloroethylene, 1,2- (Mixed Isomers)	μg/L	-	2.68E+01		1.63E+01	1.63E+01	-
156-59-2	Dichloroethylene, cis-1,2-	μg/L	-	2.52E+00	2.52E+00	3.47E+00	3.47E+00	7.00E+01
156-60-5	Dichloroethylene, trans-1,2-	μg/L	-	7.32E+00		6.78E+00	6.78E+00	1.00E+02
60-57-1	Dieldrin	μg/L	1.75E-03	5.95E-02	1.75E-03	3.81E-02	1.75E-03	-
1746-01-6	Dioxins/Furans, Total (as TCDD)h	μg/L	1.19E-07	1.83E-06	1.19E-07	1.20E-06	1.19E-07	3.00E-05
37871-00-4	~HpCDD, 2,3,7,8-	μg/L	1.19E-05	1.83E-04	1.19E-05	1.20E-04	1.19E-05	•
38998-75-3	~HpCDF, 2,3,7,8-	μg/L	1.19E-05	1.83E-04	1.19E-05	1.20E-04	1.19E-05	ı
34465-46-8	~HxCDD	μg/L	5.99E-06	2.34E-05	5.99E-06	1.40E-05	5.99E-06	-
55684-94-1	~HxCDF, 2,3,7,8-	μg/L	5.99E-06	2.34E-05	5.99E-06	1.40E-05	5.99E-06	-
3268-87-9	~OCDD	μg/L	2.00E-03	7.79E-03	2.00E-03	4.68E-03	2.00E-03	-
39001-02-0	~OCDF	μg/L	2.00E-03	7.79E-03	2.00E-03	4.68E-03	2.00E-03	-
36088-22-9	~PeCDD, 2,3,7,8-	μg/L	5.99E-07	2.34E-06	5.99E-07	1.40E-06	5.99E-07	-
57117-41-6	~PeCDF, 1,2,3,7,8-	μg/L	2.00E-05	7.79E-05	2.00E-05	4.68E-05	2.00E-05	-
57117-31-4	~PeCDF, 2,3,4,7,8-	μg/L	2.00E-06	7.79E-06	2.00E-06	4.68E-06	2.00E-06	-
1746-01-6	~TCDD, 2,3,7,8-	μg/L	1.19E-07	1.83E-06	1.19E-07	1.20E-06	1.19E-07	3.00E-05
51207-31-9	~TCDF, 2,3,7,8-	μg/L	1.19E-06	1.83E-05	1.19E-06	1.20E-05	1.19E-06	-
100-41-4	Ethylbenzene	μg/L	1.50E+00	7.01E+01	1.50E+00	5.00E+01	1.50E+00	7.00E+02
206-44-0	Fluoranthene	μg/L	-	1.33E+02	1.33E+02	8.02E+01	8.02E+01	-
86-73-7	Fluorene	μg/L	-	4.59E+01	4.59E+01	2.94E+01	2.94E+01	-
118-74-1	Hexachlorobenzene	μg/L	9.76E-03	3.34E-02	9.76E-03	2.01E-02	9.76E-03	1.00E+00
91-20-3	Naphthalene	μg/L	1.17E-01	6.16E-01	1.17E-01	6.11E-01	1.17E-01	-
88-74-4	Nitroaniline, 2-	μg/L	-	3.13E+01	3.13E+01	1.89E+01	1.89E+01	-
621-64-7	Nitroso-di-N-propylamine, N-	μg/L	1.08E-02	-	1.08E-02	-	1.08E-02	-
87-86-5	Pentachlorophenol	μg/L	4.13E-02	3.49E+00	4.13E-02	2.27E+00	4.13E-02	1.00E+00
85-01-8	Phenanthrene ^f	μg/L	-	8.41E+01	8.41E+01	5.35E+01	5.35E+01	-
1336-36-3	Polychlorinated Biphenyls, Total	μg/L	4.36E-02	-	4.36E-02	-	4.36E-02	5.00E-01
12674-11-2	~Aroclor 1016	μg/L	2.24E-01	2.34E-01	2.24E-01	1.40E-01	1.40E-01	-
11104-28-2	~Aroclor 1221	μg/L	4.71E-03	-	4.71E-03	-	4.71E-03	-
11141-16-5	~Aroclor 1232	μg/L	4.71E-03	-	4.71E-03	-	4.71E-03	-
53469-21-9	~Aroclor 1242	μg/L	7.85E-03	-	7.85E-03	-	7.85E-03	-
12672-29-6	~Aroclor 1248	μg/L	7.85E-03	-	7.85E-03	-	7.85E-03	-
11097-69-1	~Aroclor 1254	μg/L	7.85E-03	6.67E-02	7.85E-03	4.01E-02	7.85E-03	-
11096-82-5	~Aroclor 1260	μg/L	7.85E-03	-	7.85E-03	-	7.85E-03	-

Table A.5a. Groundwater No Action Levels and Primary MCLs for Significant COPCs at PGDP (Continued)

(Values were calculated in October 2022 and are based on best available information.)

CAS			Resident	Adult 1	Resident	Child F	Resident	Primary
Number	Analyte	Units	Cancera	Hazard	No Action	Hazard	No Action	MCLsj
50-32-8	Polycyclic aromatic hydrocarbons, Total Carcinogenic ⁱ	μg/L	2.51E-02	1.00E+00	2.51E-02	6.02E-01	2.51E-02	2.00E-01
56-55-3	~Benz[a]anthracene	μg/L	2.98E-02	-	2.98E-02	-	2.98E-02	-
50-32-8	~Benzo[a]pyrene	μg/L	2.51E-02	1.00E+00	2.51E-02	6.02E-01	2.51E-02	2.00E-01
205-99-2	~Benzo[b]fluoranthene	μg/L	2.51E-01	-	2.51E-01	-	2.51E-01	-
207-08-9	~Benzo[k]fluoranthene	μg/L	2.51E+00	-	2.51E+00	-	2.51E+00	-
218-01-9	~Chrysene	μg/L	2.51E+01	-	2.51E+01	-	2.51E+01	-
53-70-3	~Dibenz[a,h]anthracene	μg/L	2.51E-02	-	2.51E-02	-	2.51E-02	-
193-39-5	~Indeno[1,2,3-cd]pyrene	μg/L	2.51E-01	-	2.51E-01	-	2.51E-01	-
129-00-0	Pyrene	μg/L	-	1.86E+01	1.86E+01	1.21E+01	1.21E+01	-
127-18-4	Tetrachloroethylene	μg/L	1.13E+01	5.03E+00	5.03E+00	4.06E+00	4.06E+00	5.00E+00
108-88-3	Tolueneg	μg/L	-	1.68E+02	1.68E+02	1.10E+02	1.10E+02	1.00E+03
76-13-1	Trichloro-1,2,2-trifluoroethane, 1,1,2- (Freon-113) ^g	μg/L	-	1.03E+03	1.03E+03	1.02E+03	1.02E+03	-
71-55-6	Trichloroethane, 1,1,1-	μg/L	-	8.81E+02	8.81E+02	8.01E+02	8.01E+02	2.00E+02
79-00-5	Trichloroethane, 1,1,2-	μg/L	2.75E-01	4.16E-02	4.16E-02	4.15E-02	4.15E-02	5.00E+00
79-01-6	Trichloroethylene	μg/L	4.94E-01	3.23E-01	3.23E-01	2.83E-01	2.83E-01	5.00E+00
75-01-4	Vinyl Chloride	μg/L	1.88E-02	5.97E+00	1.88E-02	4.21E+00	1.88E-02	2.00E+00
108-38-3	Xylene, m-	μg/L	-	1.98E+01	1.98E+01	1.93E+01	1.93E+01	-
95-47-6	Xylene, o-	μg/L	-	1.99E+01	1.99E+01	1.93E+01	1.93E+01	-
106-42-3	Xylene, p-	μg/L	-	1.99E+01	1.99E+01	1.93E+01	1.93E+01	-
1330-20-7	Xylene, Mixture	μg/L	-		1.99E+01	1.93E+01	1.93E+01	1.00E+04

NOTES: Values are provided in these tables for significant COPCs for PGDP. Values for other COPCs can be obtained using the RAIS Chemical PRG online calculator, as modified using PGDP-specific inputs. NALs are not adjusted for solubility limits.

^a For the resident, ELCRs (i.e., cancer NALs) were calculated using the child/adult age-adjusted lifetime scenario (i.e., lifetime exposure).

^b Chromium (Total) NAL should utilize Chromium III or Chromium VI, as appropriate.

^c Lead values should be checked prior to use to ensure they are still current.

^d MCL is for mercury (elemental).

^e Based on recommendation from EPA, NALs for uranium (soluble salts) now use the RfD and the RfC for soluble compounds of uranium derived from ATSDR (EPA 2016). NALs for uranium (insoluble compounds) use the RfD for uranium (soluble salts), which is available in IRIS: the RfC for insoluble compounds of uranium are derived from ATSDR.

^f Acenaphthylene and phenanthrene use values for acenaphthene as a surrogate.

g Analytes are not PGDP significant COPCs (Table 2.1), but are provided for project support.

h Total dioxins/furans uses values for 2,3,7,8-TCDD, see screening note 9f in the Appendix A introduction, "Screening Levels," on pages A-3-A-5.

Total carcinogenic PAHs uses values for BaP, see screening note 9d in the Appendix A introduction, "Screening Levels," on pages A-3-A-5.

^j Radionuclides use only the ingestion risk values.

^j Accessed at https://www.epa.gov/sites/default/files/2016-06/documents/npwdr_complete_table.pdf, accessed October 27, 2022.

^k MCL is for the sum of the concentrations for trihalomethanes.

Table A.5b. Groundwater No Action Levels and Primary MCLs for Significant COPCs at PGDP

(Values were calculated in October 2022 and are based on best available information.)

CAC				Residenta		D:
CAS Number	Analyte	Units	Peak Risk Infinite ^b	Peak Risk 1,000 Years	Secular Equilibrium	Primary MCLs
14596-10-2	Am-241	pCi/L	5.15E-01	5.15E-01	6.77E-02	1.50E+01°
10045-97-3	Cs-137	pCi/L	2.28E+00	2.28E+00	1.71E+00	d
13994-20-2	Np-237	pCi/L	9.64E-02	7.61E-01	7.83E-02	1.50E+01e
13981-16-3	Pu-238	pCi/L	4.40E-01	4.40E-01	1.56E-02	1.50E+01 ^f
15117-48-3	Pu-239	pCi/L	3.87E-01	3.87E-01	6.03E-02	1.50E+01 ^f
14119-33-6	Pu-240	pCi/L	3.87E-01	3.87E-01	3.18E-02	1.50E+01 ^f
14133-76-7	Tc-99	pCi/L	1.90E+01	1.90E+01	1.90E+01	g
14269-63-7	Th-230	pCi/L	1.81E-02	4.65E-02	1.66E-02	1.50E+01 ^h
7440-29-1	Th-232i	pCi/L	3.63E-02	3.63E-02	3.63E-02	1.50E+01 ^h
13966-29-5	U-234	pCi/L	2.74E-02	6.86E-01	1.62E-02	j
15117-96-1	U-235	pCi/L	7.15E-02	6.14E-01	7.14E-02	j
7440-61-1	U-238	pCi/L	1.58E-02	5.99E-01	1.58E-02	j

NOTES: Values are provided in these tables for significant radionuclide COPCs for PGDP. Values for other radionuclides can be obtained using the EPA Radionuclide PRG calculator, as modified using PGDP-specific inputs.

Radionuclide NALs are based on the cancer endpoint using a target ELCR of 1.0E-6. The groundwater radionuclide NALs are based on the ingestion pathway only.

^a For the resident, cancer NALs were calculated using the child/adult age-adjusted lifetime scenario (i.e., lifetime exposure).

^b The time frame for calculating peak risk over an infinite period in the EPA Radionuclide PRG calculator is 1 trillion years. NASA most recently has estimated the age of the universe as being 13.7 billion years, with an uncertainty of only 200 million years (https://imagine.gsfc.nasa.gov).

^c Additional information regarding Am-241 can be found in "EPA Facts about Americium-241," dated July 2002, at the following link: https://semspub.epa.gov/work/HQ/176297.pdf; accessed October 21, 2022.

^d The EPA MCL for Cs-137 is 4 mrem/year. The value derived by the EPA from the 4 mrem/year MCL for Cs-137 is 200 pCi/L (see "Limits for Beta Particles and Photon Emitters at 4 millrems/year" found on https://www.epa.gov/sites/production/files/2015-09/documents/guide radionuclides table-betaphotonemitters.pdf; accessed October 5, 2022.

e"Maximum Contaminant Level's in EPA's PRG and Dose Compliance Concentration Calculators," revised September 2015, found on https://epa-prgs.ornl.gov/radionuclides/MCLs_2015.pdf; accessed October 5, 2022.

f Additional information regarding plutonium can be found at the following link: http://www.epa.gov/radiation/radionuclides.

EThe value derived by EPA from the 4 mrem/year MCL for Tc-99 is 900 pCi/L, (see https://www.epa.gov/sites/production/files/2015-09/documents/guide_radionuclides_table-betaphotonemitters.pdf, accessed October 5, 2022. An alternate value derived by EPA from the 4 mrem/year MCL is 3,790 pCi/L and was proposed in the July 18, 1991, Federal Register. See Table A.9 for Tc-99 dose-based groundwater screening levels resulting in a 4 mrem/year dose based on more recent dosimetry.

h Additional information regarding thorium can be found at the following link: http://www.epa.gov/radiation/radionuclides.

¹ Analyte is not a PGDP significant COPC (Table 2.1), but it is provided for project support.

¹ The uranium MCL is 30 μg/L and can be assumed to be at a 1:1 ratio for pCi/L (or 30 pCi/L). The MCL also can be converted to 20 pCi/L for total uranium using a uranium activity expected at PGDP. Isotopic uranium values derived from this conversion are 10.24 pCi/L for U-234, 0.466 pCi/L for U-235, and 9.99 pCi/L for U-238, assuming natural occurring uranium at 0.725% U-235 and the following ratios:

U-234/U-235 ranges 21–22 obtained from conversion approximately 21.9

U-235/U-238 ranges 0.04–0.05 obtained from conversion approximately 0.045

CAS			Oı	utdoor Wor	ker ^a	Exc	avation Wo	rker ^a	Industrial Worker ^a		
Number	Analyte	Units	Cancer	Hazard	No Action	Cancer	Hazard	No Action	Cancer	Hazard	No Action
7429-90-5	Aluminum	μg/L	-	5.17E+06	5.17E+06	-	5.17E+06	5.17E+06	-	1.27E+06	1.27E+06
7440-36-0	Antimony (metallic)	μg/L	-	3.10E+02	3.10E+02	-	3.10E+02	3.10E+02	-	7.64E+01	7.64E+01
7440-38-2	Arsenic, Inorganic	μg/L	9.66E+01	1.55E+03	9.66E+01	4.83E+02	1.55E+03	4.83E+02	2.38E+01	3.82E+02	2.38E+01
7440-39-3	Barium	μg/L	-	7.24E+04	7.24E+04	-	7.24E+04	7.24E+04	-	1.78E+04	1.78E+04
7440-41-7	Beryllium and compounds	μg/L	-	7.24E+01	7.24E+01	-	7.24E+01	7.24E+01	1	1.78E+01	1.78E+01
7440-42-8	Boron And Borates Only	μg/L	-	1.03E+06	1.03E+06	-	1.03E+06	1.03E+06	1	2.55E+05	2.55E+05
7440-43-9	Cadmium (Water)	μg/L	-	2.59E+01	2.59E+01	-	2.59E+01	2.59E+01	-	6.37E+00	6.37E+00
16065-83-1	Chromium(III), Insoluble Salts	μg/L	-	1.01E+05	1.01E+05	-	1.01E+05	1.01E+05	-	2.48E+04	2.48E+04
18540-29-9	Chromium(VI)	μg/L	3.62E+00	1.94E+02	3.62E+00	1.81E+01	1.94E+02	1.81E+01	8.92E-01	4.78E+01	8.92E-01
7440-47-3	Chromium (Total) ^b	μg/L	-	-	-	-	-	-	-	-	-
7440-48-4	Cobalt	μg/L	-	3.88E+03	3.88E+03	-	3.88E+03	3.88E+03	-	9.55E+02	9.55E+02
7440-50-8	Copper	μg/L	-	2.07E+05	2.07E+05	-	2.07E+05	2.07E+05	-	5.09E+04	5.09E+04
16984-48-8	Fluoride	μg/L	-	2.07E+05	2.07E+05	-	2.07E+05	2.07E+05	ı	5.09E+04	5.09E+04
7439-89-6	Iron	μg/L	-	3.62E+06	3.62E+06	-	3.62E+06	3.62E+06	-	8.92E+05	8.92E+05
7439-92-1	Lead ^c	μg/L	-	-	1.50E+01	-	-	1.50E+01	-	-	1.50E+01
7439-96-5	Manganese (Non-Diet)	μg/L	-	4.97E+03	4.97E+03	-	4.97E+03	4.97E+03	-	1.22E+03	1.22E+03
Various	Mercury, Inorganic Salts	μg/L	-	1.09E+02	1.09E+02	-	1.09E+02	1.09E+02	-	2.67E+01	2.67E+01
7439-98-7	Molybdenum	μg/L	-	2.59E+04	2.59E+04	-	2.59E+04	2.59E+04	-	6.37E+03	6.37E+03
7440-02-0	Nickel Soluble Salts	μg/L	-	2.07E+04	2.07E+04	-	2.07E+04	2.07E+04	1	5.09E+03	5.09E+03
7782-49-2	Selenium	μg/L	-	2.59E+04	2.59E+04	-	2.59E+04	2.59E+04	1	6.37E+03	6.37E+03
7440-22-4	Silver	μg/L	-	1.72E+03	1.72E+03	-	1.72E+03	1.72E+03	1	4.25E+02	4.25E+02
7440-28-0	Thallium (Soluble Salts)	μg/L	-	5.17E+01	5.17E+01	-	5.17E+01	5.17E+01	1	1.27E+01	1.27E+01
N/A	Uranium (Insoluble Compounds) ^d	μg/L	-	1.55E+04	1.55E+04	-	1.55E+04	1.55E+04	1	3.82E+03	3.82E+03-
7440-61-1	Uranium (Soluble Salts) ^d	μg/L	-	1.03E+03	1.03E+03	-	1.03E+03	1.03E+03	1	2.55E+02	2.55E+02
7440-62-2	Vanadium and Compounds	μg/L	-	6.78E+02	6.78E+02	-	6.78E+02	6.78E+02	1	1.67E+02	1.67E+02
7440-66-6	Zinc and Compounds	μg/L	-	2.59E+06	2.59E+06	-	2.59E+06	2.59E+06	ı	6.37E+05	6.37E+05
83-32-9	Acenaphthene	μg/L	-	3.71E+03	3.71E+03	-	3.71E+03	3.71E+03	ı	5.84E+02	5.84E+02
208-96-8	Acenaphthylene ^e	μg/L	-	3.71E+03	3.71E+03	-	3.71E+03	3.71E+03	ı	5.84E+02	5.84E+02
107-13-1	Acrylonitrile	μg/L	2.20E+02	4.25E+04	2.20E+02	1.10E+03	4.25E+04	1.10E+03	4.92E+01	9.48E+03	4.92E+01
120-12-7	Anthracene	μg/L	-	1.10E+04	1.10E+04	-	1.10E+04	1.10E+04	-	1.53E+03	1.53E+03
71-43-2	Benzene	μg/L	1.72E+02	1.35E+03	1.72E+02	8.60E+02	1.35E+03	8.60E+02	3.67E+01	2.89E+02	3.67E+01
117-81-7	Bis(2-ethylhexyl)phthalate ^f	μg/L	-	-	_	-	-	-	-	-	-
75-27-4	Bromodichloromethane	μg/L	4.83E+02	8.56E+03	4.83E+02	2.42E+03	8.56E+03	2.42E+03	8.61E+01	1.52E+03	8.61E+01
86-74-8	Carbazole	μg/L	1.26E+02	-	1.26E+02	6.29E+02	-	6.29E+02	2.00E+01	-	2.00E+01
56-23-5	Carbon Tetrachloride	μg/L	1.12E+02	1.12E+03	1.12E+02	5.60E+02	1.12E+03	5.60E+02	2.00E+01	2.00E+02	2.00E+01

CAS			Ou	tdoor Work	xer ^a	Exca	avation Wo	rker ^a	Industrial Worker ^a		
Number	Analyte	Units	Cancer	Hazard	No Action	Cancer	Hazard	No Action	Cancer	Hazard	No Action
67-66-3	Chloroform	μg/L	6.23E+02	6.90E+03	6.23E+02	3.12E+03	6.90E+03	3.12E+03	1.24E+02	1.37E+03	1.24E+02
75-71-8	Dichlorodifluoromethane (Freon-12) ^f	μg/L	-	1.06E+05	1.06E+05	-	1.06E+05	1.06E+05	-	2.09E+04	2.09E+04
75-34-3	Dichloroethane, 1,1-f	μg/L	3.51E+03	1.43E+05	3.51E+03	1.76E+04	1.43E+05	1.76E+04	7.29E+02	2.97E+04	7.29E+02
107-06-2	Dichloroethane, 1,2-	μg/L	3.51E+02	6.85E+03	3.51E+02	1.76E+03	6.85E+03	1.76E+03	7.30E+01	1.42E+03	7.30E+01
75-35-4	Dichloroethylene, 1,1-	μg/L	-	2.10E+04	2.10E+04	-	2.10E+04	2.10E+04	-	4.35E+03	4.35E+03
540-59-0	Dichloroethylene, 1,2- (Mixed Isomers)	μg/L	-	4.01E+03	4.01E+03	-	4.01E+03	4.01E+03	-	8.31E+02	8.31E+02
156-59-2	Dichloroethylene, cis-1,2-	μg/L	-	8.91E+02	8.91E+02	-	8.91E+02	8.91E+02	-	1.85E+02	1.85E+02
156-60-5	Dichloroethylene, trans-1,2-	μg/L	-	8.91E+03	8.91E+03	-	8.91E+03	8.91E+03	-	1.85E+03	1.85E+03
60-57-1	Dieldrin	μg/L	9.40E-02	2.69E+00	9.40E-02	4.70E-01	2.69E+00	4.70E-01	1.32E-02	3.77E-01	1.32E-02
1746-01-6	Dioxins/Furans, Total (as TCDD)g	μg/L	-	-	-	-	-	-	-	-	-
37871-00-4	~HpCDD, 2,3,7,8-	μg/L	-	-	-	-	-	-	-	-	-
38998-75-3	~HpCDF, 2,3,7,8-	μg/L	-	-	-	-	-	-	-	-	-
34465-46-8	~HxCDD	μg/L	-	-	-	-	-	-	-	-	-
55684-94-1	~HxCDF, 2,3,7,8-	μg/L	-	-	-	-	-	-	-	-	-
3268-87-9	~OCDD	μg/L	-	-	-	-	-	-	-	-	-
39001-02-0	~OCDF	μg/L	-	-	-	-	-	-	-	-	-
36088-22-9	~PeCDD, 2,3,7,8-	μg/L	-	-	-	-	-	-	-	-	-
57117-41-6	~PeCDF, 1,2,3,7,8-	μg/L	-	-	-	-	-	-	-	-	-
57117-31-4	~PeCDF, 2,3,4,7,8-	μg/L	-	-	-	-	-	-	-	-	-
1746-01-6	~TCDD, 2,3,7,8-	μg/L	-	-	-	-	-	-	-	-	-
51207-31-9	~TCDF, 2,3,7,8-	μg/L	-	-	-	-	-	-	-	-	-
100-41-4	Ethylbenzene	μg/L	2.78E+02	5.47E+03	2.78E+02	1.39E+03	5.47E+03	1.39E+03	5.41E+01	1.06E+03	5.41E+01
206-44-0	Fluoranthene	μg/L	-	-	-	-	-	-	-	-	-
86-73-7	Fluorene	μg/L	-	1.92E+03	1.92E+03	-	1.92E+03	1.92E+03	-	2.78E+02	2.78E+02
118-74-1	Hexachlorobenzene	μg/L	-	-	-	-	-	-	-	-	-
91-20-3	Naphthalene	μg/L	2.60E+01	2.23E+03	2.60E+01	1.30E+02	2.23E+03	1.30E+02	4.77E+00	4.09E+02	4.77E+00
88-74-4	Nitroaniline, 2-	μg/L	-	1.02E+04	1.02E+04	-	1.02E+04	1.02E+04	-	1.94E+03	1.94E+03
621-64-7	Nitroso-di-N-propylamine, N-	μg/L	7.85E+00	-	7.85E+00	3.92E+01	-	3.92E+01	1.53E+00	-	1.53E+00
87-86-5	Pentachlorophenol	μg/L	1.80E+00	1.28E+02	1.80E+00	8.98E+00	1.28E+02	8.98E+00	2.52E-01	1.80E+01	2.52E-01
85-01-8	Phenanthrene ^e	μg/L	-	3.71E+03	3.71E+03	-	3.71E+03	3.71E+03	-	5.84E+02	5.84E+02
1336-36-3	Polychlorinated Biphenyls, Total	μg/L	-	-	-	-	-	-	-	-	-
12674-11-2	~Aroclor 1016	μg/L	-	-	-	-	-	-	-	-	-
11104-28-2	~Aroclor 1221	μg/L	4.16E-01	-	4.16E-01	2.08E+00	-	2.08E+00	5.66E-02	-	5.66E-02
11141-16-5	~Aroclor 1232	μg/L	4.16E-01	-	4.16E-01	2.08E+00	-	2.08E+00	5.66E-02	-	5.66E-02
53469-21-9	~Aroclor 1242	μg/L	-	•	-	-	-	1	-	-	-
12672-29-6	~Aroclor 1248	μg/L	-	-	-	-	-	-	-	-	-
11097-69-1	~Aroclor 1254	μg/L	-	-	-	-	-	-	-	-	-
11096-82-5	~Aroclor 1260	μg/L	-	-	-	-	-	-	-	-	-

CAS			Ou	tdoor Work	:er ^a	Exca	avation Wor	rker ^a	Industrial Worker ^a		
Number	Analyte	Units	Cancer	Hazard	No Action	Cancer	Hazard	No Action	Cancer	Hazard	No Action
50-32-8	Polycyclic aromatic hydrocarbons, Total Carcinogenic ^h	μg/L	-	-	-	1	-	-	1	1	-
56-55-3	~Benz[a]anthracene	μg/L	-	-	-	1	-	-	1	ı	-
50-32-8	~Benzo[a]pyrene	μg/L	-	-	-	1	-	-	1	ı	-
205-99-2	~Benzo[b]fluoranthene	μg/L	-	-	-	1	-	-	1	ı	-
207-08-9	~Benzo[k]fluoranthene	μg/L	-	-	-	1	-	-	1	ı	-
218-01-9	~Chrysene	μg/L	-	-	-	ı	-	-	ı	ı	-
53-70-3	~Dibenz[a,h]anthracene	μg/L	-	-	-	ı	-	-	ı	ı	-
193-39-5	~Indeno[1,2,3-cd]pyrene	μg/L	-	-	-	ı	-	-	ı	ı	-
129-00-0	Pyrene	μg/L	-	6.91E+02	6.91E+02	-	6.91E+02	6.91E+02	-	9.28E+01	9.28E+01
127-18-4	Tetrachloroethylene	μg/L	1.85E+03	8.32E+02	8.32E+02	9.24E+03	8.32E+02	8.32E+02	3.07E+02	1.38E+02	1.38E+02
108-88-3	Toluene ^f	μg/L	-	1.34E+04	1.34E+04	-	1.34E+04	1.34E+04	-	2.75E+03	2.75E+03
76-13-1	Trichloro-1,2,2-trifluoroethane, 1,1,2- (Freon-113) ^f	μg/L	-	7.17E+06	7.17E+06	-	7.17E+06	7.17E+06	-	1.17E+06	1.17E+06
71-55-6	Trichloroethane, 1,1,1-	μg/L	-	7.45E+05	7.45E+05	1	7.45E+05	7.45E+05	1	1.42E+05	1.42E+05
79-00-5	Trichloroethane, 1,1,2-	μg/L	4.47E+02	3.64E+03	4.47E+02	2.24E+03	3.64E+03	2.24E+03	8.62E+01	7.02E+02	8.62E+01
79-01-6	Trichloroethylene	μg/L	2.45E+02	2.02E+02	2.02E+02	1.22E+03	2.02E+02	2.02E+02	4.70E+01	3.88E+01	3.88E+01
75-01-4	Vinyl Chloride	μg/L	2.32E+01	1.79E+03	2.32E+01	1.16E+02	1.79E+03	1.16E+02	5.09E+00	3.93E+02	5.09E+00
108-38-3	Xylene, m-	μg/L	-	2.05E+04	2.05E+04	-	2.05E+04	2.05E+04	-	3.96E+03	3.96E+03
95-47-6	Xylene, o-	μg/L	-	2.28E+04	2.28E+04	-	2.28E+04	2.28E+04	-	4.44E+03	4.44E+03
106-42-3	Xylene, p-	μg/L	-	2.19E+04	2.19E+04	-	2.19E+04	2.19E+04	-	4.25E+03	4.25E+03
1330-20-7	Xylene, Mixture	μg/L	-	2.16E+04	2.16E+04	-	2.16E+04	2.16E+04	-	4.20E+03	4.20E+03

		Recreational User Swimming		eational User nming		ational User Iming		ational User iming
Analyte	Units	Canceri	Hazard	No Action	Hazard	No Action	Hazard	No Action
Aluminum	μg/L	-	2.23E+05	2.23E+05	4.57E+04	4.57E+04	8.42E+04	8.42E+04
Antimony (metallic)	μg/L	-	4.48E+01	4.48E+01	1.35E+01	1.35E+01	2.31E+01	2.31E+01
Arsenic, Inorganic	μg/L	1.71E+00	6.70E+01	1.71E+00	1.37E+01	1.71E+00	2.52E+01	1.71E+00
Barium	μg/L	-	1.34E+04	1.34E+04	5.01E+03	5.01E+03	8.14E+03	8.14E+03
Beryllium and compounds	μg/L	_	1.72E+01	1.72E+01	9.31E+00	9.31E+00	1.36E+01	1.36E+01
Boron And Borates Only	μg/L	-	4.47E+04	4.47E+04	9.14E+03	9.14E+03	1.68E+04	1.68E+04
Cadmium (Water)	μg/L	-	5.15E+00	5.15E+00	2.10E+00	2.10E+00	3.33E+00	3.33E+00
Chromium(III), Insoluble Salts	μg/L	-	2.33E+04	2.33E+04	1.20E+04	1.20E+04	1.78E+04	1.78E+04
Chromium(VI)	μg/L	1.89E-01	4.50E+01	1.89E-01	2.32E+01	1.89E-01	3.44E+01	1.89E-01
Chromium (Total) ^b	μg/L	-	-	-	-	-	-	-
Cobalt	μg/L	-	7.50E+01	7.50E+01	1.42E+01	1.42E+01	2.65E+01	2.65E+01
Copper	μg/L	-	8.94E+03	8.94E+03	1.83E+03	1.83E+03	3.37E+03	3.37E+03
Fluoride	μg/L	-	8.94E+03	8.94E+03	1.83E+03	1.83E+03	3.37E+03	3.37E+03
Iron	μg/L	-	1.56E+05	1.56E+05	3.20E+04	3.20E+04	5.89E+04	5.89E+04
Lead ^c	μg/L	-	-	1.50E+01	-	1.50E+01	-	1.50E+01
Manganese (Non-Diet)	μg/L	-	1.03E+03	1.03E+03	4.40E+02	4.40E+02	6.90E+02	6.90E+02
Mercury, Inorganic Salts	μg/L	-	2.01E+01	2.01E+01	7.51E+00	7.51E+00	1.22E+01	1.22E+01
Molybdenum	μg/L	-	1.12E+03	1.12E+03	2.29E+02	2.29E+02	4.21E+02	4.21E+02
Nickel Soluble Salts	μg/L	-	2.62E+03	2.62E+03	7.32E+02	7.32E+02	1.27E+03	1.27E+03
Selenium	μg/L	-	1.12E+03	1.12E+03	2.29E+02	2.29E+02	4.21E+02	4.21E+02
Silver	μg/L	-	3.23E+02	3.23E+02	1.22E+02	1.22E+02	1.98E+02	1.98E+02
Thallium (Soluble Salts)	μg/L	-	2.23E+00	2.23E+00	4.57E-01	4.57E-01	8.42E-01	8.42E-01
Uranium (Insoluble Compounds) ^d	μg/L	-	6.71E+02	6.71E+02	1.37E+02	1.37E+02	2.52E+02	2.52E+02
Uranium (Soluble Salts)d	μg/L	-	4.47E+01	4.47E+01	9.14E+00	9.14E+00	1.68E+01	1.68E+01
Vanadium and Compounds	μg/L	-	1.48E+02	1.48E+02	6.92E+01	6.92E+01	1.06E+02	1.06E+02
Zinc and Compounds	μg/L	-	7.21E+04	7.21E+04	1.41E+04	1.41E+04	2.61E+04	2.61E+04
Acenaphthene	μg/L	-	5.62E+02	5.62E+02	3.02E+02	3.02E+02	4.42E+02	4.42E+02
Acenaphthylene ^e	μg/L	-	5.62E+02	5.62E+02	3.02E+02	3.02E+02	4.42E+02	4.42E+02
Acrylonitrile	μg/L	4.60E+00	2.11E+03	4.60E+00	4.48E+02	4.60E+00	8.19E+02	4.60E+00
Anthracene	μg/L	-	1.50E+03	1.50E+03	8.35E+02	8.35E+02	1.21E+03	1.21E+03
Benzene	μg/L	1.84E+01	2.27E+02	1.84E+01	8.98E+01	1.84E+01	1.44E+02	1.84E+01
Bis(2-ethylhexyl)phthalate ^f	μg/L	2.01E+02	5.42E+03	2.01E+02	9.75E+02	2.01E+02	1.83E+03	2.01E+02
Bromodichloromethane	μg/L	2.71E+01	8.94E+02	2.71E+01	2.70E+02	2.71E+01	4.62E+02	2.71E+01
Carbazole	μg/L	1.41E+01	-	1.41E+01	-	1.41E+01	-	1.41E+01
Carbon Tetrachloride	μg/L	1.13E+01	1.69E+02	1.13E+01	7.26E+01	1.13E+01	1.14E+02	1.13E+01
Chloroform	μg/L	4.69E+01	9.09E+02	4.69E+01	3.02E+02	4.69E+01	5.05E+02	4.69E+01
Dichlorodifluoromethane (Freon-12) ^f	μg/L	-	1.50E+04	1.50E+04	5.39E+03	5.39E+03	8.86E+03	8.86E+03

		Recreational User Swimming	Adult Recre Swim			ational User Iming		eational User
Analyte	Units	Canceri	Hazard	No Action	Hazard	No Action	Hazard	No Action
Dichloroethane, 1,1-f	μg/L	2.65E+02	1.91E+04	2.65E+02	6.21E+03	2.65E+02	1.05E+04	2.65E+02
Dichloroethane, 1,2-	μg/L	2.01E+01	7.58E+02	2.01E+01	2.16E+02	2.01E+01	3.74E+02	2.01E+01
Dichloroethylene, 1,1-	μg/L	-	3.28E+03	3.28E+03	1.24E+03	1.24E+03	2.01E+03	2.01E+03
Dichloroethylene, 1,2- (Mixed Isomers)	μg/L	-	6.19E+02	6.19E+02	2.29E+02	2.29E+02	3.74E+02	3.74E+02
Dichloroethylene, cis-1,2-	μg/L	-	1.37E+02	1.37E+02	5.09E+01	5.09E+01	8.30E+01	8.30E+01
Dichloroethylene, trans-1,2-	μg/L	-	1.37E+03	1.37E+03	5.09E+02	5.09E+02	8.30E+02	8.30E+02
Dieldrin	μg/L	9.75E-03	3.66E-01	9.75E-03	2.00E-01	9.75E-03	2.90E-01	9.75E-03
Dioxins/Furans, Total (as TCDD)g	μg/L	2.17E-05	1.90E-04	2.17E-05	3.41E-05	2.17E-05	6.41E-05	2.17E-05
~HpCDD, 2,3,7,8-	μg/L	2.17E-03	1.90E-02	2.17E-03	3.41E-03	2.17E-03	6.41E-03	2.17E-03
~HpCDF, 2,3,7,8-	μg/L	2.17E-03	1.90E-02	2.17E-03	3.41E-03	2.17E-03	6.41E-03	2.17E-03
~HxCDD	μg/L	2.17E-04	1.90E-03	2.17E-04	3.41E-04	2.17E-04	6.41E-04	2.17E-04
~HxCDF, 2,3,7,8-	μg/L	2.17E-04	1.90E-03	2.17E-04	3.41E-04	2.17E-04	6.41E-04	2.17E-04
~OCDD	μg/L	7.23E-02	6.33E-01	7.23E-02	1.14E-01	7.23E-02	2.14E-01	7.23E-02
~OCDF	μg/L	7.23E-02	6.33E-01	7.23E-02	1.14E-01	7.23E-02	2.14E-01	7.23E-02
~PeCDD, 2,3,7,8-	μg/L	2.17E-05	1.90E-04	2.17E-05	3.41E-05	2.17E-05	6.41E-05	2.17E-05
~PeCDF, 1,2,3,7,8-	μg/L	7.23E-04	6.33E-03	7.23E-04	1.14E-03	7.23E-04	2.14E-03	7.23E-04
~PeCDF, 2,3,4,7,8-	μg/L	7.23E-05	6.33E-04	7.23E-05	1.14E-04	7.23E-05	2.14E-04	7.23E-05
~TCDD, 2,3,7,8-	μg/L	2.17E-05	1.90E-04	2.17E-05	3.41E-05	2.17E-05	6.41E-05	2.17E-05
~TCDF, 2,3,7,8-	μg/L	2.17E-04	1.90E-03	2.17E-04	3.41E-04	2.17E-04	6.41E-04	2.17E-04
Ethylbenzene	μg/L	3.63E+01	9.83E+02	3.63E+01	4.90E+02	3.63E+01	7.34E+02	3.63E+01
Fluoranthene	μg/L	-	1.08E+04	1.09E+04	1.95E+03	1.95E+03	3.66E+03	3.66E+03
Fluorene	μg/L	-	2.70E+02	2.70E+02	1.48E+02	1.48E+02	2.15E+02	2.15E+02
Hexachlorobenzene	μg/L	1.76E+00	2.71E+00	1.76E+00	4.87E-01	4.87E-01	9.15E-01	9.15E-01
Naphthalene	μg/L	3.22E+00	3.79E+02	3.22E+00	1.90E+02	3.22E+00	2.84E+02	3.22E+00
Nitroaniline, 2-	μg/L	-	1.13E+03	1.13E+03	3.40E+02	3.40E+02	5.82E+02	5.82E+02
Nitroso-di-N-propylamine, N-	μg/L	3.01E-01	-	3.01E-01	-	3.01E-01	-	3.01E-01
Pentachlorophenol	μg/L	1.92E-01	1.77E+01	1.92E-01	9.96E+00	1.92E-01	1.43E+01	1.92E-01
Phenanthrene ^e	μg/L	-	5.62E+02	5.62E+02	3.02E+02	3.02E+02	4.42E+02	4.42E+02
Polychlorinated Biphenyls, Total	μg/L	7.05E+00	-	7.05E+00	-	7.05E+00	-	7.05E+00
~Aroclor 1016	μg/L	4.03E+01	1.90E+01	1.90E+01	3.41E+00	3.41E+00	6.41E+00	6.41E+00
~Aroclor 1221	μg/L	4.29E-02	-	4.29E-02	-	4.29E-02	-	4.29E-02
~Aroclor 1232	μg/L	4.29E-02	-	4.29E-02	-	4.29E-02	-	4.29E-02
~Aroclor 1242	μg/L	1.41E+00	-	1.41E+00	-	1.41E+00	-	1.41E+00
~Aroclor 1248	μg/L	1.41E+00	-	1.41E+00	-	1.41E+00	-	1.41E+00
~Aroclor 1254	μg/L	1.41E+00	5.42E+00	1.41E+00	9.75E-01	9.75E-01	1.83E+00	1.41E+00
~Aroclor 1260	μg/L	1.41E+00	-	1.41E+00	-	1.41E+00	-	1.41E+00

		Recreational User Swimming		eational User nming		ational User iming		eational User nming
Analyte	Units	Cancer ⁱ	Hazard	No Action	Hazard	No Action	Hazard	No Action
Polycyclic aromatic hydrocarbons, Total Carcinogenic ^h	μg/L	6.85E-01	8.14E+01	6.85E-01	1.46E+01	6.85E-01	2.75E+01	6.85E-01
~Benz[a]anthracene	μg/L	6.85E+00	-	6.85E+00	-	6.85E+00	-	6.85E+00
~Benzo[a]pyrene	μg/L	6.85E-01	8.14E+01	6.85E-01	1.46E+01	6.85E-01	2.75E+01	6.85E-01
~Benzo[b]fluoranthene	μg/L	6.85E+00	-	6.85E+00	-	6.85E+00	-	6.85E+00
~Benzo[k]fluoranthene	μg/L	6.85E+01	-	6.85E+01	-	6.85E+01	-	6.85E+01
~Chrysene	μg/L	6.85E+02	-	6.85E+02	-	6.85E+02	-	6.85E+02
~Dibenz[a,h]anthracene	μg/L	6.85E-01	-	6.85E-01	1	6.85E-01	-	6.85E-01
~Indeno[1,2,3-cd]pyrene	μg/L	6.85E+00	-	6.85E+00	-	6.85E+00	-	6.85E+00
Pyrene	μg/L	-	9.15E+01	9.15E+01	5.17E+01	5.17E+01	7.43E+01	7.43E+01
Tetrachloroethylene	μg/L	2.04E+02	1.27E+02	1.27E+02	6.27E+01	6.27E+01	9.42E+01	9.42E+01
Toluene ^f	μg/L	-	2.43E+03	2.43E+03	1.13E+03	1.13E+03	1.73E+03	1.73E+03
Trichloro-1,2,2-trifluoroethane, 1,1,2- (Freon-113) ^f	μg/L	-	1.02E+06	1.02E+06	4.63E+05	4.63E+05	7.14E+05	7.14E+05
Trichloroethane, 1,1,1-	μg/L	-	1.12E+05	1.12E+05	4.45E+04	4.45E+04	7.13E+04	7.13E+04
Trichloroethane, 1,1,2-	μg/L	2.85E+01	4.25E+02	2.85E+01	1.32E+02	2.85E+01	2.24E+02	2.85E+01
Trichloroethylene	μg/L	1.52E+01	3.01E+01	1.52E+01	1.17E+01	1.17E+01	1.88E+01	1.52E+01
Vinyl Chloride	μg/L	4.95E-02	2.64E+02	4.95E-02	8.89E+01	4.95E-02	1.48E+02	4.95E-02
Xylene, m-	μg/L	-	3.68E+03	3.68E+03	1.85E+03	1.85E+03	2.77E+03	2.77E+03
Xylene, o-	μg/L	-	4.09E+03	4.09E+03	2.03E+03	2.03E+03	3.04E+03	3.04E+03
Xylene, p-	μg/L	-	3.93E+03	3.93E+03	1.96E+03	1.96E+03	2.94E+03	2.94E+03
Xylene, Mixture	μg/L	-	3.88E+03	3.88E+03	1.94E+03	1.94E+03	2.90E+03	2.90E+03

		Recreational User Wading ^a	Adult Recrea			eational User ding ^a		ational User ding ^a
Analyte	Units	Canceri	Hazard	No Action	Hazard	No Action	Hazard	No Action
Aluminum	μg/L	-	2.04E+06	2.04E+06	4.56E+05	4.56E+05	5.88E+05	5.88E+05
Antimony (metallic)	μg/L	-	1.22E+02	1.22E+02	2.73E+01	2.73E+01	3.53E+01	3.53E+01
Arsenic, Inorganic	μg/L	1.41E+01	6.11E+02	1.41E+01	1.37E+02	1.41E+01	1.76E+02	1.41E+01
Barium	μg/L	-	2.85E+04	2.85E+04	6.38E+03	6.38E+03	8.24E+03	8.24E+03
Beryllium and compounds	μg/L	-	2.85E+01	2.85E+01	6.38E+00	6.38E+00	8.24E+00	8.24E+00
Boron And Borates Only	μg/L	-	4.08E+05	4.08E+05	9.12E+04	9.12E+04	1.18E+05	1.18E+05
Cadmium (Water)	μg/L	-	1.02E+01	1.02E+01	2.28E+00	2.28E+00	2.94E+00	2.94E+00
Chromium(III), Insoluble Salts	μg/L	-	3.97E+04	3.97E+04	8.89E+03	8.89E+03	1.15E+04	1.15E+04
Chromium(VI)	μg/L	1.39E-01	7.64E+01	1.39E-01	1.71E+01	1.39E-01	2.21E+01	1.39E-01
Chromium (Total) ^b	μg/L	-	-	-	-	-	-	-
Cobalt	μg/L	-	1.53E+03	1.53E+03	3.42E+02	3.42E+02	4.41E+02	4.41E+02
Copper	μg/L	-	8.15E+04	8.15E+04	1.82E+04	1.82E+04	2.35E+04	2.35E+04
Fluoride	μg/L	-	8.15E+04	8.15E+04	1.82E+04	1.82E+04	2.35E+04	2.35E+04
Iron	μg/L	-	1.43E+06	1.43E+06	3.19E+05	3.19E+05	4.12E+05	4.12E+05
Lead ^c	μg/L	-	-	1.50E+01	-	1.50E+01	-	1.50E+01
Manganese (Non-Diet)	μg/L	-	1.96E+03	1.96E+03	4.38E+02	4.38E+02	5.65E+02	5.65E+02
Mercury, Inorganic Salts	μg/L	-	4.28E+01	4.28E+01	9.57E+00	9.57E+00	1.24E+01	1.24E+01
Molybdenum	μg/L	-	1.02E+04	1.02E+04	2.28E+03	2.28E+03	2.94E+03	2.94E+03
Nickel Soluble Salts	μg/L	-	8.15E+03	8.15E+03	1.82E+03	1.82E+03	2.35E+03	2.35E+03
Selenium	μg/L	-	1.02E+04	1.02E+04	2.28E+03	2.28E+03	2.94E+03	2.94E+03
Silver	μg/L	-	6.79E+02	6.79E+02	1.52E+02	1.52E+02	1.96E+02	1.96E+02
Thallium (Soluble Salts)	μg/L	-	2.04E+01	2.04E+01	4.56E+00	4.56E+00	5.88E+00	5.88E+00
Uranium (Insoluble Compounds) ^d	μg/L	-	6.11E+03	6.11E+03	1.37E+03	1.37E+03	1.76E+03	1.76E+03
Uranium (Soluble Salts) ^d	μg/L	-	4.08E+02	4.08E+02	9.12E+01	9.12E+01	1.18E+02	1.18E+02
Vanadium and Compounds	μg/L	-	2.67E+02	2.67E+02	5.97E+01	5.97E+01	7.71E+01	7.71E+01
Zinc and Compounds	μg/L	-	1.02E+06	1.02E+06	2.28E+05	2.28E+05	2.94E+05	2.94E+05
Acenaphthene	μg/L	-	9.34E+02	9.34E+02	2.09E+02	2.09E+02	2.70E+02	2.70E+02
Acenaphthylene ^e	μg/L	-	9.34E+02	9.34E+02	2.09E+02	2.09E+02	2.70E+02	2.70E+02
Acrylonitrile	μg/L	2.91E+01	1.52E+04	2.91E+01	3.39E+03	2.91E+01	4.38E+03	2.91E+01
Anthracene	μg/L	-	2.45E+03	2.45E+03	5.49E+02	5.49E+02	7.09E+02	7.09E+02
Benzene	μg/L	2.18E+01	4.62E+02	2.18E+01	1.03E+02	2.18E+01	1.33E+02	2.18E+01
Bis(2-ethylhexyl)phthalatef	μg/L	-	-	-	-	-	-	-
Bromodichloromethane	μg/L	5.10E+01	2.44E+03	5.10E+01	5.46E+02	5.10E+01	7.04E+02	5.10E+01
Carbazole	μg/L	1.18E+01	-	1.18E+01	-	1.18E+01	-	1.18E+01
Carbon Tetrachloride	μg/L	1.19E+01	3.20E+02	1.19E+01	7.16E+01	1.19E+01	9.25E+01	1.19E+01
Chloroform	μg/L	7.34E+01	2.19E+03	7.34E+01	4.91E+02	7.34E+01	6.33E+02	7.34E+01
Dichlorodifluoromethane (Freon-12)f	μg/L	-	3.34E+04	3.34E+04	7.47E+03	7.47E+03	9.65E+03	9.65E+03

		Recreational User Wading ^a		eational User ling ^a		eational User ding ^a		eational User ding ^a
Analyte	Units	Cancer ⁱ	Hazard	No Action	Hazard	No Action	Hazard	No Action
Dichloroethane, 1,1-f	μg/L	4.32E+02	4.75E+04	4.32E+02	1.06E+04	4.32E+02	1.37E+04	4.32E+02
Dichloroethane, 1,2-	μg/L	4.32E+01	2.28E+03	4.32E+01	5.09E+02	4.32E+01	6.57E+02	4.32E+01
Dichloroethylene, 1,1-	μg/L	-	6.95E+03	6.95E+03	1.56E+03	1.56E+03	2.01E+03	2.01E+03
Dichloroethylene, 1,2- (Mixed Isomers)	μg/L	-	1.33E+03	1.33E+03	2.97E+02	2.97E+02	3.84E+02	3.84E+02
Dichloroethylene, cis-1,2-	μg/L	-	2.95E+02	2.95E+02	6.61E+01	6.61E+01	8.53E+01	8.53E+01
Dichloroethylene, trans-1,2-	μg/L	-	2.95E+03	2.95E+03	6.61E+02	6.61E+02	8.53E+02	8.53E+02
Dieldrin	μg/L	7.82E-03	6.03E-01	7.82E-03	1.35E-01	7.82E-03	1.74E-01	7.82E-03
Dioxins/Furans, Total (as TCDD)g	μg/L	-	-	-	-	-	-	-
~HpCDD, 2,3,7,8-	μg/L	-	-	-	-	-	-	-
~HpCDF, 2,3,7,8-	μg/L	-	-	-	-	-	-	-
~HxCDD	μg/L	-	-	-	-	-	-	-
~HxCDF, 2,3,7,8-	μg/L	-	-	-	-	-	-	-
~OCDD	μg/L	-	-	-	-	-	-	-
~OCDF	μg/L	-	-	-	-	-	-	-
~PeCDD, 2,3,7,8-	μg/L	-	-	-	-	-	-	-
~PeCDF, 1,2,3,7,8-	μg/L	-	-	-	-	-	-	-
~PeCDF, 2,3,4,7,8-	μg/L	-	-	-	-	-	-	-
~TCDD, 2,3,7,8-	μg/L	-	-	-	-	-	-	-
~TCDF, 2,3,7,8-	μg/L	-	-	-	-	-	-	-
Ethylbenzene	μg/L	3.21E+01	1.70E+03	3.21E+01	3.80E+02	3.21E+01	4.91E+02	3.21E+01
Fluoranthene	μg/L	-	-	-	-	-	-	-
Fluorene	μg/L	-	4.45E+02	4.45E+02	9.95E+01	9.95E+01	1.28E+02	1.28E+02
Hexachlorobenzene	μg/L	-	-	-	-	-	-	-
Naphthalene	μg/L	2.83E+00	6.54E+02	2.83E+00	1.46E+02	2.83E+00	1.89E+02	2.83E+00
Nitroaniline, 2-	μg/L	-	3.11E+03	3.11E+03	6.95E+02	6.95E+02	8.97E+02	8.97E+02
Nitroso-di-N-propylamine, N-	μg/L	9.07E-01	-	9.07E-01	-	9.07E-01	-	9.07E-01
Pentachlorophenol	μg/L	1.49E-01	2.88E+01	1.49E-01	6.44E+00	1.49E-01	8.31E+00	1.49E-01
Phenanthrene ^e	μg/L	-	9.34E+02	9.34E+02	2.09E+02	2.09E+02	2.70E+02	2.70E+02
Polychlorinated Biphenyls, Total	μg/L	-	-	-	-	-	-	-
~Aroclor 1016	μg/L	-	-	-	-	-	-	-
~Aroclor 1221	μg/L	3.35E-02	-	3.35E-02	-	3.35E-02	-	3.35E-02
~Aroclor 1232	μg/L	3.35E-02	-	3.35E-02	-	3.35E-02	-	3.35E-02
~Aroclor 1242	μg/L	-	-	-	-	-	-	-
~Aroclor 1248	μg/L	-	-	-	-	-	-	-
~Aroclor 1254	μg/L	-	-	-	-	-	-	-
~Aroclor 1260	μg/L	-	-	-	-	-	-	-

Table A.6a. Surface Water No Action Levels for Significant COPCs at PGDP (Continued)

(Values were calculated in October 2022 and are based on best available information.)

		Recreational User Wading ^a		eational User ding ^a		eational User ding ^a		ational User ding ^a
Analyte	Units	Canceri	Hazard	No Action	Hazard	No Action	Hazard	No Action
Polycyclic aromatic hydrocarbons, Total Carcinogenich	μg/L	-	ı	1	-	-	-	1
~Benz[a]anthracene	μg/L	-	1	1	-	-	-	1
~Benzo[a]pyrene	μg/L	-	ı	1	-	-	-	1
~Benzo[b]fluoranthene	μg/L	-	ı	1	-	-	-	1
~Benzo[k]fluoranthene	μg/L	-	ı	1	-	-	-	-
~Chrysene	μg/L	-	-	-	-	-	-	-
~Dibenz[a,h]anthracene	μg/L	-	-	-	-	-	-	-
~Indeno[1,2,3-cd]pyrene	μg/L	-	-	-	-	-	-	-
Pyrene	μg/L	-	1.49E+02	1.49E+02	3.32E+01	3.32E+01	4.29E+01	4.29E+01
Tetrachloroethylene	μg/L	1.82E+02	2.21E+02	1.82E+02	4.95E+01	4.95E+01	6.38E+01	6.38E+01
Toluene ^f	μg/L	-	4.39E+03	4.39E+03	9.83E+02	9.83E+02	1.27E+03	1.27E+03
Trichloro-1,2,2-trifluoroethane, 1,1,2- (Freon-113) ^f	μg/L	-	1.88E+06	1.88E+06	4.20E+05	4.20E+05	5.42E+05	5.42E+05
Trichloroethane, 1,1,1-	μg/L	-	2.27E+05	2.27E+05	5.08E+04	5.08E+04	6.56E+04	6.56E+04
Trichloroethane, 1,1,2-	μg/L	5.11E+01	1.12E+03	5.11E+01	2.51E+02	5.11E+01	3.24E+02	5.11E+01
Trichloroethylene	μg/L	1.78E+01	6.21E+01	1.78E+01	1.39E+01	1.39E+01	1.79E+01	1.78E+01
Vinyl Chloride	μg/L	2.31E-01	6.28E+02	2.31E-01	1.41E+02	2.31E-01	1.81E+02	2.31E-01
Xylene, m-	μg/L	-	6.34E+03	6.34E+03	1.42E+03	1.42E+03	1.83E+03	1.83E+03
Xylene, o-	μg/L	-	7.10E+03	7.10E+03	1.59E+03	1.59E+03	2.05E+03	2.05E+03
Xylene, p-	μg/L	-	6.80E+03	6.80E+03	1.52E+03	1.52E+03	1.96E+03	1.96E+03
Xylene, Mixture	μg/L	-	6.71E+03	6.71E+03	1.50E+03	1.50E+03	1.94E+03	1.94E+03

NOTES: Values are provided in these tables for significant COPCs for PGDP. Values for other COPCs can be obtained using the RAIS Chemical PRG online calculator, as modified using PGDP-specific inputs. NALs are not adjusted for solubility limits.

^a Recreational User Wading and all Worker scenarios consider dermal contact only.

^b Chromium (Total) should utilize Chromium III or Chromium VI NALs, as appropriate.

^c Lead values should be checked prior to use to ensure they are still current.

d Based on recommendation from EPA, NALs for uranium (soluble salts) now use the RfD and the RfC for soluble compounds of uranium derived from ATSDR (EPA 2016). NALs for uranium (insoluble compounds) use the RfD for uranium (soluble salts), which are available in the IRIS; the RfC for insoluble compounds of uranium are derived from ATSDR.

^e Acenaphthylene and phenanthrene use values for acenaphthene as a surrogate.

^f Analytes are not PGDP significant COPCs (Table 2.1), but are provided for project support.

g Total dioxins/furans uses values for 2,3,7,8-TCDD, see screening note 9f in the Appendix A introduction, "Screening Levels," on pages A-3-A-5.

h Total carcinogenic PAHs uses values for BaP, see screening note 9d in the Appendix A introduction, "Screening Levels," on pages A-3-A-5.

For the recreational user, ELCRs (i.e., cancer NALs) were calculated using the child/teen/adult age-adjusted lifetime scenario.

CAS Number	Analyte	Units	Outdoor Worker Wading	Excavation Worker Wading	Industrial Worker Wading
14596-10-2	Am-241	pCi/L			
10045-97-3	Cs-137	pCi/L			
13994-20-2	Np-237	pCi/L			
13981-16-3	Pu-238	pCi/L			
15117-48-3	Pu-239	pCi/L			
14119-33-6	Pu-240	pCi/L	No Sam	agning Values for the Weding Co.	amani all
14133-76-7	Tc-99	pCi/L	No Scr	eening Values for the Wading Sce	enario"
14269-63-7	Th-230	pCi/L			
7440-29-1	Th-232 ^c	pCi/L			
13966-29-5	U-234	pCi/L			
15117-96-1	U-235	pCi/L			
7440-61-1	U-238	pCi/L			

Table A.6b. Surface Water No Action Levels for Significant Radionuclide COPCs at PGDP (Continued)

(Values were calculated in October 2022 and are based on best available information.)

CAS	A a lasta	TI:4"	I	Recreational User Swimming	.b	F	Recreational Us Wading	er
Number	Analyte	Units	Peak Risk Infinite ^c	Peak Risk 1,000 Years	Secular Equilibrium	Peak Risk Infinite	Peak Risk 1,000 Years	Secular Equilibrium
14596-10-2	Am-241	pCi/L	3.49E+01	3.49E+01	4.59E+00			
10045-97-3	Cs-137	pCi/L	1.54E+02	1.54E+02	1.16E+02			
13994-20-2	Np-237	pCi/L	6.53E+00	5.15E+01	5.30E+00			
13981-16-3	Pu-238	pCi/L	2.98E+01	2.98E+01	1.06E+00			
15117-48-3	Pu-239	pCi/L	2.62E+01	2.62E+01	4.08E+00			
14119-33-6	Pu-240	pCi/L	2.62E+01	2.62E+01	2.16E+00	No Screen	ing Values for t	he Wading
14133-76-7	Tc-99	pCi/L	1.29E+03	1.29E+03	1.29E+03		Scenarioa	
14269-63-7	Th-230	pCi/L	1.22E+00	3.15E+00	1.12E+00			
7440-29-1	Th-232 ^d	pCi/L	2.46E+00	2.46E+00	2.46E+00			
13966-29-5	U-234	pCi/L	1.86E+00	4.64E+01	1.10E+00			
15117-96-1	U-235	pCi/L	4.84E+00	4.16E+01	4.84E+00			
7440-61-1	U-238	pCi/L	1.07E+00	4.05E+01	1.07E+00			

NOTES: Values are provided in these tables for significant radionuclide COPCs for PGDP. Values for other radionuclides can be obtained using the EPA Radionuclide PRG calculator, as modified using PGDP-specific inputs.

Radionuclide NALs are based on the cancer endpoint using a target ELCR of 1.0E-6. The surface water radionuclide NALs are based on the ingestion pathway only.

^a The wading exposure scenarios shown in this table only consider dermal contact with surface water, not indicental surface water ingestion (see Figure 8.1). Dermal absorption is not evaluated for radionuclides so no screening values are derived for the wading exposure scenarios.

^b For the recreational user, the radionuclide ALs were calculated using the child/adult age-adjusted scenario (i.e., 26-year exposure duration, with 6 years as a child and 20 years as an adult). The EPA Radiological PRG calculator only allows exposure parameter entries for two life stages (i.e., child and adult), rather than the four life stage groups provided in the EPA and RAIS chemical online calculators. Thus, the EPA Radiological PRG calculator does not allow entry of the teen exposure parameters shown in Table B.5.

^c The time frame for calculating peak risk over an infinite period in the EPA Radionuclide PRG calculator is 1 trillion years. NASA most recently has estimated the age of the universe as being 13.7 billion years, with an uncertainty of only 200 million years (https://imagine.gsfc.nasa.gov).

^d Analyte is not a PGDP significant COPC (Table 2.1), but it is provided for project support.

Table A.7a. Risk-Based SSLs for Protection of RGA Groundwater for Significant COPCs at PGDP (Values were calculated in October 2022 and are based on best available information.)

CAS		SS	SLs for EPA M	CLa	SSLs PG	DP NALs for th (See Table A.5a	
Number	Chemical	SSL 1	SSL 20	GW Conc.b	SSL 1	SSL 20	GW Conc.b
		(mg/kg)	(mg/kg)	(µg/L)	(mg/kg)	(mg/kg)	(µg/L)
7429-90-5	Aluminum	-	-	-	3.00E+03	5.99E+04	2.00E+03
7440-36-0	Antimony (metallic)	2.71E-01	5.42E+00	6.00E+00	3.52E-02	7.04E-01	7.79E-01
7440-38-2	Arsenic, Inorganic	2.92E-01	5.84E+00	1.00E+01	1.51E-03	3.02E-02	5.17E-02
	Barium	8.24E+01	1.65E+03	2.00E+03	1.55E+01	3.11E+02	3.77E+02
7440-41-7	Beryllium and compounds	3.16E+00	6.32E+01	4.00E+00	1.95E+00	3.89E+01	2.46E+00
7440-42-8	Boron And Borates Only	-	-	-	1.28E+00	2.56E+01	3.99E+02
7440-43-9	Cadmium (Water)	3.76E-01	7.52E+00	5.00E+00	1.38E-02	2.77E-01	1.84E-01
16065-83-1	Chromium (III), Insoluble Salts	-	-	-	4.04E+06	8.09E+07	2.25E+03
18540-29-9	Chromium (VI)	-	-	-	6.72E-04	1.34E-02	3.50E-02
7440-47-3	Chromium (Total) ^c	1.80E+05	3.60E+06	1.00E+02	-	-	-
7440-48-4	Cobalt	-	-	-	2.71E-02	5.43E-01	6.01E-01
7440-50-8	Copper	4.58E+01	9.15E+02	1.30E+03	2.81E+00	5.62E+01	7.99E+01
16984-48-8		6.01E+02	1.20E+04	4.00E+03	1.20E+01	2.40E+02	7.99E+01
7439-89-6	Iron	-	-	-	3.52E+01	7.04E+02	1.40E+03
7439-92-1	Lead	1.35E+01	2.70E+02	1.50E+01	_	-	-
7439-96-5	Manganese (Non-Diet)	-	-	-	2.83E+00	5.65E+01	4.34E+01
Various	Mercury, Inorganic Salts d	1.04E-01	2.09E+00	2.00E+00	2.95E-02	5.91E-01	5.66E-01
7439-98-7	Molybdenum	-	-	-	2.02E-01	4.03E+00	9.98E+00
7440-02-0	Nickel Soluble Salts	-	-	-	2.56E+00	5.12E+01	3.92E+01
7782-49-2	Selenium	2.60E-01	5.20E+00	5.00E+01	5.19E-02	1.04E+00	9.98E+00
7440-22-4	Silver	-	-	-	7.99E-02	1.60E+00	9.41E+00
7440-28-0	Thallium (Soluble Salts)	1.42E-01	2.85E+00	2.00E+00	1.42E-03	2.84E-02	2.00E-02
N/A	Uranium (Insoluble Compounds) e	1.35E+01	2.70E+02	3.00E+01	2.70E+00-	5.39E+01-	5.99E+00-
7440-61-1	Uranium (Soluble Salts) e	1.35E+01	2.70E+02	3.00E+01	1.80E-01	3.60E+00	3.99E-01
7440-62-2	Vanadium and Compounds	-	-	-	8.64E+00	1.73E+02	8.64E+00
7440-66-6	Zinc and Compounds	-	-	-	3.73E+01	7.46E+02	6.00E+02
83-32-9	Acenaphthene	-	-	-	5.49E-01	1.10E+01	5.35E+01
208-96-8	Acenaphthylene ^f	-	-	-	5.49E-01	1.10E+01	5.35E+01
107-13-1	Acrylonitrile	-	-	-	1.14E-05	2.28E-04	5.23E-02
120-12-7	Anthracene	-	-	-	5.81E+00	1.16E+02	1.77E+02
71-43-2	Benzene	2.56E-03	5.12E-02	5.00E+00	2.33E-04	4.66E-03	4.55E-01
117-81-7	Bis(2-ethylhexyl)phthalate ^g	1.44E+00	2.87E+01	6.00E+00	1.33E+00	2.66E+01	5.56E+00
75-27-4	Bromodichloromethane	2.17E-02	4.34E-01	8.00E+01	3.65E-05	7.30E-04	1.34E-01
86-74-8	Carbazole	-	-	-	3.76E-02	7.51E-01	2.03E+00
56-23-5	Carbon Tetrachloride	1.94E-03	3.89E-02	5.00E+00	1.77E-04	3.54E-03	4.55E-01
67-66-3	Chloroform	2.22E-02	4.43E-01	8.00E+01	6.12E-05	1.22E-03	2.21E-01

Table A.7a. Risk-Based SSLs for Protection of RGA Groundwater for Significant COPCs at PGDP (Continued) (Values were calculated in October 2022 and are based on best available information.)

CAS	GL 1.1	SS	SLs for EPA M	CLa	SSLs PG	DP NALs for the	
Number	Chemical	SSL 1	SSL 20	GW Conc.b	SSL 1	SSL 20	GW Conc.b
		(mg/kg)	(mg/kg)	(µg/L)	(mg/kg)	(mg/kg)	(µg/L)
75-71-8	Dichlorodifluoromethane (Freon-12) g	-	-	-	3.04E-02	6.08E-01	1.97E+01
75-34-3	Dichloroethane, 1,1-g	-	-	-	7.82E-04	1.56E-02	2.75E+00
107-06-2	Dichloroethane, 1,2-	1.42E-03	2.84E-02	5.00E+00	4.84E-05	9.69E-04	1.71E-01
75-35-4	Dichloroethylene, 1,1-	2.51E-03	5.03E-02	7.00E+00	1.02E-02	2.04E-01	2.85E+01
540-59-0	Dichloroethylene, 1,2- (Mixed Isomers)	-	-	-	4.78E-03	9.56E-02	1.63E+01
	Dichloroethylene, cis-1,2-	2.06E-02	4.12E-01	7.00E+01	1.06E-03	2.12E-02	3.61E+00
156-60-5	Dichloroethylene, trans-1,2-	3.13E-02	6.27E-01	1.00E+02	2.12E-03	4.24E-02	6.78E+00
60-57-1	Dieldrin	-	-	-	7.08E-05	1.42E-03	1.75E-03
1746-01-6	Dioxins/Furans, Total (as TCDD) h	1.50E-05	2.99E-04	3.00E-05	5.91E-08	1.18E-06	1.19E-07
37871-00-4	~HpCDD, 2,3,7,8-	-	-	-	2.75E-05	5.51E-04	1.19E-05
	~HpCDF, 2,3,7,8-	-	-	-	1.54E-05	3.08E-04	1.19E-05
34465-46-8	~HxCDD	-	-	-	8.33E-06	1.67E-04	5.99E-06
55684-94-1	~HxCDF, 2,3,7,8-	-	-	-	4.76E-06	9.52E-05	5.99E-06
3268-87-9	~OCDD	-	-	-	7.75E-03	1.55E-01	2.00E-03
39001-02-0	~OCDF	-	-	-	4.34E-03	8.68E-02	2.00E-03
36088-22-9	~PeCDD, 2,3,7,8-	-	-	-	5.19E-07	1.04E-05	5.99E-07
57117-41-6	~PeCDF, 1,2,3,7,8-	-	-	-	9.31E-06	1.86E-04	2.00E-05
57117-31-4	~PeCDF, 2,3,4,7,8-	-	-	-	9.31E-07	1.86E-05	2.00E-06
1746-01-6	~TCDD, 2,3,7,8-	1.50E-05	2.99E-04	3.00E-05	5.91E-08	1.18E-06	1.19E-07
51207-31-9	~TCDF, 2,3,7,8-	-	-	-	3.31E-07	6.62E-06	1.19E-06
100-41-4	Ethylbenzene	7.85E-01	1.57E+01	7.00E+02	1.68E-03	3.36E-02	1.50E+00
206-44-0	Fluoranthene	-	-	-	8.91E+00	1.78E+02	8.02E+01
86-73-7	Fluorene	-	-	-	5.45E-01	1.09E+01	2.94E+01
118-74-1	Hexachlorobenzene	1.26E-02	2.52E-01	1.00E+00	1.23E-04	2.46E-03	9.76E-03
91-20-3	Naphthalene	-	-	-	3.85E-04	7.70E-03	1.17E-01
88-74-4	Nitroaniline, 2-	-	-	-	8.01E-03	1.60E-01	1.89E+01
621-64-7	Nitroso-di-N-propylamine, N-	-	-	-	8.10E-06	1.62E-04	1.08E-02
87-86-5	Pentachlorophenol	1.38E-03	2.77E-02	1.00E+00	5.71E-05	1.14E-03	4.13E-02
85-01-8	Phenanthrenee	-	-	-	5.49E-01	1.10E+01	5.35E+01
1336-36-3	Polychlorinated Biphenyls, Total	7.82E-02	1.56E+00	5.00E-01	6.82E-03	1.36E-01	4.36E-02
	~Aroclor 1016	-	-	-	1.34E-02	2.68E-01	1.40E-01
11104-28-2	~Aroclor 1221	-	-	-	8.00E-05	1.60E-03	4.71E-03
11141-16-5	~Aroclor 1232	-	-	-	8.00E-05	1.60E-03	4.71E-03
	~Aroclor 1242	-	-	-	1.23E-03	2.45E-02	7.85E-03
	~Aroclor 1248	-	-	-	1.20E-03	2.41E-02	7.85E-03
	~Aroclor 1254	-	-	-	2.05E-03	4.10E-02	7.85E-03
	~Aroclor 1260	-	-	-	5.49E-03	1.10E-01	7.85E-03

Table A.7a. Risk-Based SSLs for Protection of RGA Groundwater for Significant COPCs at PGDP (Continued)

(Values were calculated in October 2022 and are based on best available information.)

CAS	Chemical	SS	SLs for EPA M	CL ^a	SSLs PG	DP NALs for th (See Table A.5)	
Number	Chemical	SSL 1 (mg/kg)	SSL 20 (mg/kg)	GW Conc. ^b (μg/L)	SSL 1 (mg/kg)	SSL 20 (mg/kg)	GW Conc. ^b (μg/L)
50-32-8	Polycyclic aromatic hydrocarbons, Total Carcinogenic i	2.35E-01	4.70E+00	2.00E-01	2.94E-02	5.89E-01	2.51E-02
56-55-3	~Benz[a]anthracene	-	-	-	1.05E-02	2.11E-01	2.98E-02
50-32-8	~Benzo[a]pyrene	2.35E-01	4.70E+00	2.00E-01	2.94E-02	5.89E-01	2.51E-02
205-99-2	~Benzo[b]fluoranthene	-	-	-	3.00E-01	6.01E+00	2.51E-01
207-08-9	~Benzo[k]fluoranthene	-	-	-	2.94E+00	5.89E+01	2.51E+00
218-01-9	~Chrysene	-	-	-	9.05E+00	1.81E+02	2.51E+01
53-70-3	~Dibenz[a,h]anthracene	-	-	-	9.58E-02	1.92E+00	2.51E-02
193-39-5	~Indeno[1,2,3-cd]pyrene	-	-	-	9.78E-01	1.96E+01	2.51E-01
129-00-0	Pyrene	-	-	-	1.32E+00	2.63E+01	1.21E+01
127-18-4	Tetrachloroethylene	2.27E-03	4.55E-02	5.00E+00	1.84E-03	3.69E-02	4.06E+00
108-88-3	Toluene ^e	6.92E-01	1.38E+01	1.00E+03	7.62E-02	1.52E+00	1.10E+02
76-13-1	Trichloro-1,2,2-trifluoroethane, 1,1,2- (Freon-113) f	-	-	-	2.56E+00	5.13E+01	1.02E+03
71-55-6	Trichloroethane, 1,1,1-	7.01E-02	1.40E+00	2.00E+02	2.81E-01	5.62E+00	8.01E+02
79-00-5	Trichloroethane, 1,1,2-	1.62E-03	3.24E-02	5.00E+00	1.35E-05	2.69E-04	4.15E-02
79-01-6	Trichloroethylene	1.79E-03	3.57E-02	5.00E+00	1.01E-04	2.02E-03	2.83E-01
75-01-4	Vinyl Chloride	6.90E-04	1.38E-02	2.00E+00	6.47E-06	1.29E-04	1.88E-02
108-38-3	Xylene, m-	-	-	-	1.88E-02	3.77E-01	1.93E+01
95-47-6	Xylene, o-	-	-	-	1.91E-02	3.81E-01	1.93E+01
106-42-3	Xylene, p-	-	-	-	1.89E-02	3.77E-01	1.93E+01
1330-20-7	Xylene, Mixture	9.90E+00	1.98E+02	1.00E+04	1.91E-02	3.82E-01	1.93E+01

^a Values in this table for SSLs of 1 and 20 were calculated using the RAIS Chemical PRG online calculator in October 2022 located at the Web site http://rais.oml.gov/cgi-bin/prg/PRG search?select=chem. Prior to using the values in this table in a quantitative risk assessment, a risk assessment, a risk assessment, a risk assessment be consulted to determine if any values need to be updated and to verify that the values are being used appropriately. SSL 1 indicates the soil screening level calculated for a DAF of 1. SSL 20 indicates the soil screening level calculated for a DAF of 20. The appropriate DAF is calculated on a project-specific basis, additional information is provided in Appendix E of this document.

Note: Default parameters from RAIS used are as follows:

Fraction organic carbon in soil (unitless) 0.002
Water-filled soil porosity (Lwater/Lsoil) 0.3
Dry soil bulk density (kg/L) 1.5
Soil particle density (kg/L) 2.65

b Groundwater concentrations represent the MCLs or risk-based NALs that are used as the groundwater target concentrations in the SSL equation.

^c Chromium (Total) should utilize chromium III or chromium VI, as appropriate. See list of screening levels note 9b (on page A-4).

^d SSLs for EPA MCL use MCL for mercury (elemental).

^e Based on recommendation from EPA 2016, SSLs for uranium (soluble salts) now use the RfD and the RfC for soluble compounds of uranium derived from ATSDR. SSLs for uranium (insoluble compounds) use the RfD for uranium (soluble salts), which is available in IRIS; the RfC for insoluble compounds of uranium are derived from ATSDR.

^f Values for acenaphthylene and phenanthrene use values for acenaphthene as a surrogate.

g Analytes are not PGDP significant COPCs (Table 2.1), but are provided for project support. SSLs for other COPCs can be derived using similar methods as needed.

^h Total dioxins/furans uses values for 2,3,7,8-TCDD, see list of screening levels note 9f (on page A-5).

¹ Total carcinogenic PAHs use values for BaP, see screening note 9d (page A-4).

Table A.7b. Risk-Based SSLs for Protection of RGA Groundwater for Radionuclide COPCs at PGDP

(Values were calculated in October 2022 and are based on best available information.)

			Peak Ris	k Infinite	Peak Risk	1,000 Years	Secular E	quilibrium
CAS Number	Radionuclide	Units	SSL 1 (pCi/g)	SSL 20 (pCi/g)	SSL 1 (pCi/g)	SSL 20 (pCi/g)	SSL 1 (pCi/g)	SSL 20 (pCi/g)
14596-10-2	Am-241	pCi/g	9.79E-01	1.96E+01	9.79E-01	1.96E+01	1.29E-01	2.57E+00
10045-97-3	Cs-137	pCi/g	6.39E-01	1.28E+01	6.39E-01	1.28E+01	4.79E-01	9.58E+00
13994-20-2	Np-237	pCi/g	6.77E-03	1.35E-01	5.34E-02	1.07E+00	5.50E-03	1.10E-01
13981-16-3	Pu-238	pCi/g	2.42E-01	4.84E+00	2.42E-01	4.84E+00	8.58E-03	1.72E-01
15117-48-3	Pu-239	pCi/g	2.13E-01	4.26E+00	2.13E-01	4.26E+00	3.32E-02	6.64E-01
14119-33-6	Pu-240	pCi/g	2.13E-01	4.26E+00	2.13E-01	4.26E+00	1.75E-02	3.50E-01
14133-76-7	Tc-99	pCi/g	7.60E-03	1.52E-01	7.60E-03	1.52E-01	7.60E-03	1.52E-01
14269-63-7	Th-230	pCi/g	5.79E-02	1.16E+00	1.49E-01	2.98E+00	5.31E-02	1.06E+00
7440-29-1	Th-232	pCi/g	1.16E-01	2.32E+00	1.16E-01	2.32E+00	1.16E-01	2.32E+00
13966-29-5	U-234	pCi/g	1.84E-03	3.67E-02	4.60E-02	9.19E-01	1.09E-03	2.17E-02
15117-96-1	U-235	pCi/g	4.79E-03	9.58E-02	4.11E-02	8.23E-01	4.78E-03	9.57E-02
7440-61-1	U-238	pCi/g	1.06E-03	2.12E-02	4.01E-02	8.03E-01	1.06E-03	2.12E-02

Values in this table for SSLs of 1 and 20 were calculated using the best available information in October 2022. Prior to using the values in this table (in a quantitative risk assessment), a risk assessor must be consulted to determine if any values need to be updated, and to verify that the values are being used appropriately. SSLs calculated using the formula Groundwater NAL*DAF*($K_d+(\theta/\rho)$)/1,000 (EPA 2000).

Where:

Groundwater NAL is the child/adult age-adjusted lifetime resident exposure scenario (see Table A.5b).

DAF is the dilution attenuation factor set at 1 and 20

K_d is the chemical-specific distribution coefficient (see below).

 θ is the porosity set at 0.3

 ρ is the density set at 1.5 g/cm³

K_d values and their references are the following:

Radionuclide	K _d (cm ³ /g)	Reference	Reference Sources
Americium-241	1.90E+03	DOE 2018	(Sheppard and Thibault 1990)
Cesium-137	2.80E+02	DOE 2018	(Sheppard and Thibault 1990)
Neptunium-237	7.00E+01	DOE 2003	Site-specific (DOE 1997b and DOE 2002)
Plutonium-238	5.50E+02	DOE 2003	(Sheppard and Thibault 1990)
Plutonium-239	5.50E+02	DOE 2003	(Sheppard and Thibault 1990)
Plutonium-240	5.50E+02	DOE 2003	(Sheppard and Thibault 1990)
Technetium-99	2.00E-01	DOE 2003	Site-specific (DOE 1997b and DOE 2002)
Thorium-230	3.20E+03	DOE 2003	(Sheppard and Thibault 1990)
Thorium-232	3.20E+03	DOE 2003	(Sheppard and Thibault 1990)
Uranium-234	6.68E+01	DOE 2003	Site-specific (based on uranium-238)
Uranium-235	6.68E+01	DOE 2003	Site-specific (based on uranium-238)
Uranium-238	6.68E+01	DOE 2003	Site-specific (DOE 1997b and DOE 2002)

Table A.8. Dose-Based Soil/Sediment Screening Levels for Site-Related Radionuclides at PGDP (Values were calculated in November 2022 and are based on best available information.)

			Outdoor Worker						
CAS	Radionuclide	Units	1	12	25	100			
Number	Kaulonuchue	Ullits	mrem/year	mrem/year	mrem/year	mrem/year			
14596-10-2	Americium-241	pCi/g	1.37E+01	1.65E+02	3.43E+02	1.37E+03			
10045-97-3	Cesium-137+D	pCi/g	2.17E+00	2.60E+01	5.42E+01	2.17E+02			
13994-20-2	Neptunium-237+D	pCi/g	5.75E+00	6.90E+01	1.44E+02	5.75E+02			
13981-16-3	Plutonium-238	pCi/g	1.28E+01	1.54E+02	3.20E+02	1.28E+03			
15117-48-3	Plutonium-239	pCi/g	1.18E+01	1.41E+02	2.95E+02	1.18E+03			
14119-33-6	Plutonium-240	pCi/g	1.18E+01	1.42E+02	2.95E+02	1.18E+03			
14133-76-7	Technetium-99	pCi/g	4.44E+03	5.32E+04	1.11E+05	4.44E+05			
14269-63-7	Thorium-230	pCi/g	1.40E+01	1.68E+02	3.50E+02	1.40E+03			
7440-29-1	Thorium-232	pCi/g	1.28E+01	1.54E+02	3.20E+02	1.28E+03			
13966-29-5	Uranium-234	pCi/g	6.12E+01	7.35E+02	1.53E+03	6.12E+03			
15117-96-1	Uranium-235+D	pCi/g	9.15E+00	1.10E+02	2.29E+02	9.15E+02			
7440-61-1	Uranium-238+D	pCi/g	3.10E+01	3.72E+02	7.74E+02	3.10E+03			

			Excavation Worker						
CAS	Radionuclide	Units	1	12	25	100			
Number	Kaulonuchue	Ullits	mrem/year	mrem/year	mrem/year	mrem/year			
14596-10-2	Americium-241	pCi/g	1.37E+01	1.65E+02	3.43E+02	1.37E+03			
10045-97-3	Cesium-137+D	pCi/g	2.17E+00	2.60E+01	5.42E+01	2.17E+02			
13994-20-2	Neptunium-237+D	pCi/g	5.75E+00	6.90E+01	1.44E+02	5.75E+02			
13981-16-3	Plutonium-238	pCi/g	1.28E+01	1.54E+02	3.20E+02	1.28E+03			
15117-48-3	Plutonium-239	pCi/g	1.18E+01	1.41E+02	2.95E+02	1.18E+03			
14119-33-6	Plutonium-240	pCi/g	1.18E+01	1.42E+02	2.95E+02	1.18E+03			
14133-76-7	Technetium-99	pCi/g	4.44E+03	5.32E+04	1.11E+05	4.44E+05			
14269-63-7	Thorium-230	pCi/g	1.40E+01	1.68E+02	3.50E+02	1.40E+03			
7440-29-1	Thorium-232	pCi/g	1.28E+01	1.54E+02	3.20E+02	1.28E+03			
13966-29-5	Uranium-234	pCi/g	6.12E+01	7.35E+02	1.53E+03	6.12E+03			
15117-96-1	Uranium-235+D	pCi/g	9.15E+00	1.10E+02	2.29E+02	9.15E+02			
7440-61-1	Uranium-238+D	pCi/g	3.10E+01	3.72E+02	7.74E+02	3.10E+03			

			Industrial Worker						
CAS	Radionuclide	Units	1	12	25	100			
Number	Radionuciuc	Cilits	mrem/year	mrem/year	mrem/year	mrem/year			
14596-10-2	Americium-241	pCi/g	5.29E+01	6.35E+02	1.32E+03	5.29E+03			
10045-97-3	Cesium-137+D	pCi/g	1.62E+00	1.94E+01	4.05E+01	1.62E+02			
13994-20-2	Neptunium-237+D	pCi/g	5.23E+00	6.28E+01	1.31E+02	5.23E+02			
13981-16-3	Plutonium-238	pCi/g	7.17E+01	8.61E+02	1.79E+03	7.17E+03			
15117-48-3	Plutonium-239	pCi/g	6.58E+01	7.90E+02	1.65E+03	6.58E+03			
14119-33-6	Plutonium-240	pCi/g	6.59E+01	7.91E+02	1.65E+03	6.59E+03			
14133-76-7	Technetium-99	pCi/g	2.00E+04	2.40E+05	5.01E+05	2.00E+06			
14269-63-7	Thorium-230	pCi/g	7.76E+01	9.31E+02	1.94E+03	7.76E+03			
7440-29-1	Thorium-232	pCi/g	7.14E+01	8.57E+02	1.79E+03	7.14E+03			
13966-29-5	Uranium-234	pCi/g	3.84E+02	4.60E+03	9.59E+03	3.84E+04			
15117-96-1	Uranium-235+D	pCi/g	7.76E+00	9.31E+01	1.94E+02	7.76E+02			
7440-61-1	Uranium-238+D	pCi/g	4.10E+01	4.92E+02	1.02E+03	4.10E+03			

Table A.8. Dose-Based Soil/Sediment Screening Levels for Site-Related Radionuclides at PGDP (Continued) (Values were calculated in November 2022 and are based on best available information.)

			Adult Recreator								
CAS	Radionuclide	T I : 4	1	12	25	100					
Number	Kadionuciide	Units	mrem/year	mrem/year	mrem/year	mrem/year					
14596-10-2	Americium-241	pCi/g	9.61E+01	1.15E+03	2.40E+03	9.61E+03					
10045-97-3	Cesium-137+D	pCi/g	4.97E+00	5.97E+01	1.24E+02	4.97E+02					
13994-20-2	Neptunium-237+D	pCi/g	1.55E+01	1.86E+02	3.88E+02	1.55E+03					
13981-16-3	Plutonium-238	pCi/g	1.06E+02	1.27E+03	2.65E+03	1.06E+04					
15117-48-3	Plutonium-239	pCi/g	9.75E+01	1.17E+03	2.44E+03	9.75E+03					
14119-33-6	Plutonium-240	pCi/g	9.76E+01	1.17E+03	2.44E+03	9.76E+03					
14133-76-7	Technetium-99	pCi/g	3.21E+04	3.85E+05	8.02E+05	3.21E+06					
14269-63-7	Thorium-230	pCi/g	1.15E+02	1.39E+03	2.89E+03	1.15E+04					
7440-29-1	Thorium-232	pCi/g	1.06E+02	1.27E+03	2.65E+03	1.06E+04					
13966-29-5	Uranium-234	pCi/g	5.12E+02	6.15E+03	1.28E+04	5.12E+04					
15117-96-1	Uranium-235+D	pCi/g	2.33E+01	2.79E+02	5.82E+02	2.33E+03					
7440-61-1	Uranium-238+D	pCi/g	1.11E+02	1.33E+03	2.77E+03	1.11E+04					
				Child 1	Recreator						
CAS	D 11 11 1	TT */	1	12	25	100					
Number	Radionuclide	Units	mrem/year	mrem/year	mrem/year	mrem/year					
14596-10-2	Americium-241	pCi/g	3.15E+01	3.78E+02	7.87E+02	3.15E+03					
10045-97-3	Cesium-137+D	pCi/g	3.69E+00	4.43E+01	9.22E+01	3.69E+02					
13994-20-2	Neptunium-237+D	pCi/g	1.04E+01	1.25E+02	2.61E+02	1.04E+03					
13981-16-3	Plutonium-238	pCi/g	3.01E+01	3.62E+02	7.53E+02	3.01E+03					
15117-48-3	Plutonium-239	pCi/g	2.84E+01	3.41E+02	7.10E+02	2.84E+03					
14119-33-6	Plutonium-240	pCi/g	2.84E+01	3.41E+02	7.10E+02	2.84E+03					
14133-76-7	Technetium-99	pCi/g	4.05E+03	4.86E+04	1.01E+05	4.05E+05					
14269-63-7	Thorium-230	pCi/g	3.01E+01	3.61E+02	7.52E+02	3.01E+03					
7440-29-1	Thorium-232	pCi/g	2.66E+01	3.20E+02	6.66E+02	2.66E+03					
13966-29-5	Uranium-234	pCi/g	1.08E+02	1.29E+03	2.69E+03	1.08E+04					
15117-96-1	Uranium-235+D	pCi/g	1.55E+01	1.86E+02	3.89E+02	1.55E+03					
7440-61-1	Uranium-238+D	pCi/g	5.17E+01	6.21E+02	1.29E+03	5.17E+03					
				Teen I	Recreator						
CAS	D. P P.1.	TT . *4	1	12	25	100					
Number	Radionuclide	Units	mrem/year	mrem/year	mrem/year	mrem/year					
14596-10-2	Americium-241	pCi/g	7.16E+01	8.59E+02	1.79E+03	7.16E+03					
10045-97-3	Cesium-137+D	pCi/g	3.69E+00	4.43E+01	9.23E+01	3.69E+02					
13994-20-2	Neptunium-237+D	pCi/g	1.15E+01	1.38E+02	2.88E+02	1.15E+03					
13981-16-3	Plutonium-238	pCi/g	8.26E+01	9.92E+02	2.07E+03	8.26E+03					
15117-48-3	Plutonium-239	pCi/g	7.56E+01	9.08E+02	1.89E+03	7.56E+03					
14119-33-6	Plutonium-240	pCi/g	7.57E+01	9.09E+02	1.89E+03	7.57E+03					
14133-76-7	Technetium-99	pCi/g	1.95E+04	2.34E+05	4.88E+05	1.95E+06					
14269-63-7	Thorium-230	pCi/g	8.22E+01	9.86E+02	2.05E+03	8.22E+03					
7440-29-1	Thorium-232	pCi/g	7.23E+01	8.68E+02	1.81E+03	7.23E+03					
13966-29-5	Uranium-234	pCi/g	2.54E+02	3.05E+03	6.36E+03	2.54E+04					
15117-96-1	Uranium-235+D	pCi/g	1.69E+01	2.03E+02	4.24E+02	1.69E+03					
7440-61-1	Uranium-238+D	pCi/g	7.50E+01	9.00E+02	1.88E+03	7.50E+03					

Table A.8. Dose-Based Soil/Sediment Screening Levels for Site-Related Radionuclides at PGDP (Continued)

(Values were calculated in November 2022 and are based on best available information.)

Adult Resident CAS 12 100 25 Radionuclide Units Number mre<u>m/year</u> mrem/year mrem/year mrem/year 14596-10-2 Americium-241 pCi/g 2.00E+01 2.40E+02 5.00E+02 2.00E+03 10045-97-3 Cesium-137+D 5.14E-01 6.17E+00 5.14E+01 pCi/g 1.28E+01 pCi/g 13994-20-2 Neptunium-237+D 1.68E+00 2.01E+01 4.19E+01 1.68E+02 13981-16-3 Plutonium-238 pCi/g 3.04E+01 3.65E+02 7.60E+02 3.04E+03 Plutonium-239 15117-48-3 pCi/g 2.79E+013.35E+02 6.97E+022.79E+0314119-33-6 Plutonium-240 pCi/g 2.79E+01 3.35E+02 6.99E+022.79E+03 Technetium-99 14133-76-7 pCi/g 8.22E+04 6.85E+03 1.71E+05 6.85E+05 14269-63-7 Thorium-230 pCi/g 3.27E+01 3.92E+02 8.17E+02 3.27E+03 7440-29-1 Thorium-232 pCi/g 3.02E+01 3.62E+02 7.55E+02 3.02E+03 13966-29-5 Uranium-234 pCi/g 1.47E+02 1.77E+03 3.68E+03 1.47E+04 15117-96-1 Uranium-235+D pCi/g 2.47E+00 2.97E+01 6.18E+01 2.47E+02 Uranium-238+D 1.32E+01 1.59E+02 1.32E+03 7440-61-1 pCi/g 3.31E+02

			Child Resident						
CAS	Radionuclide	Units	1	12	25	100			
Number	Kaulonuchue	Units	mrem/year	mrem/year	mrem/year	mrem/year			
14596-10-2	Americium-241	pCi/g	1.06E+01	1.27E+02	2.64E+02	1.06E+03			
10045-97-3	Cesium-137+D	pCi/g	5.14E-01	6.16E+00	1.28E+01	5.14E+01			
13994-20-2	Neptunium-237+D	pCi/g	1.61E+00	1.94E+01	4.03E+01	1.61E+02			
13981-16-3	Plutonium-238	pCi/g	1.18E+01	1.42E+02	2.96E+02	1.18E+03			
15117-48-3	Plutonium-239	pCi/g	1.11E+01	1.34E+02	2.79E+02	1.11E+03			
14119-33-6	Plutonium-240	pCi/g	1.12E+01	1.34E+02	2.79E+02	1.12E+03			
14133-76-7	Technetium-99	pCi/g	1.52E+03	1.82E+04	3.79E+04	1.52E+05			
14269-63-7	Thorium-230	pCi/g	1.18E+01	1.41E+02	2.94E+02	1.18E+03			
7440-29-1	Thorium-232	pCi/g	1.04E+01	1.25E+02	2.61E+02	1.04E+03			
13966-29-5	Uranium-234	pCi/g	4.25E+01	5.10E+02	1.06E+03	4.25E+03			
15117-96-1	Uranium-235+D	pCi/g	2.38E+00	2.85E+01	5.94E+01	2.38E+02			
7440-61-1	Uranium-238+D	pCi/g	1.07E+01	1.28E+02	2.67E+02	1.07E+03			

Values in this table were calculated using the best available information in November 2022. Prior to using the values in this table (in a quantitative risk assessment), a risk assessor must be consulted to determine if any values need to be updated, and to verify that the values are being used appropriately. Screening levels are based on dose conversion factors from ICRP 72 and ICRP 60 (See Table B.4).

Screening Value = $[\Sigma 1/(Pathway-Specific ALs)]^{-1}$

Pathways include ingestion, inhalation, and external gamma (See Table B.5 for exposure parameters).

Table A.9. Dose-Based Groundwater Screening Levels for Site-Related Radionuclides at PGDP (Values were calculated in November 2022 and are based on best available information.)

			Adult Resident					
CAS			1 mrem/	4 mrem/	12 mrem/	25 mrem/	100 mrem/	
Number	Radionuclide	Units	year	year	year	year	year	
14596-10-2	Americium-241	pCi/L	1.54E+00	6.18E+00	1.85E+01	3.86E+01	1.54E+02	
10045-97-3	Cesium-137+D	pCi/L	2.38E+01	9.50E+01	2.85E+02	5.94E+02	2.38E+03	
13994-20-2	Neptunium-237+D	pCi/L	2.79E+00	1.11E+01	3.34E+01	6.97E+01	2.79E+02	
13981-16-3	Plutonium-238	pCi/L	1.34E+00	5.37E+00	1.61E+01	3.36E+01	1.34E+02	
15117-48-3	Plutonium-239	pCi/L	1.24E+00	4.94E+00	1.48E+01	3.09E+01	1.24E+02	
14119-33-6	Plutonium-240	pCi/L	1.24E+00	4.94E+00	1.48E+01	3.09E+01	1.24E+02	
14133-76-7	Technetium-99	pCi/L	4.82E+02	1.93E+03	5.79E+03	1.21E+04	4.82E+04	
14269-63-7	Thorium-230	pCi/L	1.47E+00	5.88E+00	1.77E+01	3.68E+01	1.47E+02	
7440-29-1	Thorium-232	pCi/L	1.34E+00	5.37E+00	1.61E+01	3.36E+01	1.34E+02	
13966-29-5	Uranium-234	pCi/L	6.31E+00	2.53E+01	7.58E+01	1.58E+02	6.31E+02	
15117-96-1	Uranium-235+D	pCi/L	6.53E+00	2.61E+01	7.84E+01	1.63E+02	6.53E+02	
7440-61-1	Uranium-238+D	pCi/L	6.38E+00	2.55E+01	7.66E+01	1.60E+02	6.38E+02	

			Child Resident					
CAS			1 mrem/	4 mrem/	12 mrem/	25 mrem/	100 mrem/	
Number	Radionuclide	Units	year	year	year	year	year	
14596-10-2	Americium-241	pCi/L	3.67E+00	1.47E+01	4.40E+01	9.17E+01	3.67E+02	
10045-97-3	Cesium-137+D	pCi/L	1.03E+02	4.13E+02	1.24E+03	2.58E+03	1.03E+04	
13994-20-2	Neptunium-237+D	pCi/L	6.91E+00	2.76E+01	8.29E+01	1.73E+02	6.91E+02	
13981-16-3	Plutonium-238	pCi/L	3.19E+00	1.27E+01	3.82E+01	7.96E+01	3.19E+02	
15117-48-3	Plutonium-239	pCi/L	3.00E+00	1.20E+01	3.60E+01	7.51E+01	3.00E+02	
14119-33-6	Plutonium-240	pCi/L	3.00E+00	1.20E+01	3.60E+01	7.51E+01	3.00E+02	
14133-76-7	Technetium-99	pCi/L	4.30E+02	1.72E+03	5.17E+03	1.08E+04	4.30E+04	
14269-63-7	Thorium-230	pCi/L	3.19E+00	1.27E+01	3.82E+01	7.96E+01	3.19E+02	
7440-29-1	Thorium-232	pCi/L	2.82E+00	1.13E+01	3.38E+01	7.04E+01	2.82E+02	
13966-29-5	Uranium-234	pCi/L	1.12E+01	4.49E+01	1.35E+02	2.81E+02	1.12E+03	
15117-96-1	Uranium-235+D	pCi/L	1.15E+01	4.59E+01	1.38E+02	2.87E+02	1.15E+03	
7440-61-1	Uranium-238+D	pCi/L	1.06E+01	4.26E+01	1.28E+02	2.66E+02	1.06E+03	

Values in this table were calculated using the best available information in November 2022. Prior to using the values in this table (in a quantitative risk assessment), a risk assessor must be consulted to determine if any values need to be updated and to verify that the values are being used appropriately. Screening levels are based on dose conversion factors from ICRP 72 and ICRP 60 (See Table B.4).

Screening Value = [\(\Sigma 1.7\)] (Pathway-Specific ALs)]^
Rethwaye include: [\(\Sigma 1.7\)] (Pathway-Specific ALs)]^-

Pathways include ingestion (See Table B.5 for exposure parameters).

Table A.10. Dose-Based Surface Water Screening Levels for Site-Related Radionuclides at PGDP (Values were calculated in October 2022 and are based on best available information.)

Adult Recreator (swimming)

CAS			1 mrem/	4 mrem/	12 mrem/	25 mrem/	100 mrem/
Number	Radionuclide	Units	year	year	year	year	year
14596-10-2	Americium-241	pCi/L	1.26E+02	5.02E+02	1.51E+03	3.14E+03	1.26E+04
10045-97-3	Cesium-137+D	pCi/L	1.93E+03	7.73E+03	2.32E+04	4.83E+04	1.93E+05
13994-20-2	Neptunium-237+D	pCi/L	2.27E+02	9.06E+02	2.72E+03	5.66E+03	2.27E+04
13981-16-3	Plutonium-238	pCi/L	1.09E+02	4.37E+02	1.31E+03	2.73E+03	1.09E+04
15117-48-3	Plutonium-239	pCi/L	1.00E+02	4.02E+02	1.21E+03	2.51E+03	1.00E+04
14119-33-6	Plutonium-240	pCi/L	1.00E+02	4.02E+02	1.21E+03	2.51E+03	1.00E+04
14133-76-7	Technetium-99	pCi/L	3.92E+04	1.57E+05	4.70E+05	9.80E+05	3.92E+06
14269-63-7	Thorium-230	pCi/L	1.20E+02	4.78E+02	1.43E+03	2.99E+03	1.20E+04
7440-29-1	Thorium-232	pCi/L	1.09E+02	4.37E+02	1.31E+03	2.73E+03	1.09E+04
13966-29-5	Uranium-234	pCi/L	5.13E+02	2.05E+03	6.16E+03	1.28E+04	5.13E+04
15117-96-1	Uranium-235+D	pCi/L	5.31E+02	2.12E+03	6.37E+03	1.33E+04	5.31E+04
7440-61-1	Uranium-238+D	pCi/L	5.19E+02	2.08E+03	6.23E+03	1.30E+04	5.19E+04

			Child Recreator (swimming)						
CAS			1 mrem/	4 mrem/	12 mrem/	25 mrem/	100 mrem/		
Number	Radionuclide	Units	year	year	year	year	year		
14596-10-2	Americium-241	pCi/L	8.91E+01	3.56E+02	1.07E+03	2.23E+03	8.91E+03		
10045-97-3	Cesium-137+D	pCi/L	2.51E+03	1.00E+04	3.01E+04	6.27E+04	2.51E+05		
13994-20-2	Neptunium-237+D	pCi/L	1.68E+02	6.72E+02	2.02E+03	4.20E+03	1.68E+04		
13981-16-3	Plutonium-238	pCi/L	7.74E+01	3.10E+02	9.29E+02	1.94E+03	7.74E+03		
15117-48-3	Plutonium-239	pCi/L	7.30E+01	2.92E+02	8.76E+02	1.82E+03	7.30E+03		
14119-33-6	Plutonium-240	pCi/L	7.30E+01	2.92E+02	8.76E+02	1.82E+03	7.30E+03		
14133-76-7	Technetium-99	pCi/L	1.05E+04	4.18E+04	1.26E+05	2.62E+05	1.05E+06		
14269-63-7	Thorium-230	pCi/L	7.74E+01	3.10E+02	9.29E+02	1.94E+03	7.74E+03		
7440-29-1	Thorium-232	pCi/L	6.85E+01	2.74E+02	8.22E+02	1.71E+03	6.85E+03		
13966-29-5	Uranium-234	pCi/L	2.73E+02	1.09E+03	3.28E+03	6.83E+03	2.73E+04		
15117-96-1	Uranium-235+D	pCi/L	2.79E+02	1.12E+03	3.35E+03	6.98E+03	2.79E+04		
7440-61-1	Uranium-238+D	pCi/L	2.59E+02	1.04E+03	3.11E+03	6.47E+03	2.59E+04		

			Teen Recreator (swimming)						
CAS			1 mrem/	4 mrem/	12 mrem/	25 mrem/	100 mrem/		
Number	Radionuclide	Units	year	year	year	year	year		
14596-10-2	Americium-241	pCi/L	7.70E+01	3.08E+02	9.24E+02	1.93E+03	7.70E+03		
10045-97-3	Cesium-137+D	pCi/L	1.18E+03	4.74E+03	1.42E+04	2.96E+04	1.18E+05		
13994-20-2	Neptunium-237+D	pCi/L	1.39E+02	5.55E+02	1.66E+03	3.47E+03	1.39E+04		
13981-16-3	Plutonium-238	pCi/L	7.00E+01	2.80E+02	8.40E+02	1.75E+03	7.00E+03		
15117-48-3	Plutonium-239	pCi/L	6.42E+01	2.57E+02	7.70E+02	1.60E+03	6.42E+03		
14119-33-6	Plutonium-240	pCi/L	6.42E+01	2.57E+02	7.70E+02	1.60E+03	6.42E+03		
14133-76-7	Technetium-99	pCi/L	1.88E+04	7.52E+04	2.26E+05	4.70E+05	1.88E+06		
14269-63-7	Thorium-230	pCi/L	7.00E+01	2.80E+02	8.40E+02	1.75E+03	7.00E+03		
7440-29-1	Thorium-232	pCi/L	6.16E+01	2.46E+02	7.39E+02	1.54E+03	6.16E+03		
13966-29-5	Uranium-234	pCi/L	2.08E+02	8.32E+02	2.50E+03	5.20E+03	2.08E+04		
15117-96-1	Uranium-235+D	pCi/L	2.18E+02	8.73E+02	2.62E+03	5.46E+03	2.18E+04		
7440-61-1	Uranium-238+D	pCi/L	2.17E+02	8.67E+02	2.60E+03	5.42E+03	2.17E+04		

Values in this table were calculated using the best available information in October 2022. Prior to using the values in this table (in a quantitative risk assessment), a risk assessor must be consulted to determine if any values need to be updated and to verify that the values are being used appropriately. Screening levels are based on dose conversion factors from ICRP 72 and ICRP 60 (See Table B.3).

Screening Value = [\(\text{S1}/\text{Pathway-Specific ALs}\)]^{-1}

Pathways include ingestion (See Table B.5 for exposure parameters).

Table A.11. Dose-Based SSLs for Protection of RGA Groundwater for Site-Related Radionuclides at PGDP (Values were calculated in November 2022 and are based on best available information.)

Adult Resident SSL 1 1 mrem/ 4 mrem/ 12 mrem/ 25 mrem/ 100 mrem/ **CAS Number** Radionuclide Units year year year year year 14596-10-2 Americium-241 pCi/g 2.93E+00 1.17E+01 3.52E+01 7.33E+01 2.93E+02 10045-97-3 Cesium-137+D 7.99E+01 pCi/g 6.66E+002.66E+01 1.66E+02 6.66E+0213994-20-2 Neptunium-237+D pCi/g 1.96E-01 7.83E-01 2.35E+00 4.89E+00 1.96E+01 13981-16-3 Plutonium-238 pCi/g 7.39E-01 2.96E+00 8.87E+00 1.85E+01 7.39E+01 Plutonium-239 pCi/g 15117-48-3 6.80E-01 2.72E+00 8.16E+00 1.70E+01 6.80E+01 14119-33-6 Plutonium-240 pCi/g 6.80E-01 2.72E+00 8.16E+00 1.70E+01 6.80E+01 14133-76-7 Technetium-99 pCi/g 1.93E-01 7.72E-01 2.31E+00 4.82E+00 1.93E+01 14269-63-7 Thorium-230 pCi/g 4.71E+00 1.88E+01 5.65E+01 1.18E+024.71E+02 pCi/g 7440-29-1 Thorium-232 4.30E+00 1.72E+01 5.16E+01 1.07E+02 4.30E+02 13966-29-5 Uranium-234 pCi/g 4.23E-01 1.69E+00 5.08E+00 1.06E+01 4.23E+01 15117-96-1 Uranium-235+D pCi/g 4.38E-01 1.75E+00 5.25E+00 1.09E+01 4.38E+01 7440-61-1 Uranium-238+D 1.07E+01 pCi/g 4.28E-01 1.71E+00 5.13E+00 4.28E+01

			Adult Resident SSL 20					
CAS Number	Radionuclide	Units	1 mrem/ year	4 mrem/ year	12 mrem/ year	25 mrem/ year	100 mrem/ year	
14596-10-2	Americium-241	pCi/g	5.87E+01	2.35E+02	7.04E+02	1.47E+03	5.87E+03	
10045-97-3	Cesium-137+D	pCi/g	1.33E+02	5.33E+02	1.60E+03	3.33E+03	1.33E+04	
13994-20-2	Neptunium-237+D	pCi/g	3.91E+00	1.57E+01	4.70E+01	9.78E+01	3.91E+02	
13981-16-3	Plutonium-238	pCi/g	1.48E+01	5.91E+01	1.77E+02	3.69E+02	1.48E+03	
15117-48-3	Plutonium-239	pCi/g	1.36E+01	5.44E+01	1.63E+02	3.40E+02	1.36E+03	
14119-33-6	Plutonium-240	pCi/g	1.36E+01	5.44E+01	1.63E+02	3.40E+02	1.36E+03	
14133-76-7	Technetium-99	pCi/g	3.86E+00	1.54E+01	4.63E+01	9.64E+01	3.86E+02	
14269-63-7	Thorium-230	pCi/g	9.41E+01	3.77E+02	1.13E+03	2.35E+03	9.41E+03	
7440-29-1	Thorium-232	pCi/g	8.60E+01	3.44E+02	1.03E+03	2.15E+03	8.60E+03	
13966-29-5	Uranium-234	pCi/g	8.46E+00	3.38E+01	1.02E+02	2.12E+02	8.46E+02	
15117-96-1	Uranium-235+D	pCi/g	8.75E+00	3.50E+01	1.05E+02	2.19E+02	8.75E+02	
7440-61-1	Uranium-238+D	pCi/g	8.56E+00	3.42E+01	1.03E+02	2.14E+02	8.56E+02	

Table A.11. Dose-Based SSLs for Protection of RGA Groundwater for **Site-Related Radionuclides at PGDP (Continued)**

(Values were calculated in November 2022 and are based on best available information.)

Child Resident SSL 1

CAS Number	Radionuclide	Units	1 mrem/	4 mrem/	12 mrem/	25 mrem/	100 mrem/	
			year	year	year	year	year	
14596-10-2	Americium-241	pCi/g	6.97E+00	2.79E+01	8.36E+01	1.74E+02	6.97E+02	
10045-97-3	Cesium-137+D	pCi/g	2.89E+01	1.16E+02	3.47E+02	7.23E+02	2.89E+03	
13994-20-2	Neptunium-237+D	pCi/g	4.85E-01	1.94E+00	5.82E+00	1.21E+01	4.85E+01	
13981-16-3	Plutonium-238	pCi/g	1.75E+00	7.01E+00	2.10E+01	4.38E+01	1.75E+02	
15117-48-3	Plutonium-239	pCi/g	1.65E+00	6.61E+00	1.98E+01	4.13E+01	1.65E+02	
14119-33-6	Plutonium-240	pCi/g	1.65E+00	6.61E+00	1.98E+01	4.13E+01	1.65E+02	
14133-76-7	Technetium-99	pCi/g	1.72E-01	6.89E-01	2.07E+00	4.30E+00	1.72E+01	
14269-63-7	Thorium-230	pCi/g	1.02E+01	4.08E+01	1.22E+02	2.55E+02	1.02E+03	
7440-29-1	Thorium-232	pCi/g	9.02E+00	3.61E+01	1.08E+02	2.25E+02	9.02E+02	
13966-29-5	Uranium-234	pCi/g	7.53E-01	3.01E+00	9.03E+00	1.88E+01	7.53E+01	
15117-96-1	Uranium-235+D	pCi/g	7.69E-01	3.08E+00	9.23E+00	1.92E+01	7.69E+01	
7440-61-1	Uranium-238+D	pCi/g	7.13E-01	2.85E+00	8.56E+00	1.78E+01	7.13E+01	

		_	Child Resident SSL 20					
CAS Number	Radionuclide	Units	1 mrem/	4 mrem/	12 mrem/	25 mrem/	100 mrem/	
CAS Mullibel	Radionuciuc	Cints	year	year	year	year	year	
14596-10-2	Americium-241	pCi/g	1.39E+02	5.57E+02	1.67E+03	3.48E+03	1.39E+04	
10045-97-3	Cesium-137+D	pCi/g	5.78E+02	2.31E+03	6.94E+03	1.45E+04	5.78E+04	
13994-20-2	Neptunium-237+D	pCi/g	9.70E+00	3.88E+01	1.16E+02	2.43E+02	9.70E+02	
13981-16-3	Plutonium-238	pCi/g	3.51E+01	1.40E+02	4.21E+02	8.76E+02	3.51E+03	
15117-48-3	Plutonium-239	pCi/g	3.30E+01	1.32E+02	3.96E+02	8.26E+02	3.30E+03	
14119-33-6	Plutonium-240	pCi/g	3.30E+01	1.32E+02	3.96E+02	8.26E+02	3.30E+03	
14133-76-7	Technetium-99	pCi/g	3.44E+00	1.38E+01	4.13E+01	8.61E+01	3.44E+02	
14269-63-7	Thorium-230	pCi/g	2.04E+02	8.15E+02	2.45E+03	5.10E+03	2.04E+04	
7440-29-1	Thorium-232	pCi/g	1.80E+02	7.21E+02	2.16E+03	4.51E+03	1.80E+04	
13966-29-5	Uranium-234	pCi/g	1.51E+01	6.02E+01	1.81E+02	3.76E+02	1.51E+03	
15117-96-1	Uranium-235+D	pCi/g	1.54E+01	6.15E+01	1.85E+02	3.85E+02	1.54E+03	
7440-61-1	Uranium-238+D	pCi/g	1.43E+01	5.71E+01	1.71E+02	3.57E+02	1.43E+03	

Values in this table were calculated using the best available information in November 2022 following the methods shown in Table A.7b and the values presented in Table A.9 for the Adult Resident and Child Resident.

SSL 1 indicates the soil screening level calculated for a DAF of 1. SSL 20 indicates the soil screening level calculated for a DAF of 20. SSLs calculated using the formula Dose-Based Groundwater Screening Level*DAF* $(K_d+(\theta/\rho))/1,000$ (EPA 2000).

Dose-Based Groundwater Screening Level is the adult or child resident exposure scenario (see Table A.9). DAF is the dilution attenuation factor set at 1 and 20. K_4 is the chemical-specific distribution coefficient (see below). θ is the porosity set at 0.3. ρ is the density set at 1.5 g/cm^3 . K_4 values and their references are the following:

002)
002)
3)
3)
002)
30 3) 3)

Table A.12. Background Concentrations for Surface and Subsurface Soil at PGDP Background Levels of Selected Radionuclides and Metals in Soil and Geologic Media

Background Levels of Selected Radionuclides and Metals in Soil and Geologic Media at the Paducah Gaseous Diffusion Plant (DOE 1997a)

Analyte Surface Subsurface Inorganic Chemicals (mg/kg)³ 13,000 12,000 Antimony 0.21 0.21 Arsenic 12 7.9 Barium 200 170 Beryllium 0.67 0.69 Cadmium 0.21 0.21 Calcium 200,000 6,100 Chromium (III) 16 43 Chromium (VI)° Cobalt 14 13 Copper 19 25 Cyanide (CN-)³ Iron 28,000 28,000 Lead 36 23 Magnesium 7,700 2,100 Manganese 1,500 820 Mercury 0,2 0,13 Nickel 21 22 Potassium 1,300 950 Selenium 0,8 0,7 Silver 2,3 2,7 Sodium 32,0 340		Background Value ^b					
Aluminum 13,000 12,000 Antimony 0.21 0.21 Arsenic 12 7.9 Barium 200 170 Beryllium 0.67 0.69 Cadmium 0.21 0.21 Calcium 200,000 6,100 Chromium (III) 16 43 Chromium (VI)c Cobalt 14 13 Copper 19 25 Cyanide (CN-)d Iron 28,000 28,000 Lead 36 23 Magnesium 7,700 2,100 Manganese 1,500 820 Mercury 0.2 0.13 Nickel 21 22 Potassium 1,300 950 Selenium 0.8 0,7 Silver 2.3 2,7 Sodium 320 340 Sulfided Ur	Analyte						
Antimony Arsenic Arsenic 12 Arsenic 12 Arsenic 170 Beryllium 200 170 Beryllium 0.67 0.69 Cadmium 0.21 0.21 0.21 Calcium 200,000 6,100 Chromium (III) 16 43 Chromium (VI)c Cobalt 14 13 Copper 19 25 Cyanide (CN-)d Iron 28,000 28,000 Lead 36 23 Magnesium 7,700 2,100 Manganese 1,500 820 Mercury 0,2 0,13 Nickel 21 22 Potassium 1,300 950 Selenium 1,300 950 Selenium 0,8 Silver 2,3 Selenium 0,8 Silver 2,3 Sodium 320 340 Sulfided Thallium 0,21 0,34 Tind Thallium 4,9 4,6 Vanadium 3,8 3,7 Zinc 65 60 Radiouclide (pCi/g) Cesium-137 0,49 0,28 Neptunium-237c Potassium 0,49 0,28 Neptunium-237c 0,11 Plutonium-237c Potassium-40 Radiour-26 1,5 1,5 1,5 1,5 1,5 1,5 1,5 1,5 1,5 1,5	Inorganic Chemicals (mg/kg) ^a						
Arsenic 12 7.9 Barium 2000 170 Beryllium 0.67 0.69 Cadmium 0.21 0.21 Calcium 200,000 6,100 Chromium (III) 16 43 Chromium (VI)° Cobalt 14 13 Copper 19 25 Cyanide (CN-)'d Iron 28,000 28,000 Lead 36 23 Magnesium 7,700 2,100 Manganese 1,500 820 Mercury 0.2 0.13 Nickel 21 22 Potassium 1,300 950 Selenium 0.8 0.7 Silver 2.3 2.7 Sodium 320 340 Sulfide' Thallium 0.21 0.34 Tin'd Uranium 4.9 4.6 Vanadium 38 37 Zinc 65 60 Redionuclide (pCi/g) Cesium-137 0.49 0.28 Neptunium-237° Cesium-137° 0.49 0.28 Neptunium-237° Potassium-238° Potassium-20 1.5 1.5 Neptunium-239° Potassium-40 1.6 1.6 Radium-226 1.5 1.5 Strontium-90° Technetium-99 2.5 2.8 Thorium-228 1.6 1.6 Thorium-230 1.5 1.5	Aluminum	13,000	12,000				
Barium 200 170 Beryllium 0.67 0.69 Cadmium 0.21 0.21 Calcium 200,000 6,100 Chromium (III) 16 43 Chromium (VI)c Cobalt 14 13 Copper 19 25 Cyanide (CN-)d Iron 28,000 28,000 Lead 36 23 Magnesium 7,700 2,100 Manganese 1,500 820 Mercury 0.2 0.13 Nickel 21 22 Potassium 1,300 950 Selenium 0.8 0.7 Silver 2.3 2.7 Sodium 320 340 Sulfided Tind Uranium 4.9 4.6 Vanadium 38 37 Zinc	Antimony	0.21	0.21				
Beryllium 0.67 0.69 Cadmium 0.21 0.21 Calcium 200,000 6,100 Chromium (III) 16 43 Chromium (VI)c Cobalt 14 13 Copper 19 25 Cyanide (CN-)d Iron 28,000 28,000 Lead 36 23 Magnesium 7,700 2,100 Manganese 1,500 820 Mercury 0.2 0.13 Nickel 21 22 Potassium 1,300 950 Selenium 0.8 0.7 Silver 2.3 2.7 Sodium 320 340 Sulfided Tind Uranium 4.9 4.6 Vanadium 38 37 Zine 65 60 Radionulide (pCi/g)	Arsenic	12	7.9				
Beryllium 0.67 0.69 Cadmium 0.21 0.21 Calcium 200,000 6,100 Chromium (III) 16 43 Chromium (VI)c Cobalt 14 13 Copper 19 25 Cyanide (CN-)d Iron 28,000 28,000 Lead 36 23 Magnesium 7,700 2,100 Manganese 1,500 820 Mercury 0.2 0.13 Nickel 21 22 Potassium 1,300 950 Selenium 0.8 0.7 Silver 2.3 2.7 Sodium 320 340 Sulfided Tind Uranium 4.9 4.6 Vanadium 38 37 Zine 65 60 Radionulide (pCi/g)	Barium	200	170				
Cadmium 0.21 0.21 Calcium 200,000 6,100 Chromium (III) 16 43 Chromium (VI)e Cobalt 14 13 Copper 19 25 Cyanide (CN-) ^d Iron 28,000 28,000 Lead 36 23 Magnesium 7,700 2,100 Manganese 1,500 820 Mercury 0.2 0.13 Nickel 21 22 Potassium 1,300 950 Selenium 0.8 0.7 Silver 2.3 2.7 Sodium 320 340 Sulfided Thallium 0.21 0.34 Tind Uranium 4.9 4.6 Vanadium 38 37 Zinc 65 60 Radionuclide (pCi/g) Cesium-137 0.49 0.28	Beryllium	0.67	0.69				
Calcium 200,000 6,100 Chromium (III) 16 43 Chromium (VI)e Cobalt 14 13 Copper 19 25 Cyanide (CN-)d Iron 28,000 28,000 Lead 36 23 Magnesium 7,700 2,100 Manganese 1,500 820 Mercury 0.2 0.13 Nickel 21 22 Potassium 1,300 950 Selenium 0.8 0.7 Silver 2.3 2.7 Sodium 320 340 Sulfided Thallium 0.21 0.34 Tind Vanadium 4.9 4.6 Vanadium 38 37 Zinc 65 60 Radionuclide (pCi/g) Cesium-137 0.49 0.28 Neptunium-238e 0.0025 -		0.21	0.21				
Chromium (III) 16 43 Chromium (VI)° Cobalt 14 13 Copper 19 25 Cyanide (CN-) ^d Iron 28,000 28,000 Lead 36 23 Magnesium 7,700 2,100 Manganese 1,500 820 Mercury 0.2 0.13 Nickel 21 22 Potassium 1,300 950 Selenium 0.8 0.7 Silver 2.3 2.7 Sodium 320 340 Sulfided Thallium 0.21 0.34 Tind Uranium 4.9 4.6 Vanadium 38 37 Zinc 65 60 Radionuclide (pCi/g) Cesium-137 0.49 0.28 Neptunium-238° 0.073 Plutonium-239° 0.025 <t< td=""><td>Calcium</td><td></td><td></td></t<>	Calcium						
Chromium (VI)° Cobalt 14 13 Copper 19 25 Cyanide (CN-)° Iron 28,000 28,000 Lead 36 23 Magnesium 7,700 2,100 Manganese 1,500 820 Mercury 0.2 0.13 Nickel 21 22 Potassium 1,300 950 Selenium 0.8 0.7 Silver 2.3 2.7 Sodium 320 340 Sulfided Tind Uranium 4.9 4.6 Vanadium 38 37 Zinc 65 60 Radionuclide (pCi/g) Cesium-137 0.49 0.28 Neptunium-237° 0.1 Plutonium-238° 0.073 Plutonium-239°	Chromium (III)						
Cobalt 14 13 Copper 19 25 Cyanide (CN-) ^d Iron 28,000 28,000 Lead 36 23 Magnesium 7,700 2,100 Manganese 1,500 820 Mercury 0.2 0.13 Nickel 21 22 Potassium 1,300 950 Selenium 0.8 0.7 Silver 2.3 2.7 Sodium 320 340 Sulfided Thallium 0.21 0.34 Tind Uranium 4.9 4.6 Vanadium 38 37 Zinc 65 60 Radionuclide (pCi/g) Cesium-137 0.49 0.28 Neptunium-237e 0.1 Plutonium-238e 0.073 Plutonium-239e 0.025 Potassium-40 1.6 1.6 </td <td></td> <td></td> <td></td>							
Copper 19 25 Cyanide (CN-) ^d Iron 28,000 28,000 Lead 36 23 Magnesium 7,700 2,100 Manganese 1,500 820 Mercury 0.2 0.13 Nickel 21 22 Potassium 1,300 950 Selenium 0.8 0.7 Silver 2.3 2.7 Sodium 320 340 Sulfide ^d Thallium 0.21 0.34 Tind Uranium 4.9 4.6 Vanadium 38 37 Zinc 65 60 Radionuclide (pCi/g) Cesium-137 0.49 0.28 Neptunium-238 ^e 0.073 Plutonium-239e 0.025 Potassium-40 16 16 Radiom-226 1.5 1.5 Strontium-90e 4.7		14					
Cyanide (CN-) ^d Iron 28,000 28,000 Lead 36 23 Magnesium 7,700 2,100 Manganese 1,500 820 Mercury 0.2 0.13 Nickel 21 22 Potassium 1,300 950 Selenium 0.8 0.7 Silver 2.3 2.7 Sodium 320 340 Sulfide ^d Tind 0.21 0.34 Tind Uranium 4.9 4.6 Vanadium 38 37 Zinc 65 60 Radionuclide (pCi/g) Cesium-137 0.49 0.28 Neptunium-237e 0.1 Plutonium-238e 0.073 Plutonium-239e 0.025 Potassium-40 16 16 Radium-226 1.5 1.5 Strontium-90e 4.7 <td></td> <td></td> <td></td>							
Iron 28,000 28,000 Lead 36 23 Magnesium 7,700 2,100 Manganese 1,500 820 Mercury 0.2 0.13 Nickel 21 22 Potassium 1,300 950 Selenium 0.8 0.7 Silver 2.3 2.7 Sodium 320 340 Sulfided Thallium 0.21 0.34 Tind Uranium 4.9 4.6 Vanadium 38 37 Zinc 65 60 Radionuclide (pCi/g) Cesium-137 0.49 0.28 Neptunium-237e 0.1 Plutonium-238e 0.073 Plutonium-239e 0.025 Potassium-40 16 16 Radium-226 1.5 1.5 Strontium-90e 4.7 Thorium-238 1.6 1.6	Cyanide (CN-)d						
Lead 36 23 Magnesium 7,700 2,100 Manganese 1,500 820 Mercury 0.2 0.13 Nickel 21 22 Potassium 1,300 950 Selenium 0.8 0.7 Silver 2.3 2.7 Sodium 320 340 Sulfided Thallium 0.21 0.34 Tind Uranium 4.9 4.6 Vanadium 38 37 Zinc 65 60 Radionuclide (pCi/g) Cesium-137 0.49 0.28 Neptunium-237e 0.1 Plutonium-238e 0.073 Plutonium-239e 0.025 Potassium-40 16 16 Radium-226 1.5 1.5 Strontium-90e 4.7 Technetium-99 2.5 2.8 Thorium-238 1.6 1.6 <td></td> <td></td> <td></td>							
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Manganese 1,500 820 Mercury 0.2 0.13 Nickel 21 22 Potassium 1,300 950 Selenium 0.8 0.7 Silver 2.3 2.7 Sodium 320 340 Sulfided Thallium 0.21 0.34 Tind Uranium 4.9 4.6 Vanadium 38 37 Zinc 65 60 Radionuclide (pCi/g) Cesium-137 0.49 0.28 Neptunium-237e 0.1 Plutonium-238e 0.073 Plutonium-239e 0.025 Potassium-40 16 16 Radium-226 1.5 1.5 Strontium-90e 4.7 Technetium-99 2.5 2.8 Thorium-238 1.6 1.6 Thorium-230 1.5 1.5 Thorium-232 1.5 <td< td=""><td></td><td></td><td></td></td<>							
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Vanadium 38 37 Zinc 65 60 Radionuclide (pCi/g) Cesium-137 0.49 0.28 Neptunium-237° 0.1 Plutonium-238° 0.073 Potassium-40 16 16 Radium-226 1.5 1.5 Strontium-90° 4.7 Technetium-99 2.5 2.8 Thorium-228 1.6 1.6 Thorium-230 1.5 1.4 Thorium-232 1.5 1.5							
Zinc 65 60 Radionuclide (pCi/g) Cesium-137 0.49 0.28 Neptunium-237° 0.1 Plutonium-238° 0.073 Plutonium-239° 0.025 Potassium-40 16 16 Radium-226 1.5 1.5 Strontium-90° 4.7 Technetium-99 2.5 2.8 Thorium-228 1.6 1.6 Thorium-230 1.5 1.4 Thorium-232 1.5 1.5							
Radionuclide (pCi/g) Cesium-137 0.49 0.28 Neptunium-237° 0.1 Plutonium-238° 0.073 Plutonium-239° 0.025 Potassium-40 16 16 Radium-226 1.5 1.5 Strontium-90° 4.7 Technetium-99 2.5 2.8 Thorium-228 1.6 1.6 Thorium-230 1.5 1.4 Thorium-232 1.5 1.5							
Cesium-137 0.49 0.28 Neptunium-237e 0.1 Plutonium-238e 0.073 Plutonium-239e 0.025 Potassium-40 16 16 Radium-226 1.5 1.5 Strontium-90e 4.7 Technetium-99 2.5 2.8 Thorium-228 1.6 1.6 Thorium-230 1.5 1.4 Thorium-232 1.5 1.5		65	60				
Neptunium-237e 0.1 Plutonium-238e 0.073 Plutonium-239e 0.025 Potassium-40 16 16 Radium-226 1.5 1.5 Strontium-90e 4.7 Technetium-99 2.5 2.8 Thorium-228 1.6 1.6 Thorium-230 1.5 1.4 Thorium-232 1.5 1.5							
Plutonium-238° 0.073 Plutonium-239° 0.025 Potassium-40 16 16 Radium-226 1.5 1.5 Strontium-90° 4.7 Technetium-99 2.5 2.8 Thorium-228 1.6 1.6 Thorium-230 1.5 1.4 Thorium-232 1.5 1.5		0.49	0.28				
Plutonium-239° 0.025 Potassium-40 16 16 Radium-226 1.5 1.5 Strontium-90° 4.7 Technetium-99 2.5 2.8 Thorium-228 1.6 1.6 Thorium-230 1.5 1.4 Thorium-232 1.5 1.5		0.1					
Potassium-40 16 16 Radium-226 1.5 1.5 Strontium-90° 4.7 Technetium-99 2.5 2.8 Thorium-228 1.6 1.6 Thorium-230 1.5 1.4 Thorium-232 1.5 1.5		0.073					
Radium-226 1.5 1.5 Strontium-90° 4.7 Technetium-99 2.5 2.8 Thorium-228 1.6 1.6 Thorium-230 1.5 1.4 Thorium-232 1.5 1.5	Plutonium-239e	0.025					
Strontium-90° 4.7 Technetium-99 2.5 2.8 Thorium-228 1.6 1.6 Thorium-230 1.5 1.4 Thorium-232 1.5 1.5	Potassium-40	16	16				
Technetium-99 2.5 2.8 Thorium-228 1.6 1.6 Thorium-230 1.5 1.4 Thorium-232 1.5 1.5	Radium-226	1.5	1.5				
Thorium-228 1.6 1.6 Thorium-230 1.5 1.4 Thorium-232 1.5 1.5	Strontium-90e	4.7					
Thorium-228 1.6 1.6 Thorium-230 1.5 1.4 Thorium-232 1.5 1.5	Technetium-99	2.5	2.8				
Thorium-230 1.5 1.4 Thorium-232 1.5 1.5							
Thorium-232 1.5 1.5							
Uranium-234 1.2 ^f 1.2 ^f		1.2 ^f	1.2 ^f				
Uranium-235 0.06 ^f 0.06 ^f							
Uranium-238 1.2 1.2							

Notes: Cells with "—" indicate data are not available or not applicable.

Values contained in this table were approved by EPA and the Commonwealth of Kentucky as representative background concentrations; however, the background values have not been approved for all uses by the PGDP Risk Assessment Working Group.

^a Includes inorganic chemicals found on Target Analyte List as defined by EPA in 1988 CLP Statement of Work and RCRA Appendix IX list of constituents

^b Value for use in screening to determine if inorganic chemical or radionuclide detected at naturally occurring concentration in surface or subsurface soil. Details on the derivation of the background concentrations for antimony, beryllium, cadmium, thallium, uranium, and all radionuclides are in DOE 1997a. Details on the derivation of the background concentration for all other inorganic chemicals are in DOE 1996.

^cData are not adequate to calculate a background concentration in soil for this analyte.

^d Cyanide is not expected to be naturally occurring in soil at PGDP; background values were not derived.

^eConcentrations for these radionuclides in subsurface soil were not derived.

^fThe values listed for uranium-234 and uranium-235 are not from the 1996 background study, but are derived from the natural isotopic abundance ratio and the uranium-238 values. The values for these radionuclides that appeared in the 2001 version of the Risk Methods Document (DOE 2001) were the UTLs of measured values for the individual isotopes as reported in the PGDP background study (DOE 1997a).

Table A.13. Background Concentrations for Groundwater Drawn from the RGA

and McNairy Formation at PGDP

Background Concentrations of Naturally Occurring Inorganic Chemicals and Selected Radionuclides in the Regional Gravel Aquifer and
McNairy Formation at the Paducah Gaseous Diffusion Plant, Paducah, Kentucky in
Feasibility Study for the Groundwater Operable Unit at Paducah Gaseous Diffusion Plant Volume 5 (DOE 2000)

	Over All Observations*		Over Wells*	
Analyte	RGA	McNairy	RGA	McNairy
Inorganic Chemicals (µg/L)				
Aluminum	2,189	687	1,640	750
Aluminum, Dissolved	311	579	201	587
Antimony	60 ^a	60ª	60ª	60ª
Antimony, Dissolved	60 ^a	60ª	60ª	60ª
Arsenic	5ª	5ª	5ª	5ª
Arsenic, Dissolved	5ª	5 ^a	5ª	5ª
Barium	235	296	202	265
Barium, Dissolved	200	268	179	266
Beryllium	4ª	17ª	4ª	17ª
Beryllium, Dissolved	4ª	4 ^a	4ª	4a
Cadmium	10 ^a	10 ^a	10 ^a	10 ^a
Cadmium, Dissolved	10 ^a	10 ^a	10ª	10 ^a
Calcium	41,238	38,858	40,000	39,470
Calcium, Dissolved	38,166	38,829	35,800	40,270
Chloride	91,021	19,708	89,200	20,230
Chromium	144	60ª	134	60ª
Chromium, Dissolved	50 ^a	50ª	50ª	50 ^a
Cobalt	45ª	96	45ª	72
Cobalt, Dissolved	45a	45ª	45ª	45ª
Copper	36	57	34	33
Copper, Dissolved	20	13ª	18	13ª
Fluoride	270	330	245	298
Iron	5,030	18,360	3,720	15,830
Iron, Dissolved	267	12,372	164	9,446
Lead	129	50 ^a	250	50 ^a
Lead, Dissolved	98	50a	250	50 ^a
Magnesium	16,262	13,418	15,700	16,457
Magnesium, Dissolved	16,215	14,171	15,400	16,533
Manganese	119	941	82	729
Manganese, Dissolved	68	894	48	682
Mercury	0.2^{a}	0.2^{a}	0.2^{a}	0.2^{a}
Mercury, Dissolved	0.2ª	0.2^{a}	0.2ª	0.2^{a}
Molybdenum	50 ^a	50 ^a	50 ^a	50 ^a
Molybdenum, Dissolved	50 ^a	50 ^a	50 ^a	50 ^a
Nickel	682	109ª	530°	109ª
Nickel, Dissolved	305	50 ^a	305	50 ^a
Nitrate as Nitrogen	15,561	1,474	13,500	1,430
Potassium	5,195	55,752	4,470	64,080
Potassium, Dissolved	4,096	51,205	3,700	58,750
Selenium	5ª	5 ^a	5ª	5 ^a
Selenium, Dissolved	5ª	5 ^a	5ª	5 ^a
Silica	26,401	36,000	21,100	29,400

Table A.13. Background Concentrations for Groundwater Drawn from the RGA and McNairy Formation at PGDP (Continued)

Background Concentrations of Naturally Occurring Lorganic Chemicals and Selected Radionuclides in the Regional Gravel Aquifer and McNairy Formation at the Paducah Gaseous Diffusion Plant, Paducah, Kentucky in Feasibility Study for the Groundwater Operable Unit at Paducah Gaseous Diffusion Plant Volume 5 (DOE 2000)

	Over All Observations*		Over Wells*		
Analyte	RGA	McNairy	RGA	McNairy	
Silver	11ª	50ª	11ª	50ª	
Silver, Dissolved	$60^{\rm a}$	50 ^a	60 ^a	50 ^a	
Sodium	59,450	29,200	63,500	24,920	
Sodium, Dissolved	60,433	27,980	65,700	25,900	
Sulfate	19,947	28,900	19,100	27,270	
Thallium	56ª	644	56ª	255	
Thallium, Dissolved	56ª	56ª	56ª	56 ^a	
Uranium	2^{a}	1ª	2^{a}	1ª	
Uranium, Dissolved	2^{a}	1	2^{a}	1	
Vanadium	134	126	139	119	
Vanadium, Dissolved	134	126	131	107	
Zinc	54	142	25	104	
Zinc, Dissolved	49	116	26	80	
Radionuclides (pCi/L)					
Gross Alpha	5.8	11.9	2.36	5.3	
Gross Beta	13.8	144.5	7.3	125.4	
Neptunium-237	0.8	0.5	0.21	0.13	
Plutonium-239	0.1	0.2	0.03	0.04	
Radium-226	0.6	1.2	0.1	0.29	
Radon-222	626	295	555.3	228.3	
Technetium-99	22.3	20.6	10.8	7.8	
Thorium-230	1.1	1.5	0.54	0.4	
Total Radium	1.3	0.7	0.46	0.36	
Uranium-234 ^b	0.7	0.3	0.7	0.3	
Uranium-235 ^b	0.3	0.2	0.3	0.2	
Uranium-238 ^b	0.7	0.3	0.7	0.3	

Values contained in this table have not been approved for all uses by the PGDP Risk Assessment Working Group; therefore, the values presented here are provisional values and subject to change. The issues to be resolved are the data set from which these values were derived and the statistical methods used to analyze the data set.

Gray shading indicates that the background value is greater than the comparison value for those background values that were derived qualitatively over all observations because the analyte never was detected or was detected infrequently at a concentration near the analyte's detection limit (see Table A.13, note a).

For all projects where averages within groundwater wells over time are considered, the values derived for these groundwater wells under the column heading "over wells" should be used. For all other projects, the values shown under the column heading "over all observations" should be used.

^{*}For inorganic chemicals, background concentrations were derived for both total and filtered samples over all observations within a group (i.e., both groundwater wells and soil boring data) and over only groundwater wells within a group (i.e., only groundwater wells data). For radionuclides, background concentrations were derived using total sample results only because there were too few results from filtered samples.

^a Background value was derived qualitatively over all observations because analyte was never detected or was detected infrequently at a concentration near the analyte's detection limit.

^b Uranium isotopic concentrations were derived from the mass concentration of uranium.

^c Nickel background value varies from Risk Methods Documents prior to 2013 due to an error in calculation in the source document. See PGDP Risk Assessment Working Group Meeting Minutes for December 5, 2012, February 6, 2013, and March 6, 2013 (DOE 2017, Appendix E).

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APPENDIX B DERIVATION OF PRELIMINARY REMEDIATION GOALS



B.1. DERIVATION OF RISK-BASED PRELIMINARY REMEDIATION GOALS

This appendix presents the methods used to derive the direct contact risk-based action levels (ALs) and no action levels (NALs) [i.e., preliminary remediation goals (PRGs)] presented in Appendix A. These PRGs were derived using calculators sponsored and maintained by the U.S. Environmental Protection Agency (EPA) and, where appropriate, Paducah Gaseous Diffusion Plant (PGDP) site-specific parameters. PRGs for chemicals and radionuclides of potential concern (COPCs) were derived using either the Risk Assessment Information System (RAIS) Chemical PRG or the EPA Radionuclide PRG (RAIS 2022, EPA 2022a) online calculator. Links to those sites are as follows:

- https://rais.ornl.gov/cgi-bin/prg/PRG search?select=chem
- https://epa-prgs.ornl.gov/cgi-bin/radionuclides/rprg_search

EPA maintains a separate online calculator for determining regional screening levels for nonradionuclides (EPA 2022b). The link to that site is as follows:

• https://epa-prgs.ornl.gov/cgi-bin/chemicals/csl search

B.1.1 INTRODUCTION

No action and action direct contact risk-based PRGs may be derived using a modification of methods described in Risk Assessment Guidance for Superfund (RAGS), Part B (EPA 1991). In RAGS, Part B, risk-based PRGs are developed by rearranging the equations used to calculate risk or hazard in a risk assessment so that the equations solve for a concentration or activity concentration of an analyte that "yields" a target risk or hazard. To derive the direct contact PRGs, the linear, direct relationship between the concentration or activity concentration of an analyte in an environmental medium and the risk or hazard that exposure to this analyte can present were used. Although this method differs from that in RAGS, Part B, the ultimate results of the modified calculations match those that are received by rearranging the risk or hazard equations. PRGs for vapor intrusion are evaluated consistent with Section 3.3 (see also Appendix E, Section E.9).

B.1.2 MATERIALS

In order to derive risk-based PRGs, several pieces of information are required. These are the receptors of interest, the routes through which the receptors may be exposed and equations describing these routes, carcinogenic (cancer) and noncarcinogenic (hazard) toxicity values, and target risk and hazard values. Each of these is discussed in the following subsections, and they are included on several tables within the appendix. Tables within the subsections that summarize information for deriving risk-based PRGs are as follows:

- Table B.1. Action and No Action Risk-Based Screening Levels for Chemicals and Radionuclides Derived for PGDP by Medium;
- Table B.2. Peak Risk Time Frames for Significant Radionuclides at PGDP;
- Table B.3. Action and No Action Risk-Based Screening Levels and SSLs for Radionuclides Derived for PGDP by Medium;

- Table B.4. Dose Conversion Factors for Radionuclides of Interest;
- Table B.5. Default Exposure Parameters Used in Calculation of RME;
- Table B.6. Toxicity Values and Information;
- Table B.7. Soil Parameters for VF Calculations; and
- Table B.8. Volatilization Parameters.

B.1.2.1 Receptors

Table B.1 provides a matrix showing the medium-receptor combinations for which PRGs were derived. As shown there, over all media, the receptors for which no action and action direct contact risk-based PRGs were derived are the industrial worker; the resident (adult and child); the recreational user (adult, child, and teen); the excavation worker; and the outdoor worker. These receptors were chosen because they represent the most likely current and future receptors for most areas and units at PGDP. Also, it is believed that the PRGs derived for these receptors yield a range of values that is most useful for determining the cleanup priority for the various areas and units at PGDP. Note: Outdoor worker PRGs (used for surface soil) can be used for a construction/excavation worker (used for surface and subsurface soil); however, because the duration and frequency of exposure for a construction/excavation worker would be markedly less than that for an outdoor worker, scenario-specific PRGs for the construction/excavation worker based on site-specific conditions should be derived, as appropriate. The teen recreational user is broken out separately from the adult recreational user, because the exposure parameters for teen recreational users are significantly different than for the adult.

Table B.1 also includes a series of notes that discusses how the PRGs are to be applied to data during site scoping. These notes should be considered before site scoping is attempted.

Table B.1. Action and No Action Risk-Based Screening Levels for Chemicals and Radionuclides Derived for PGDP by Medium

Outdoor Worker Excavation Worker Industrial Worker Adult Recreator		Medium							
	Groundwater	Surface Water	Soil/Sediment						
Outdoor Worker	No	Yes	Yes						
Excavation Worker	No	Yes	Yes						
Industrial Worker	No	Yes	Yes						
Adult Recreator	No	Yes	Yes						
Teen Recreator	No	Yes	Yes						
Child Recreator	No	Yes	Yes						
Adult Resident	Yes	No	Yes						
Child Resident	Yes	No	Yes						

Notes:

Screening values for the residential scenario are used in data screening to develop the list of COPCs in a baseline human health risk assessment (see Section 3.3.3.2 "Procedures to screen or evaluate data to determine COPCs" of the main text for additional information). Additional scenarios/receptors should be used to determine early action screening.

^{2.} All groundwater screening is to be performed using the resident. Of the two receptors (i.e., child and adult), use of the smaller child screening value is more protective of human health when considering noncancer effects. For cancer effects, the lifetime values for the resident, which considers both child and adult exposures, should be used. Note that values for soil deemed protective of groundwater also are available and are based on the resident only.

Table B.1. Action and No Action Risk-Based Screening Levels for Chemicals and Radionuclides Derived for PGDP by Medium (Continued)

- 3. The surface water screening values selected are a location-specific decision. For all areas along effluent ditches or along creeks carrying effluent, the industrial worker screening values are appropriate. Additionally, at areas outside the industrialized areas, use of the recreator values are appropriate. Of the three recreator values available, the child recreator values are the smallest and most protective of human health when considering noncancer effects. For cancer effects, the lifetime values for the recreator, which consider integrated exposure over multiple life stages (e.g., child, teen, adult), should be used. Note that the EPA Radionuclide PRG calculator only allows exposure parameter inputs for child and adult life stages, while the RAIS Chemical PRG online calculator allows exposure parameter inputs for infant, child, adolescent/teen, and adult life stages. Note that two different sets of recreator values are available; these are a set for screening shallow water courses under a wading scenario and a set for screening deeper water courses under a swimming scenario. While which of these two recreator screening values to use is a location-specific decision, general guidance should be to use the wading values for most areas. If exposure by a resident to surface water is of concern, use of the recreator values is appropriate, because rates of contact for the recreator were selected assuming that the individual would be a local resident.
- 4. Determining which soil or sediment screening value is appropriate is a location-specific decision. For all locations inside the industrialized area at PGDP where surface soil contamination is of concern, use of the industrial worker risk-based screening values is appropriate. However, if the scenario involves outdoor maintenance type activities, the outdoor worker risk-based screening values also should be considered. For locations inside the industrialized area at PGDP where contact with surface soil and subsurface soil is of concern (i.e., soil from the surface down to 10 or 16 ft bgs, as appropriate), use of the excavation worker risk-based screening values is appropriate. For locations, outside the industrialized area where surface soil contamination is of concern, screening using the recreator and/or resident risk-based screening values is appropriate. As with the surface water values, the child screening values (for both the recreator and resident) are the most protective of human health when considering noncancer effects. For cancer effects, the lifetime values for the recreator or resident scenarios, which consider exposures across multiple age groupings, should be used. Generally, the recreator risk-based screening values are more appropriate for areas along ditches and creeks (i.e., for bank soils), and the resident risk-based screening values are more appropriate for grassy fields. Finally, the outdoor worker risk-based screening values also can be considered for contact with soil in locations outside the industrialized area if this scenario is appropriate for the locations considered. If screening needs to consider shorter-term exposures to both surface and subsurface soil in locations outside the industrialized area, excavation worker screening values can be used.
- 5. As mentioned above, values for soil for protection of groundwater also are available. These should be used in all areas.

B.1.2.2 Exposure Routes and Equations

The exposure routes considered for the various media-scenario combinations are provided below. Included in this list are the tables from Appendix D that display the equations used to derive chronic daily intake or absorbed dose. The sources for these exposure parameters are provided in the tables in Appendix D. These exposure parameters are summarized in a table at the back of this appendix (Table B.5). Table B.5 also includes a series of notes that provide information on how the exposure parameters shown in the table were used in the RAIS Chemical PRG online calculator and the EPA Radionuclide PRG calculator to derive the PGDP PRGs presented in Appendix A. The equations used in these calculators are similar to the COPC intake equations presented in Appendix D except that they have been rearranged to solve for COPC concentrations in environmental media. The equations in Appendix D should be used to calculate reasonable maximum exposure (RME) intakes in a PGDP baseline human health risk assessment, as shown in Figure B.1.

It is important to note that PRGs are not derived for industrial use of groundwater. These are not derived because they would not be useful to remedial decision making, as indicated in the following material taken from RAGS, Part B, Section 3.2.1 (EPA 1991).

Once ground water is determined to be suitable for drinking, risk-based concentrations should be based on residential exposures....Similarly, for surface water that is to be used for drinking, the risk-based PRGs should be calculated for residential populations, and not simply worker populations.

Note that the number of exposure routes included in these calculations exceeds that presented in RAGS, Part B, for each scenario. Including exposure routes beyond those discussed in RAGS, Part B, is consistent with material in Section 3.1.1 of RAGS, Part B, where it is stated: "Additional exposure pathways (e.g., dermal absorption) are possible and may be significant at some sites for some contaminants, while perhaps only one exposure pathway (e.g., direct ingestion of water only) may be relevant in others. In any case, the risk-based PRG for each chemical should be calculated by considering all of the relevant exposure

pathways." For example, the equations used to derive the radionuclide PRGs do not consider exposure via dermal contact, but they do consider exposure via the external gamma pathway.

Exposure Route	Residential Scenario (Child and Adult)	Industrial Worker Scenario	Outdoor and Excavation Worker Scenarios	Recreational User Scenario (Child, Teen, and Adult)
Groundwater, Chemicals	(Clina and Addit)	Worker Seemand	Worker Section 103	Addity
Ingestion of water	Table D.1	Table D.1*	N/A	N/A
Inhalation of vapors emitted from water during household uses (including	Table D.2	Table D.2*	N/A	N/A
showering) Dermal contact with water during showering	Table D.3	Table D.3*	N/A	N/A
Groundwater, Radionuclides				
Ingestion of water	Table D.1	Table D.1*	N/A	N/A
Soil and Sediment, Chemicals				
Incidental ingestion of contaminated soil or sediment	Table D.4	Table D.4	Table D.4	Table D.4
Dermal contact with contaminated soil or sediment	Table D.5	Table D.5	Table D.5	Table D.5
Inhalation of particulates emitted from soil or sediment	Table D.6	Table D.6	Table D.6	Table D.6
Inhalation of vapors emitted from soil or sediment	Table D.6	Table D.6	Table D.6	Table D.6
Soil and Sediment, Radionuclides				
Incidental ingestion of contaminated soil or sediment	Table D.4	Table D.4	Table D.4	Table D.4
Inhalation of particulates emitted from soil or sediment	Table D.6	Table D.6	Table D.6	Table D.6
Inhalation of vapors emitted from soil or sediment	Table D.6	Table D.6	Table D.6	Table D.6
External exposure to ionizing radiation from soil or sediment	Table D.7	Table D.7	Table D.7	Table D.7
Surface Water, Chemicals				
Dermal contact with contaminated surface water (wading)	Table D.9	Table D.9	N/A	Table D.9
Incidental ingestion of contaminated surface water (swimming)	Table D.8	N/A	N/A	Table D.8
Dermal contact with contaminated surface water (swimming)	Table D.10	N/A	N/A	Table D.10
Surface Water, Radionuclides				
Incidental ingestion of contaminated surface water (swimming)	Table D.8	N/A	N/A	Table D.8

^{*}Because future use of groundwater at PGDP is uncertain, the industrial worker exposure to groundwater scenario is provided for informational purposes only. This hypothetical future exposure pathway (i.e., the industrial worker) should represent in most, if not all, locations an incomplete exposure pathway.

Figure B.1. Equations for Calculating RME intakes in a PGDP Baseline Human Health Risk Assessment

B.1.2.3 Toxicity Values

The toxicity values used in the derivation of the PRGs are taken from a variety of sources. The sources of these values are discussed in Section 3.3.5 of the main text. Table B.6a and Table B.6b (located at the end of this Appendix) provide the toxicity values and other information used in the derivation risk-based PRGs for chemicals and radionuclides, respectively.

N/A = not applicable

B.1.2.4 Target Risk and Hazard Values

The target risk [i.e., target excess lifetime cancer risk (ELCR)] and hazard [i.e., target hazard quotient (HQ)] values used when deriving the PRGs for no action are 1E-06 and 0.1, respectively. The target risk and hazard values used when deriving the PRGs for action are 1E-04 and 3, respectively. Note, if five or more constituents are detected at a site, it may be appropriate during project scoping to reduce the chemical-specific target risk used to derive the PRGs for no action.

B.1.2.5 Information Specific to Risk-Based Radionuclide PRG Derivation

The radionuclide PRGs presented in Appendix A were calculated using the EPA Radionuclide PRG calculator and the PGDP-specific exposure factors provided in Table B.5. Previous versions of the Risk Methods Document used the "+D" designation for four of the COPCs (i.e., Cs-137+D, Np-237+D, U-235+D, U-238+D). For these radionuclides, the PRGs used "+D" slope factors that incorporated the contribution from an ingrowth of progeny for a period of 100 years. The intent of the "+D" slope factors was to make the PRGs protective by accounting for the risk contributions from short-lived decay products that are typically difficult to measure in the environment.

The EPA Radionuclide PRG calculator has been updated in recent years and no longer includes the "+D" slope factors. Instead, the calculator now provides four PRG output options that incorporate different assumptions related to radioactive decay and progeny ingrowth. The radionuclide PRGs presented in Appendix A represent two of the EPA Radionuclide PRG calculator output options (i.e., peak risk, secular equilibrium), which are described in more detail as follows:

- Peak Risk: This option generates a PRG for the parent radionuclide that is based on the time period when the parent and progeny activities present the most risk. The peak risk PRG option assumes that a pure isotope was released and progeny begin ingrowth and decay. The calculator identifies all of the progeny in the decay chain, and each progeny is modeled for decay and progeny ingrowth independently from the parent. The PRG output is the inverse sum of the reciprocal PRGs of the parent and all of the progeny present within the time period of peak risk. For this calculation option, PRGs were calculated assuming the following two different time frames.
 - Infinite Time Frame—This is the default setting for the peak risk PRG option. The calculator uses a Bateman equation solver to determine the parent and progeny activities out to a trillion years.² The radiological activity time curve is converted to risk, and the calculator selects the time period with the most risk from the parent and progeny and then calculates a PRG that corresponds to that time period.
 - 1,000-Year Time Frame—The calculator operates as described in the previous bullet except that it determines the parent and progeny activities out to 1,000 years. The radiological activity time curve is converted to risk, and the calculator selects the time period with the most risk from the parent and progeny and then calculates a PRG that corresponds to that time period. The 1,000-year time frame is consistent with the U.S. Department of Energy's (DOE's) requirements for disposal facilities for uranium and thorium wastes, which must be designed to remain effective for 1,000 years, to the extent reasonably achievable [DOE Order (O) 458.1]. Table B.2

¹ The time period on which the peak risk PRG is based on corresponds to the receptor-specific exposure duration (ED) used in the calculation of the PRG (e.g., industrial worker = 25 years, excavation worker = 5 years). Depending on the parent radionuclide, this time period may occur immediately or thousands of years following the release.

² The Bateman equations are a set of first-order differential equations that describe quantities and activities in a radioactive decay chain as a function of time based on the decay rates and initial quantities.

summarizes the low and high end of the peak risk time periods for the radionuclide and receptor scenarios included in this Risk Methods Document.

• Secular Equilibrium: This option generates a PRG for the parent radionuclide, assuming that the parent and all of the progeny are in secular equilibrium and that the parent is continually being renewed (no source decay). The PRG is based on secular equilibrium of the full chain. For straight chain decay, all of the progeny are at the same activity of the parent radionuclide.

Table B.2. Peak Risk Time Frames for Significant Radionuclides at PGDP

CAS Number	Radionuclide	Peak Risk Period Range (years)
14596-10-2	Am-241	Immediate
10045-97-3	Cs-137	Immediate
13994-20-2	Np-237	535,000–634,000
13981-16-3	Pu-238	Immediate
15117-48-3	Pu-239	Immediate
14119-33-6	Pu-240	Immediate
14133-76-7	Tc-99	Immediate
14269-63-7	Th-230	8,900–8,950
7440-29-1	Th-232	169–196
13966-29-5	U-234	185,000–187,000
15117-96-1	U-235	446,000–448,000
7440-61-1	U-238	3,520,000–3,520,000

B.1.3 METHOD OF DERIVATION

Each risk-based PRG is calculated using the RAIS Chemical PRG online calculator or the EPA Radionuclide PRG calculator. Equations for the derivation for PRGs can be found at the following links.

- https://rais.ornl.gov/tools/rais chemical prg guide.html
- https://epa-prgs.ornl.gov/radionuclides/

B.2. DERIVATION OF DOSE-BASED PRELIMINARY REMEDIATION GOALS FOR RADIONUCLIDES

The following describes the methods used to derive direct-contact dose-based screening levels. Methods for deriving the groundwater protection soil screening levels (SSLs) also are provided for comparison to direct-contact PRGs.

B.2.1 INTRODUCTION

Direct contact dose-based PRGs for radionuclides were derived using a modification of methods described by RAGS, Part B. This modified approach is similar to that used to develop risk-based PRGs for PGDP (except for two additional modifications). These modifications are as follows: (1) the ED and averaging time (AT) terms were dropped because dose limits are based on annual dose and not lifetime exposure, and (2) slope factors and reference doses were replaced with radiation dose conversion factors (DCFs), which represent the radionuclide-specific dose equivalent per unit intake (i.e., ingestion or inhalation) or external

gamma exposure. The DCFs are used to convert radionuclide activity concentrations in environmental media to radiation doses.

B.2.2 MATERIALS

In order to derive dose-based screening levels, several pieces of information are required. These are the receptors of interest, the routes through which the receptors may be exposed, exposure parameters and equations describing these routes, activity concentration-to-DCFs, and target dose values. Each of these is discussed in the following sections.

B.2.2.1 Receptors

The receptors considered in dose-based screening level calculations are described in the derivation of risk-based PRGs. The description is not repeated here, although it is noted that the ED term is not relevant for dose calculations. This is because dose-based values generally call for yearly rather than lifetime values and represent radionuclide-specific activity concentrations in environmental media that would yield the target dose in a given year [e.g., in units of millirem (mrem/year)]. Direct contact screening levels were derived for the industrial worker, the excavation worker, the outdoor worker, the resident (adult and child), and the recreational user (adult, child, and teen). These receptors were chosen because they represent the most likely current and future receptors for most areas and units at PGDP. Also, it is believed that the screening levels derived for these receptors yield a range of values that are most useful for determining the cleanup priority for the various areas and units at PGDP.

Table B.3 lists the media evaluated, by receptor, and includes a series of notes that discuss how the screening levels are to be applied to data during site scoping. These notes should be considered before site scoping is attempted.

Table B.3. Action and No Action Dose-Based Screening Levels for Radionuclides Derived for PGDP by Medium

Caanaria/Dagantar		Medium							
Scenario/Receptor	Groundwater	Surface Water	Soil/Sediment						
Outdoor Worker	No	No	Yes						
Excavation Worker	No	No	Yes						
Industrial Worker	No	No	Yes						
Adult Recreator	No	Yes	Yes						
Teen Recreator	No	Yes	Yes						
Child Recreator	No	Yes	Yes						
Adult Resident	Yes	No	Yes						
Child Resident	Yes	No	Yes						

Notes:

- Screening values for the residential scenario are used in data screening to develop the list of COPCs in a baseline human health risk assessment (see Section 3.3.3.2 "Procedures to screen or evaluate data to determine COPCs" of the main text for additional information). Additional scenarios/receptors should be used to determine early action screening.
- 2. All groundwater screening is to be performed using the resident. Note that values for soil deemed protective of groundwater also are available and are based on the resident only.
- 3. Dose-based values for surface water are provided only for recreators.
- 4. Determining which soil or sediment screening value is appropriate is a location-specific decision. For all locations inside the industrialized area at PGDP where surface soil contamination is of concern, use of the industrial worker dose-based screening values is appropriate. However, if the scenario involves outdoor maintenance type activities, the outdoor worker dose-based screening values also should be considered. For locations inside the industrialized area at PGDP where contact with surface soil and subsurface soil is of concern (i.e., soil from the surface down to 10 or 16 ft bgs, as appropriate), use of the excavation worker dose-based screening values is appropriate. For locations, outside the industrialized area where surface soil contamination is of concern, screening using the recreator and/or resident dose-based screening values is appropriate. As with the surface water values, the child resident dose-based screening values are the smallest and most protective of human health. Generally, the recreator dose-based screening values are more appropriate for areas along ditches and creeks (i.e., for bank soils), and the resident dose-based screening values are more appropriate for grassy fields. Finally, the outdoor worker dose-based screening values also can be considered for

Table B.3. Action and No Action Dose-Based Screening Levels for Radionuclides Derived for PGDP by Medium (Continued)

contact with soil in locations outside the industrialized area if this scenario is appropriate for the locations considered. If screening needs to consider shorter-term exposures to both surface and subsurface soil in locations outside the industrialized area, excavation worker screening values can be used.

- 5. As mentioned above, values for soil for protection of groundwater also are available. These should be used in all areas.
- Screening values for the Outdoor Worker, Excavation Worker, and Industrial Worker do not include exposure to volatiles from groundwater. See Appendix E, Section E.9, for vapor intrusion considerations.

B.2.2.2 Exposure Routes and Equations

The same exposure routes are considered in the equations used to calculate the dose-based PRGs and the risk-based radionuclide PRGs. The equations used to calculate the dose-based PRGs are generally consistent with equations used in the EPA Radionuclide Dose Compliance Concentrations calculator.³ They are similar to what are normally used to calculate the risk-based radionuclide PRGs, except for the removal of the ED and AT terms and the use of DCFs in the place of slope factors—given that the human health-based limits are radiological doses (in units mrem/year) rather than carcinogenic risk. The general equations used to derive the PGDP dose-based PRGs are provided below. For the soil/sediment PRG equations and exposure scenarios, dose-based PRGs were calculated for target doses of 1, 12, 25, and 100 mrem/year. For groundwater and surface water equations and exposure scenarios, dose-based PRGs were calculated for target doses of 1, 4, 12, 25, and 100 mrem/year. DCFs for specific exposure pathways and receptor life stages used are provided in Table B.4. The other exposure parameter values are provided in Table B.5.

Soil/Sediment PRG Equations:

Soil Ingestion PRG_{ing} (pCi/g) = $\frac{TD}{0.001 \text{ g/mg} \times EF \times IR_S \times DCF_{ing}}$ $Particulate\ Inhalation\ PRG_{inh}(_{pCi/g})\ = \frac{_{TD}}{_{EF\times\,ET\times\,InhR\times\left(\frac{1}{PEF}\right)\times\,DCF_{inh}\times\,1000\frac{g}{k_{rr}}}}$ External Dose $PRG_{ext}(pci/g) = \frac{TD}{EF \times (1-S_e) \times T_e \times DCF_{ext}}$ Combined Pathways $PRG_{comb} = \frac{1}{(1/PRG_{ing}) + (1/PRG_{inh}) + (1/PRG_{ext})}$ Where: = target dose (1 mrem/year, 12 mrem/year, 25 mrem/year, and 100 mrem/year) TD EF = exposure frequency (days/year) = soil ingestion rate (mg/day) DCF_{ing} = dose conversion factor—ingestion (mrem/pCi) ET = exposure time (hours/day) InhR = inhalation rate (m³/hour) PEF = particulate emission factor (m³/kg) DCF_{inh} = dose conversion factor—inhalation (mrem/pCi) = gamma shielding factor (unitless) S_{e} = gamma exposure frequency [(days/year) / (days/year)] EF_{ext} = gamma exposure time factor [(hours/day) / (hours/day)] DCF_{ext} = dose conversion factor—external gamma (mrem/year per pCi/g)

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³ https://epa-dccs.ornl.gov/

Groundwater PRG Equation:

Groundwater Ingestion PRG (pCi/L) =
$$\frac{TD}{EF \times IR_{GW} \times DCF_{ing}}$$

Where:

TD = target dose (1 mrem/year, 4 mrem/year, 12 mrem/year, 25 mrem/year, and 100 mrem/year)

EF = exposure frequency (days/year) IR_{GW} = groundwater ingestion rate (L/day)

DCF_{ing} = dose conversion factor - ingestion (mrem/pCi)

Surface Water PRG Equation:

Surface Water Ingestion PRG (pCi/L) =
$$\frac{TD}{EF \times ET \times IR_{SW} \times DCF_{ing}}$$

Where:

TD = target dose (1 mrem/year, 4 mrem/year, 12 mrem/year, 25 mrem/year, and 100 mrem/year)

EF = exposure frequency (days/year) ET = exposure time (hours/day)

 IR_{SW} = surface water ingestion rate (L/hour)

DCF_{ing} = dose conversion factor - ingestion (mrem/pCi)

Dose-Based SSL Equation

Dose-Based SSL (pci/g) = Dose-Based PRG_{GW}× DAF×
$$\left[KD + \left(\frac{\theta}{\rho}\right)\right] \times 0.001 \frac{kg}{g}$$

Where:

 PRG_{GW} = dose-based groundwater PRG for the adult resident (Table A.9)

DAF = dilution attenuation factor set at 1 and 20 (unitless)

Kd = radionuclide-specific distribution coefficient (Lwater/kg) (Table A.9).

 θ = soil porosity set at 0.3 Lwater/Lsoil

ρ = is the density set at 1.5 kg/Lsoil (so θ/ρ = 0.2)

B.2.2.3 Toxicity Values

The toxicity values (i.e., DCFs) used in the derivation of the dose-based concentrations are taken from the latest version of Residual Radioactivity Materials Model (RESRAD) output (i.e., RESRAD 7.2). DCFs are consistent with International Commission on Radiological Protection (ICRP) Publication 60 and Publication 72. The use of ICRP 60 and 72 is consistent with the requirements established by DOE O 458.1. These DCFs are given in unit mrem/pCi for the inhalation and ingestion pathways or mrem/year per pCi/g (i.e., pCi/g in soil/sediment) for the external gamma pathway. The values are provided in Table B.4.

Table B.4. Dose Conversion Factors for Radionuclides of Interest

	Pathway (units)							
	Ingestion ^a	a Inhalation ^a External Gamm						
Radionuclide	(mrem/pCi)	(mrem/pCi)	(mrem/year per pCi/g)					
Adult								
Americium-241	7.40E-04	3.55E-01	3.72E-02					
Cesium-137+D	4.81E-05	1.44E-04	3.38E+00 ^b					
Neptunium-237+D	4.10E-04	1.85E-01	1.01E+00°					
Plutonium-238	8.51E-04	4.07E-01	1.17E-04					

Table B.4. Dose Conversion Factors for Radionuclides of Interest (Continued)

		Pathway (units)	
	Ingestion ^a	Inhalationa	External Gamma ^a
Radionuclide	(mrem/pCi)	(mrem/pCi)	(mrem/year per pCi/g)
Adult (Continued)			
Plutonium-239	9.25E-04	4.44E-01	2.64E-04
Plutonium-240	9.25E-04	4.44E-01	1.13E-04
Technetium-99	2.37E-06	4.81E-05	1.09E-04
Thorium-230	7.77E-04	3.70E-01	1.07E-03
Thorium-232	8.51E-04	4.07E-01	4.56E-04
Uranium-234	1.81E-04	3.48E-02	3.44E-04
Uranium-235+D	1.75E-04	3.15E-02	6.92E-01 ^d
Uranium-238+D	1.79E-04	2.96E-02	1.20E-01°
Teen			•
Americium-241	7.40E-04	3.40E-01	3.72E-02
Cesium-137+D	4.81E-05	1.55E-04	3.38E+00 ^b
Neptunium-237+D	4.11E-04	1.74E-01	1.01E+00°
Plutonium-238	8.14E-04	3.70E-01	1.17E-04
Plutonium-239	8.88E-04	4.07E-01	2.64E-04
Plutonium-240	8.88E-04	4.07E-01	1.13E-04
Technetium-99	3.03E-06	5.55E-05	1.09E-04
Thorium-230	8.14E-04	3.66E-01	1.07E-03
Thorium-232	9.25E-04	4.44E-01	4.56E-04
Uranium-234	2.74E-04	3.70E-02	3.44E-04
Uranium-235+D	2.61E-04	3.40E-02	6.92E-01 ^d
Uranium-238+D	2.63E-04	3.22E-02	1.20E-01°
Child			•
Americium-241	9.99E-04	4.44E-01	3.72E-02
Cesium-137+D	3.55E-05	2.59E-04	3.38E+00 ^b
Neptunium-237+D	5.30E-04	2.22E-01	1.01E+00°
Plutonium-238	1.15E-03	5.18E-01	1.17E-04
Plutonium-239	1.22E-03	5.55E-01	2.64E-04
Plutonium-240	1.22E-03	5.55E-01	1.13E-04
Technetium-99	8.51E-06	8.88E-05	1.09E-04
Thorium-230	1.15E-03	5.18E-01	1.07E-03
Thorium-232	1.30E-03	5.92E-01	4.56E-04
Uranium-234	3.26E-04	7.03E-02	3.44E-04
Uranium-235+D	3.19E-04	6.29E-02	6.92E-01 ^d
Uranium-238+D	3.44E-04	5.93E-02	1.20E-01e

^a From RESRAD version 7.2 output, November 2017. These values are consistent with ICRP 60 and 72, using ages 15 and 5 for the teen and child, respectively.

^b External DCFs for cesium-137+D are calculated by summing external DCFs for cesium-137 and barium-137m. Other DCFs for cesium-137 daughters are not summed.

^c External DCFs for neptunium-237+D are calculated by summing external DCFs for neptunium-237, protactinium-233, and uranium-233.

d External DCFs for uranium-235+D are calculated by summing external DCFs for uranium-235 and thorium-231.

^eExternal DCFs for uranium-238+D are calculated by summing external DCFs for uranium-238, thorium-234, and protactinium-234m.

B.2.2.4 Target Dose Values

The target dose values used when deriving the dose-based concentrations in soil and sediment are 1, 12, 25, and 100 mrem/year. An additional target dose of 4 mrem/year was added for the surface water and groundwater media in consideration of the federal drinking water standard (standards available at https://www.epa.gov/ground-water-and-drinking-water/national-primary-drinking-water-regulations#Radionuclides) although these standards are applicable to public drinking water supplies.

B.2.3 METHOD OF DERIVATION

The dose-based radionuclide PRGs were calculated in Microsoft Excel using the equations provided in Section B.2.2.2. These PRGs are provided in Appendix A, Table A.8 (soil/sediment), Table A.9 (groundwater), Table A.10 (surface water), and Table A.11 (dose-based SSLs).

B.3. EVALUATION FOR LEAD

For sites for which the concentration in soil exceeds the 400 mg/kg screening level, risks from lead should be analyzed using the Integrated Exposure Uptake Biokinetic (IEUBK) model (EPA 2021a, EPA 2021b). The model should be run using the EPA-recommended 5 μ g/dl blood lead level cutoff and any site-specific values. The analysis of risks from lead also should show the probability of exceeding the recommended Commonwealth of Kentucky blood lead level of 2.5 μ g/dl (note that this probability distribution can be developed in the IEUBK model from the previous model run by changing the cutoff value in the graph menu). The uncertainty section of the risk assessment should include text indicating that there is no safe level of lead exposure to children and comparing the risks predicted by the IEUBK analyses based on the two cutoff values.

EPA has published *Recommendations for Sieving Soil and Dust Samples at Lead Sites for Assessment of Incidental Ingestion* (EPA 2016). Historical data not sampled in accordance with EPA's recommendation should be evaluated as uncertain.

The IEUBK model calculates a blood lead level that includes the contribution from off-site sources such as lead in food and water. To make the model more site-specific, regional or site-specific concentrations of lead in food and water, if available, can be used in the model. Site-specific values can be substituted with concurrence from regulatory agencies. At this time, site-specific values are not available for PGDP.

For recreational exposures, the time on-site versus the total time spent outdoors can be included in the model. The model allows only one soil concentration to be entered, but the exposure to on and off-site soil can be incorporated by weighting the soil concentration by the proportion of time spent on and off-site. This method and its limitations are described fully in Appendix A of EPA's review of the human health risk assessment for the Couer d'Alene Basin (EPA 2000).

For industrial or outdoor worker scenarios, the Adult Lead Model is used to develop a PRG for soil (EPA 2017). This model includes a default blood lead level and geometric standard deviation based on the National Health and Nutrition Examination Survey (NHANES) for all races combined, other measured adult blood lead concentrations from state or regional databases may be used in place of the default value if such values are available.

B.4. VOLATILIZATION

Volatilization factors (VFs) are developed for each chemical based on its physical properties. The soil parameters used in the calculation of VFs and the chemical-specific parameters used in the calculation of VFs and the VF values are presented in Tables B.7 and B.8.

Table B.7. Soil Parameters for VF Calculations

Parameter	Definition (units)		Default
Q/C	Inverse the mean conc. at the center of a	Residential	64.177
	0.5 acres square source (g/m ² -s per kg/m ³⁾	Industrial/commercial	43.07
T	Exposure interval (s)		8.2E+08
ρ_{b}	Dry soil bulk density (g/cm³)		1.5
θ_{a}	Air filled soil porosity (L _{air} /L _{soil})		0.28
n	Total soil porosity (L _{pore} /L _{soil})		0.43
θ w	Water-filled soil porosity (L _{water} /L _{soil})		0.15

Information compiled April 2022.

Table B.8. Volatilization Parameters

CAS		Di	Dw	Unitless	Koc
Number	Chemical	(cm ² /s)	(cm ² /s)	Н'	(cm ³ /g)
83-32-9	Acenaphthene	5.06E-02	8.33E-06	7.52E-03	5.03E+03
208-96-8	Acenaphthylene	4.50E-02	6.98E-06	4.66E-03	5.03E+03
107-13-1	Acrylonitrile	1.14E-01	1.23E-05	5.64E-03	8.51E+00
120-12-7	Anthracene	3.90E-02	7.85E-06	2.27E-03	1.64E+04
71-43-2	Benzene	8.95E-02	1.03E-05	2.27E-01	1.46E+02
75-27-4	Bromodichloromethane	5.63E-02	1.07E-05	8.67E-02	3.18E+01
56-23-5	Carbon Tetrachloride	5.71E-02	9.78E-06	1.13E+00	4.39E+01
67-66-3	Chloroform	7.69E-02	1.09E-05	1.50E-01	3.18E+01
75-71-8	Dichlorodifluoromethane (Freon-12)*	7.60E-02	1.08E-05	1.40E+01	4.39E+01
75-34-3	Dichloroethane, 1,1-*	8.36E-02	1.06E-05	2.30E-01	3.18E+01
107-06-2	Dichloroethane, 1,2-	8.57E-02	1.10E-05	4.82E-02	3.96E+01
75-35-4	Dichloroethylene, 1,1-	8.63E-02	1.10E-05	1.07E+00	3.18E+01
540-59-0	Dichloroethylene, 1,2- (Mixed Isomers)	8.79E-02	1.12E-05	1.67E-01	3.96E+01
156-59-2	Dichloroethylene, 1,2-cis-	8.84E-02	1.13E-05	1.67E-01	3.96E+01
156-60-5	Dichloroethylene, 1,2-trans-	8.76E-02	1.12E-05	3.83E-01	3.96E+01
1746-01-6	Dioxins/Furans (Total) (as TCDD)	4.70E-02	6.76E-06	2.04E-03	2.49E+05
100-41-4	Ethylbenzene	6.85E-02	8.46E-06	3.22E-01	4.46E+02
86-73-7	Fluorene	4.40E-02	7.89E-06	3.93E-03	9.16E+03
118-74-1	Hexachlorobenzene	2.90E-02	7.85E-06	6.95E-02	6.20E+03
91-20-3	Naphthalene	6.05E-02	8.38E-06	1.80E-02	1.54E+03
85-01-8	Phenanthrene	3.45E-02	6.69E-06	1.73E-03	1.67E+04
1336-36-3	Polychlorinated Biphenyls	2.43E-02	6.27E-06	1.70E-02	7.81E+04
12674-11-2	~Aroclor 1016	2.54E-02	6.56E-06	8.18E-03	4.77E+04
11104-28-2	~Aroclor 1221	3.25E-02	7.23E-06	9.32E-03	8.40E+03
11141-16-5	~Aroclor 1232	3.34E-02	7.52E-06	3.01E-02	8.40E+03
53469-21-9	~Aroclor 1242	2.39E-02	6.11E-06	1.40E-02	7.81E+04
12672-29-6	~Aroclor 1248	2.41E-02	6.18E-06	1.80E-02	7.65E+04
11097-69-1	~Aroclor 1254	2.37E-02	6.10E-06	1.16E-02	1.31E+05
11096-82-5	~Aroclor 1260	2.20E-02	5.61E-06	1.37E-02	3.50E+05
	cPAHs				
56-55-3	~Benz[a]anthracene	2.61E-02	6.75E-06	4.91E-04	1.77E+05
129-00-0	Pyrene	2.78E-02	7.25E-06	4.87E-04	5.43E+04
127-18-4	Tetrachloroethylene	5.05E-02	9.46E-06	7.24E-01	9.49E+01
108-88-3	Toluene*	7.78E-02	9.20E-06	2.71E-01	2.34E+02
71-55-6	Trichloroethane, 1,1,1-	6.48E-02	9.60E-06	7.03E-01	4.39E+01
79-00-5	Trichloroethane, 1,1,2-	6.69E-02	1.00E-05	3.37E-02	6.07E+01
79-01-6	Trichloroethylene	6.87E-02	1.02E-05	4.03E-01	6.07E+01

Table B.8. Volatilization Parameters (Continued)

CAS		Di	$\mathbf{D}_{\mathbf{w}}$	Unitless	Koc
Number	Chemical	(cm ² /s)	(cm^2/s)	Н'	(cm ³ /g)
76-13-1	Trichloro-1,2,2-trifluoroethane, 1,1,2- (Freon-113)*	3.76E-02	8.59E-06	2.15E+01	1.97E+02
1330-20-7	Xylene, Mixture	6.85E-02	8.46E-06	2.71E-01	3.83E+02
108-38-3	Xylene, m-	6.85E-02	8.46E-06	2.94E-01	3.75E+02
95-47-6	Xylene, o-	6.91E-02	8.56E-06	2.12E-01	3.83E+02
106-42-3	Xylene, P-	6.84E-02	8.45E-06	2.82E-01	3.75E+02

Values taken from RAIS (http://rais.ornl.gov/) in October 2022.

B.5. REFERENCES

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- EPA 2021b. Integrated Exposure Uptake Biokinetic Model for Lead in Children, Windows® Version (IEUBKwin v2), available at https://www.epa.gov/superfund/lead-superfund-sites-software-and-users-manuals, May.
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- EPA 2022b. EPA Regional Screening Level (RSL) Calculator, available at https://epa-prgs.ornl.gov/cgi-bin/chemicals/csl search/ (accessed October 2022).
- RAIS (Risk Assessment Information System) 2022. RAIS website, developed by Oak Ridge National Laboratory and the University of Tennessee, available at https://rais.ornl.gov/ (accessed October 2022).

^{*} Analytes are not PGDP-significant COPCs (Table 2.1), but are provided for project support.

Table B.5. Default Exposure Parameters Used in Calculation of RME

		Default Industrial	Outdoor	Excavation	Adult	Child	Adult Recreational	Teen Recreational	Child Recreational
Pathway Variable	Units	Worker	Worker	Worker	Resident	Resident	User	User	User
General Parameters Used in All Intake Models (unless otherwise noted)									
Exposure frequency (EF)	days/year	250a	185 ^b	185 ^b	350a	350a	104°	140°	140°
Exposure duration (ED)	years	25 ^a	25ª	5 ^b	20a	6 ^a	10 ^a	10 ^a	6 ^a
Body weight (BW)	kg	80 ^a	80 ^a	80a	80 ^a	15a	80 ^a	44 ^d	15 ^a
	year ×	70 ×	70 ×	70 ×	70 ×	70 ×	70 ×	70 ×	70 ×
Averaging time-cancer (AT-C)	days/year	365	365	365	365	365	365	365	365
	year ×	25 ×	25 ×	5 ×	20 ×	6 ×	10 ×	10 ×	6 ×
Averaging time-noncancer (AT-N)	days/year	365	365	365	365	365	365	365	365
		0.693/	0.693/	0.693/	0.693/	0.693/	0.693/	0.693/	0.693/
Decay constant (λ)	unitless	half-life ^e							
					$3 \times (10/70)^{x}$	10×(2/70) x			10×(2/70) x
Age-dependent adjustment factor (ADAF)	unitless	N/A	N/A	N/A	$1 \times (10/70)^{x}$	$3 \times (4/70)^{x}$	1×(10/70) x	3×(10/70) x	3×(4/70) ^x
Ingestion of Water (Table D.1)									
Drinking water ingestion rate (IR)	L/day	1°	N/A	N/A	2.5a	0.78^{a}	N/A	N/A	N/A
Inhalation RGA Groundwater (Table D.2)		•							
Indoor inhalation rate	m ³ /hour	2.5°	N/A	N/A	0.833°	0.833°	N/A	N/A	N/A
Exposure time in the shower (ET _{shower})	hours/day	0.71a	N/A	N/A	0.71a	0.54a	N/A	N/A	N/A
Time of shower (t ₁)	hour	$0.43^{\rm f}$	N/A	N/A	$0.43^{\rm f}$	0.32^{f}	N/A	N/A	N/A
Time after shower (t ₂)	hour	$0.28^{\rm f}$	N/A	N/A	$0.28^{\rm f}$	0.22^{f}	N/A	N/A	N/A
Fraction volatilized while showering (f _{shower})	unitless	0.75^{g}	N/A	N/A	0.75^{g}	0.75^{g}	N/A	N/A	N/A
Water flow rate (F _w)	L/hour	890°	N/A	N/A	890°	890°	N/A	N/A	N/A
Bathroom volume (Va)	m^3	11°	N/A	N/A	11°	11°	N/A	N/A	N/A
	hour/day × year	24×70			24×70	24×70			
Averaging time-cancer (AT-C)	× day/year	× 365	N/A	N/A	× 365	× 365	N/A	N/A	N/A
	hour/day × year	24 × 25			24×20	24 × 6			
Averaging time-noncancer (AT-N)	× day/year	× 365	N/A	N/A	× 365	× 365	N/A	N/A	N/A
Exposure time household use (ET _{house})	hours/day	N/A	N/A	N/A	24°	24°	N/A	N/A	N/A
Exchange rate (ER)	changes/day	N/A	N/A	N/A	10 ^c	10°	N/A	N/A	N/A
Mixing coefficient (MC)	unitless	N/A	N/A	N/A	0.5°	0.5°	N/A	N/A	N/A
Fraction volatilized household use (fhouse)	unitless	N/A	N/A	N/A	0.5°	0.5°	N/A	N/A	N/A
Water flow rate (WHF)	L/day	N/A	N/A	N/A	890°	890°	N/A	N/A	N/A
House volume (HV)	m ³ /change	N/A	N/A	N/A	450°	450°	N/A	N/A	N/A
Dermal Contact with RGA Groundwater (sh									
Body surface area exposed r (SA)	m ²	1.9652a	N/A	N/A	1.9652a	0.6365a	N/A	N/A	N/A
Event time (t _{event})	hours/event	0.71 ^a	N/A	N/A	0.71 ^a	0.54 ^a	N/A	N/A	N/A
Event frequency (EV)	events/day	1	N/A	N/A	1	1	N/A	N/A	N/A
Fraction absorbed (FA)	unitless	1°	N/A	N/A	1°	1°	N/A	N/A	N/A

B-1/

Table B.5. Default Exposure Parameters Used in Calculation of RME (Continued)

Pathway Variable	Units	Default Industrial Worker	Outdoor Worker	Excavation Worker	Adult Resident	Child Resident	Adult Recreational User	Teen Recreational User	Child Recreational User
Incidental Ingestion of Soil/Sediment (Table									
Incidental ingestion rate (IR)	mg/day	50 ^a	480 ^b	480 ^b	100 ^a	200a	100 ^a	100a	200a
Fraction ingested	unitless	1 ^j	1 ^j	1 ^j	1 ^j	1 ^j	1 ^j	1 ^j	1 ^j
Dermal Contact with Soil/Sediment (Table I	0.5)				•				
Body surface area exposed s (SA)	m ² /day	0.3527a	0.3527a	0.3527a	0.6032a	0.2373a	0.6032a	0.75°	0.2373a
Soil-to-skin adherence factor (AF)	mg/cm ² -day	1 ^b	1 ^b	1 ^b	1 ^b	1 ^b	1 ^b	1 ^b	1 ^b
Inhalation of Vapors and Particulates Emitt	ed from Soil/Sed	ment (Tabl	e D.6)						
Exposure time (ET) (soil)	hours/day	8ª	8 ^a	8ª	24ª	24ª	5°	5°	5°
Exposure time (ET) (sediment)	hours/day	2.6°	N/A	N/A	N/A	N/A	N/A	N/A	N/A
	hour/day × year		24×70	24 × 70	24 × 70	24×70	24 × 70	24 × 70	24×70
Averaging time-cancer (AT-C)	× day/year	× 365	× 365	× 365	× 365	× 365	× 365	× 365	× 365
	hour/day × year	24×25	24 × 25	24×25	24×20	24 × 6	24×10	24 × 10	24 × 6
Averaging time-noncancer (AT-N)	× day/year	× 365	× 365	× 365	× 365	× 365	× 365	× 365	× 365
			6.20E+08°						
Particulate emission factor ^t (PEF)	m ³ /kg	1.36E+09	1.36E+09	1.36E+09	1.36E+09	1.36E+09	1.36E+09	1.36E+09	1.36E+09
Inhalation rate for dose-based calcs (InhR)	m ³ /hour	2.5	2.5	2.5	0.833	0.833	2.5	2.5	2.5
External Exposure to Ionizing Radiation fro	m Soil/Sediment	(Table D.7)							
	(days/year)/								
Gamma exposure frequency (EF)	(days/year)	250/365a	185/365 ^b	185/365 ^b	350/365a	350/365a	104/365°	140/365°	140/365°
Gamma shielding factor (Se)	unitless	0.2h	$0.2^{\rm h}$	0.2 ^h	$0.2^{\rm h}$	0.2^{h}	0 i	0 i	0 i
	(hour/day)/								
Gamma exposure time factor (Te) (soil)	(hour/day)	8/24 ^c	8/24 ^c	8/24°	18/24 ^u	18/24 ^u	5/24°	5/24°	5/24°
(1: ()	(hour/day)/	0.6/0.46	0/2/46	0/0/46	10/240	10/240	5 /O Ac	5/0.46	5/0.46
(sediment)	(hour/day)	2.6/24 ^c	8/24 ^c	8/24°	18/24 ^u	18/24 ^u	5/24°	5/24 ^c	5/24°
Incidental Ingestion of Surface Water (swim								1	1
Ingestion rate (IR)	L/hour	N/A	N/A	N/A	N/A	N/A	0.092 ^y	0.150 ^y	0.096 ^y
Exposure time (ET)	hour/day	N/A	N/A	N/A	N/A	N/A	2.6°	2.6°	2.6°
Exposure frequency (EF)	day/year	N/A	N/A	N/A	N/A	N/A	45°	45°	45°
Dermal Contact with Surface Water (wading			•	•	•		•		•
Body surface area exposed ^v (SA)	m ²	0.3527a	0.3527a	0.3527a	N/A	N/A	1.06°	0.75°	0.33°
Exposure frequency (EF)	day/year	250a	20 ^b	20 ^b	N/A	N/A	52°	140°	140°
Exposure time (ET)	hour/day	2.6°	8ª	8 ^a	N/A	N/A	2.6°	2.6°	2.6°
Dermal Contact with Surface Water (swimm									
Body surface area exposed ^w (SA)	m ²	N/A	N/A	N/A	N/A	N/A	1.9652a	1.31°	0.6365a
Exposure frequency (EF)	days/year	N/A	N/A	N/A	N/A	N/A	45°	45°	45°
Exposure time (ET)	hours/day	N/A	N/A	N/A	N/A	N/A	2.6°	2.6°	2.6°
Event (EV)	events/day	N/A	N/A	N/A	N/A	N/A	1	1	1

Table B.5. Default Exposure Parameters Used in Calculation of RME (Continued) a

Pathway Variable	Units	Default Industrial Worker	Outdoor Worker	Excavation Worker	Adult Resident	Child Resident	Adult Recreational User	Teen Recreational User	Child Recreational User
Consumption of Fish (Table D.11)									
Diet fraction (FI)	unitless	N/A	N/A	N/A	N/A	N/A	1 ^j	1 ^j	1 ^j
Ingestion rate ¹ (IR)	kg/day	N/A	N/A	N/A	N/A	N/A	0.029 ^k	0.029^{k}	0.029^{k}
Exposure Frequency (EF)	days/year	N/A	N/A	N/A	N/A	N/A	365	365	365
Consumption of Deer (Table D.13)									
Diet fraction (FI)	unitless	N/A	N/A	N/A	N/A	N/A	1 ^j	1 ^j	1 ^j
Ingestion rate ¹ (IR)	kg/day	N/A	N/A	N/A	N/A	N/A	0.032 ^m	0.032 ^m	$0.007^{\rm m}$
Exposure Frequency (EF)	days/year	N/A	N/A	N/A	N/A	N/A	350 ^m	350 ^m	$350^{\rm m}$
Consumption of Rabbit (Table D.15)		•				•			
Diet fraction (FI)	unitless	N/A	N/A	N/A	N/A	N/A	1 ^j	1 ^j	1 ^j
Ingestion rate ¹ (IR)	kg/meal	N/A	N/A	N/A	N/A	N/A	0.0165 ⁿ	0.0082 ⁿ	$0.0033^{\rm n}$
Exposure Frequency (EF)	meals/year	N/A	N/A	N/A	N/A	N/A	350 ⁿ	350 ⁿ	350 ⁿ
Consumption of Quail (Table D.17)						•			
Diet fraction (FI)	unitless	N/A	N/A	N/A	N/A	N/A	1 ^j	1 ^j	1 ^j
Ingestion rate (IR)	kg/meal	N/A	N/A	N/A	N/A	N/A	0.0047°	0.0024°	0.00094°
Exposure Frequency (EF)	meals/year	N/A	N/A	N/A	N/A	N/A	350°	350°	350°
Consumption of Homegrown Vegetables (Ta	ble D.19)								
Diet fraction (FI)	unitless	N/A	N/A	N/A	0.4 ^p	0.4 ^p	N/A	N/A	N/A
Ingestion rate ¹ (IR)	kg/day	N/A	N/A	N/A	0.72^{q}	0.29^{q}	N/A	N/A	N/A
Consumption of Beef (Table D.21)		•		-		•	<u> </u>		
Diet fraction (FI)	unitless	N/A	N/A	N/A	1 ^j	1 ^j	N/A	N/A	N/A
Ingestion rate ¹ (IR)	kg/day	N/A	N/A	N/A	0.19^{q}	0.07^{q}	N/A	N/A	N/A
Consumption of Milk (Table D.23)						•	<u> </u>		
Diet fraction (FI)	unitless	N/A	N/A	N/A	1 ^j	1 ^j	N/A	N/A	N/A
Ingestion rate ¹ (IR)	kg/day	N/A	N/A	N/A	1.25 ^q	0.9^{q}	N/A	N/A	N/A
Consumption of Poultry (Table D.25)		•		•		•		•	
Diet fraction (FI)	unitless	N/A	N/A	N/A	1 ^j	1 ^j	N/A	N/A	N/A
Ingestion rate ¹ (IR)	kg/day	N/A	N/A	N/A	0.17^{q}	0.07^{q}	N/A	N/A	N/A
Consumption of Pork (Table D.27)		•		•		•		•	
Diet fraction (FI)	unitless	N/A	N/A	N/A	1 ^j	1 ^j	N/A	N/A	N/A
Ingestion rate ¹ (IR)	kg/day	N/A	N/A	N/A	0.08^{q}	0.03^{q}	N/A	N/A	N/A
Consumption of Eggs (Table D.29)									
Diet fraction (FI)	unitless	N/A	N/A	N/A	1 ^j	1 ^j	N/A	N/A	N/A
Ingestion rate ¹ (IR)	kg/day	N/A	N/A	N/A	0.11 ^q	0.06^{q}	N/A	N/A	N/A
Information compiled October 2022									

Information compiled October 2022.

N/A = not applicable

^a EPA 2014, "Human Health Evaluation Manual, Supplemental Guidance: Update of Standard Default Exposure Factors," OSWER Directive 9200.1-120, Assessment and Remediation Division, February 6 (accessed at https://www.epa.gov/sites/production/files/2015-11/documents/oswer_directive_9200.1-120_exposurefactors_corrected2.pdf on October 10, 2017).

^b RAWG Meeting Minutes, September 2014 (see DOE 2017, Appendix E).

Table B.5. Default Exposure Parameters Used in Calculation of RME (Continued)

- ^c KDEP (Kentucky Department for Environmental Protection) 2002. Kentucky Risk Assessment Guidance, Risk Assessment Branch, Kentucky Department for Environmental Protection, Commonwealth of Kentucky.
- d Frederick, T. 2015. U.S. EPA e-mail "RE: Paducah Risk Assessment Working Group: Poll Question re: Paducah-Specific Exposure Parameter," to Garner, L., et al., October 20 (see DOE 2017, Appendix E).
- ^e See the RAIS Web site for additional information (http://rais.ornl.gov/).
- f RAWG Meeting Minutes, June 15, 2016 (see DOE 2017, Appendix E).
- g Value selected by 2009 work group because KDEP (2002) does not specify this value for showering.
- h EPA 1991. Risk Assessment Guidance for Superfund: Volume I-Human Health Evaluation Manual (Part B, Development of Risk-based Preliminary Remediation Goals), OSWER Directive 9285.7-01B.
- RAWG 2007. Discussion on removing gamma shielding factor for recreational receptor, RAWG teleconference call, December (see DOE 2017, Appendix E).
- ^j Maximum Value used; equivalent to 100%.
- k Knuth, B. A., N. A. Connelly, and M. A. Shapiro 1993. Angler Attitudes and Behaviors Associated with Ohio River Health Advisories, Human Dimensions Research Unit (HDRU) Publication 93-6, Department of Natural Resources, New York State College of Agriculture and Life Sciences, Cornell University, Ithaca, NY, 163 p.
- ¹ Ingestion values represent the 95th percentile of individuals who consume this food group.
- ^m Based on taking 2 deer per year (consistent with regulation in the state of Kentucky), a 50% success rate (Kentucky Department of Fish and Wildlife, 1992, Deer Surveys, Project No: W-45-24), a dressed weight averaging 108.5 pounds per deer for Ballard and McCracken counties, 60% of venison recovered per deer carcass, 2.5 persons per household in Ballard and McCracken counties, and a child consumption rate 20% of that for adults. Intake values above correspond to 0.467 g/kg bw-day for the child, 0.744 g/kg bw-day for the teen, and 0.457 g/kg bw-day for the adult receptor.
- "Based on 20 rabbits bagged per year at West Kentucky Wildlife Management Area, a personal communication stating that dressed weight equals 60% of average 1.2 kg rabbit, 2.5 persons per household in Ballard and McCracken counties, a child consumption rate 20% of that for adults, and a teen consumption rate 50% of that for adults. Intake values above correspond to 0.220 g/kg bw-day for the child, 0.191 g/kg bw-day for the teen, and 0.236 g/kg bw-day for the adult receptor.
- Based on 20 quail bagged per year at West Kentucky Wildlife Management Area, personal communication stating dressed weight equals 75% of average 0.183 kg quail, 2.5 persons per household in Ballard and McCracken counties, a child consumption rate 20% of that for adults, and a teen consumption rate 50% of that for adults. Intake values above correspond to 0.063 g/kg bw-day for the child, 0.558 g/kg bw-day for the teen, and 0.067 g/kg bw-day for the adult receptor.
- PEPA 1989. Exposure Factors Handbook, EPA/600/8-89/043.
- ^q EPA 2003. "CSFII Analysis of Food Intake Distributions," EPA/600/R-03/029, Washington, DC.
- ^r Entire surface area of body for both adult and child.
- ^s Includes areas of face, forearms, lower legs, and hands for adults; face, arms, hands, legs, and feet for teens; and face, forearms, hands, lower legs, and feet for children for residents and recreational users. Includes area of hands, arms, and head for workers.
- '.The PEF is the default value used in the RAIS Chemical PRG online calculator, which is based on a calculated air dispersion factor "Q/C_{wind}" value that represents the 90th percentile of Q/C_{wind} values calculated for 29 cities selected to be representative of the range of meteorological conditions across the United States, as described in EPA 2002. Supplemental Guidance for Developing Soil Screening Levels for Superfund Sites, OSWER 9355.4-24, December. For development of the dose-based PRGs, the Outdoor Worker, Excavation Worker, and Industrial Worker scenarios use the PEF value of 6.20E+08 and the Recreator and Resident scenarios use the PEF value of 9.30E+08.
- ^u RAWG Meeting Minutes, December 2012 (see DOE 2017, Appendix E).
- VIncludes areas of arms, hands, legs, and feet for adult, teen, and child for recreational users. Includes area of arms, hands, and head for workers.
- w Includes whole body area for adult, teen, and child.
- * EPA 2005. Supplemental Guidance for Assessing Susceptibility from Early-Life Exposure to Carcinogens, Risk Assessment Forum, EPA/630/R-03/003F, March.
- y EPA 2019. Update for Chapter 3 of the Exposure Factors Handbook: Ingestion of Water and Other Select Liquids, National Center for Environmental Assessment, EPA/600/R-18/259F, February. Values are taken from Table 3-7, the upper percentile for the age group 6 to < 11 for the child, 11 to < 16 for the teen (approximate), and 21+ for the adult.

Notes on use of Table B.5 exposure parameters in the calculation of ALs/NALs.

Chemical COPCs:

- Industrial Worker (soil and sediment)—The ALs/NALs for the industrial worker scenario were calculated using the Composite Worker module in the RAIS Chemical PRG online calculator, with the relevant exposure parameters listed in the table above.
- Outdoor and Excavation Worker (soil and sediment)—The ALs/NALs for the outdoor worker and excavation worker scenarios were calculated using the Outdoor Worker module in the RAIS Chemical PRG online calculator, with the relevant exposure parameters listed in the table above.
- Industrial, Excavation, and Outdoor Worker (wading in surface water)—The ALs/NALs for the industrial worker, outdoor worker, and excavation worker scenarios were calculated using the Recreator module in the RAIS Chemical PRG online calculator, with the relevant exposure parameters listed in the table above. The Recreator module in the RAIS Chemical PRG online calculator is a site-specific option that allows for input of receptor-specific exposure parameter values.
- Resident (soil, sediment, and groundwater)—The ALs/NALs for the child, adult, and age-adjusted lifetime resident scenarios were calculated using the Resident module in the RAIS Chemical PRG online calculator, with the relevant exposure parameters listed in the table above.
- Recreational User (soil, sediment, and surface water)—The ALs/NALs for the child, adult, and age-adjusted lifetime recreational user scenarios were calculated using the Recreator module in the RAIS Chemical PRG online calculator, with the relevant exposure parameters listed in the table above.

Radionuclide COPCs:

- Industrial, Excavation, and Outdoor Worker (soil and sediment)—The ALs/NALs for the industrial worker, outdoor worker, and excavation worker scenarios were calculated using the Indoor Worker module in the EPA Radionuclide PRG calculator, with the relevant exposure parameters listed in the table above.
- Resident (soil, sediment, and groundwater)—The ALs/NALs for the age-adjusted lifetime resident scenario were calculated using the Resident module in the EPA Radionuclide PRG calculator, with the relevant child and adult exposure parameters listed in the table above. The EPA Radionuclide PRG calculator does not allow entry of the exposure parameters for the teen life stage shown in the table above.
- Recreational User (soil, sediment, and surface water)—The ALs/NALs for the age-adjusted lifetime recreational user scenario were calculated using the Recreator module in the EPA Radionuclide PRG calculator with the relevant child and adult exposure parameters listed in the table above. The Radionuclide PRG calculator does not allow entry of the exposure parameters for the teen life stage shown in the table above.

Table B.6a. Toxicity Values and Information¹

Chemical Abstract Number	Analyte	Oral Slope Factor (SFo) ²	Sfo Ref	Inhalation Unit Risk (IUR) ³	IUR Ref	Oral RfD (RfDo) ⁴	RfDo Ref	Inhalation (RfCi) ⁵	RfCi Ref	Volatile Organic? ⁶	Muta- gen? ⁷	Gastro intestinal (GI) Absorption Factor (Unitless)	EPA ABS (Unitless)8	ABS Ref
7429-90-5	Aluminum	-	-	-	-	1.00E+00	P	5.00E-03	P	NO	NO	1.00E+00	-	RAGSE
7440-36-0	Antimony (metallic)	-	-	-	-	4.00E-04	I	3.00E-04	A	NO	NO	1.50E-01	-	RAGSE
7440-38-2	Arsenic, Inorganic	1.50E+00	I	4.30E-06	I	3.00E-04	I	1.50E-05	С	NO	NO	1.00E+00	3.00E-02	RAGSE
7440-39-3	Barium	-	-	-	-	2.00E-01	I	5.00E-04	Н	NO	NO	7.00E-02	-	RAGSE
7440-41-7	Beryllium and compounds	-	-	2.40E-06	I	2.00E-03	I	2.00E-05	I	NO	NO	7.00E-03	-	RAGSE
7440-42-8	Boron And Borates Only	-	-	-	-	2.00E-01	I	2.00E-02	Н	NO	NO	1.00E+00	-	RAGSE
7440-43-9	Cadmium (Diet)	-	-	1.80E-06	I	1.00E-04	I	1.00E-05	A	NO	NO	2.50E-02	1.00E-03	RAGSE
7440-43-9	Cadmium (Water)	-	-	1.80E-06	I	5.00E-04	I	1.00E-05	A	NO	NO	5.00E-02	1.00E-03	RAGSE
7440-47-3	Chromium (Total) ^a	-	-	-	-	-	-	-	-	NO	NO	1.30E-02	-	-
16065-83-1	Chromium(III), Insoluble Salts	-	-	-	-	1.50E+00	I	-	-	NO	NO	1.30E-02	-	RAGSE
18540-29-9	Chromium(VI)	5.00E-01	С	8.40E-05	S	3.00E-03	I	1.00E-04	I	NO	YES	2.50E-02	-	RAGSE
7440-48-4	Cobalt	-	-	9.00E-06	P	3.00E-04	P	6.00E-06	P	NO	NO	1.00E+00	-	RAGSE
7440-50-8	Copper	-	-	-	-	4.00E-02	Н	-	-	NO	NO	1.00E+00	-	RAGSE
16984-48-8	Fluoride	-	-	-	-	4.00E-02	С	1.30E-02	С	NO	NO	1.00E+00	-	RAGSE
7439-89-6	Iron	-	-	-	-	7.00E-01	P	-	-	NO	NO	1.00E+00	-	RAGSE
7439-92-1	Lead ^e	-	-	-	-	-	-	-	-	NO	NO	-	-	-
7439-96-5	Manganese (Non-Diet)	-	-	-	-	2.40E-02	S	5.00E-05	I	NO	NO	4.00E-02	-	RAGSE
various	Mercury, Inorganic Salts	-	-	-	-	3.00E-04	S	-	-	NO	NO	7.00E-02	-	RAGSE
7439-98-7	Molybdenum	-	-	-	-	5.00E-03	I	2.00E-03	A	NO	NO	1.00E+00	-	RAGSE
7440-02-0	Nickel Soluble Salts	-	-	2.60E-07	С	2.00E-02	I	9.00E-05	A	NO	NO	4.00E-02	-	RAGSE
7782-49-2	Selenium	-	-	-	-	5.00E-03	I	2.00E-02	С	NO	NO	1.00E+00	-	RAGSE
7440-22-4	Silver	-	-	-	-	5.00E-03	I	-	-	NO	NO	4.00E-02	-	RAGSE
7440-28-0	Thallium (Soluble Salts)	-	-	-	-	1.00E-05	PS	-	-	NO	NO	1.00E+00	-	RAGSE
N/A	Uranium (Insoluble Compounds)	-	-	-	-	3.00E-03	I	4.00E-05	A	NO	NO	1.00E+00	-	RAGSE
N/A	Uranium (Soluble Salts)	-	-	-	-	2.00E-04	A	4.00E-05	A	NO	NO	1.00E+00	-	RAGSE
7440-62-2	Vanadium and Compounds	-	-	-	-	5.04E-03	S	1.00E-04	A	NO	NO	2.60E-02	-	RAGSE
7440-66-6	Zinc and Compounds	-	-	-	-	3.00E-01	I	-	-	NO	NO	1.00E+00	-	RAGSE
83-32-9	Acenaphthene	-	-	-	-	6.00E-02	I	-	-	YES	NO	1.00E+00	1.30E-01	RAGSE
208-96-8	Acenaphthylene ^b	-	-	-	-	-	-	-	-	YES	NO	1.00E+00	1.30E-01	RAGSE
107-13-1	Acrylonitrile	5.40E-01	I	6.80E-08	I	1.00E-02	Α	2.00E-03	I	YES	NO	1.00E+00	-	RAGSE
120-12-7	Anthracene	-	-	-	-	3.00E-01	I	-	-	YES	NO	1.00E+00	1.30E-01	RAGSE
71-43-2	Benzene	5.50E-02	I	7.80E-09	I	4.00E-03	I	3.00E-02	I	YES	NO	1.00E+00	-	RAGSE
117-81-7	Bis(2-ethylhexyl)phthalate ^c	1.40E-02	I	2.40E-09	С	2.00E-02	I	-	-	NO	NO	1.00E+00	1.00E-01	RAGSE
75-27-4	Bromodichloromethane	6.20E-02	I	3.70E-08	С	8.00E-03	P	-	-	YES	NO	1.00E+00	-	RAGSE
86-74-8	Carbazole	2.00E-02	Н	-	-	-	_	-	-	NO	NO	1.00E+00	1.00E-01	RAGSE
56-23-5	Carbon Tetrachloride	7.00E-02	I	6.00E-09	I	4.00E-03	I	1.00E-01	I	YES	NO	1.00E+00	-	RAGSE
67-66-3	Chloroform	3.10E-02	С	2.30E-08	I	1.00E-02	I	9.77E-02	A	YES	NO	1.00E+00	-	RAGSE
75-71-8	Dichlorodifluoromethane (Freon-12) ^c	_	i _	_	-	2.00E-01	I	1.00E-01	PS	YES	NO	1.00E+00	-	RAGSE

Table B.6a. Toxicity Values and Information (Continued)

Chemical Abstract		Oral Slope Factor	Sfo	Inhalation Unit Risk	IUR	Oral RfD	RfDo	Inhalation	RfCi	Volatile	Muta	Gastro intestinal (GI) Absorption Factor	EPA ABS	ABS
Number	Analyte	(SFo) ²	Ref	(IUR) ³	Ref	(RfDo) ⁴	Ref	(RfCi) ⁵		Organic?6	gen? ⁷	(Unitless)	(Unitless) ⁸	Ref
75-34-3	Dichloroethane, 1,1-°	5.70E-03	C	1.60E-09	С	2.00E-01	P	5.00E-01	Н	YES	NO	1.00E+00	-	RAGSE
107-06-2	Dichloroethane, 1,2-	9.10E-02	I	2.60E-08	I	6.00E-03	PS	7.00E-03	P	YES	NO	1.00E+00	-	RAGSE
75-35-4	Dichloroethylene, 1,1-	-	-	-	-	5.00E-02	I	2.00E-01	I	YES	NO	1.00E+00	-	RAGSE
540-59-0	Dichloroethylene, 1,2- (Mixed Isomers)	-	-	-	-	9.00E-03	Н	-	-	YES	NO	1.00E+00	-	RAGSE
156-59-2	Dichloroethylene, 1,2-cis-	-	-	-	-	2.00E-03	I	4.00E-02	PS	YES	NO	1.00E+00	-	RAGSE
156-60-5	Dichloroethylene, 1,2-trans-	-	-	-	-	2.00E-02	I	4.00E-02	PS	YES	NO	1.00E+00	-	RAGSE
60-57-1	Dieldrin	1.60E+01	I	4.60E-06	I	5.00E-05	I	-	-	NO	NO	1.00E+00	1.00E-01	RAGSE
1746-01-6	Dioxins/Furans, Total ^f	-	-	-	-	-	-	-	-	-	-	-	-	-
37871-00-4	~HpCDD	1.30E+03	W	3.80E-04	W	7.00E-08	W	4.00E-06	W	YES	NO	1.00E+00	3.00E-02	RAGSE
38998-75-3	~HpCDF, 2,3,7,8-	1.30E+03	W	3.80E-04	W	7.00E-08	W	4.00E-06	W	YES	NO	1.00E+00	3.00E-02	RAGSE
34465-46-8	~HxCDD, 2,3,7,8-	1.30E+04	W	3.80E-03	W	7.00E-09	W	4.00E-07	W	NO	NO	1.00E+00	3.00E-02	RAGSE
55684-94-1	~HxCDF, 2,3,7,8-	1.30E+04	W	3.80E-03	W	7.00E-09	W	4.00E-07	W	NO	NO	1.00E+00	3.00E-02	RAGSE
3268-87-9	~OCDD	3.90E+01	W	1.14E-05	W	2.33E-06	W	1.33E-04	W	NO	NO	1.00E+00	3.00E-02	RAGSE
39001-02-0	~OCDF	3.90E+01	W	1.14E-05	W	2.33E-06	W	1.33E-04	W	NO	NO	1.00E+00	3.00E-02	RAGSE
36088-22-9	~PeCDD, 2,3,7,8-	1.30E+05	W	3.80E-02	W	7.00E-10	W	4.00E-08	W	NO	NO	1.00E+00	3.00E-02	RAGSE
57117-41-6	~PeCDF, 1,2,3,7,8-	3.90E+03	W	1.14E-03	W	2.33E-08	W	1.33E-06	W	NO	NO	1.00E+00	3.00E-02	RAGSE
57117-31-4	~PeCDF, 2,3,4,7,8-	3.90E+04	W	1.14E-02	W	2.33E-09	W	1.33E-07	W	NO	NO	1.00E+00	3.00E-02	RAGSE
1746-01-6	~TCDD, 2,3,7,8-	1.30E+05	С	3.80E-02	С	7.00E-10	I	4.00E-08	С	YES	NO	1.00E+00	3.00E-02	RAGSE
51207-31-9	~TCDF, 2,3,7,8-	1.30E+04	W	3.80E-03	W	7.00E-09	W	4.00E-07	W	YES	NO	1.00E+00	3.00E-02	RAGSE
100-41-4	Ethylbenzene	1.10E-02	С	2.50E-09	С	5.00E-02	P	1.00E+00	I	YES	NO	1.00E+00	-	RAGSE
206-44-0	Fluoranthene	-	-	-	1	4.00E-02	I	-	-	NO	NO	1.00E+00	1.30E-01	RAGSE
86-73-7	Fluorene	-	-	-	-	4.00E-02	I	-	-	YES	NO	1.00E+00	1.30E-01	RAGSE
118-74-1	Hexachlorobenzene	1.60E+00	I	4.60E-07	I	1.00E-05	P	-	-	YES	NO	1.00E+00	-	RAGSE
91-20-3	Naphthalene	1.20E-01	С	3.40E-08	С	2.00E-02	I	3.00E-03	I	YES	NO	1.00E+00	1.30E-01	RAGSE
88-74-4	Nitroaniline, 2-	-	-	-	1	1.00E-02	PS	5.00E-05	PS	NO	NO	1.00E+00	1.00E-01	RAGSE
621-64-7	Nitroso-di-N-propylamine, N-	7.00E+00	I	2.00E-06	С	-	-	-	-	NO	NO	1.00E+00	1.00E-01	RAGSE
87-86-5	Pentachlorophenol	4.00E-01	I	5.10E-09	С	5.00E-03	I	-	-	NO	NO	1.00E+00	2.50E-01	RAGSE
85-01-8	Phenanthrene ^b	-	-	-	1	-	-	-	-	YES	NO	1.00E+00	1.30E-01	RAGSE
1336-36-3	Polychlorinated Biphenyls (high risk)	2.00E+00	I	5.71E-07	I	-	-	-	-	YES	NO	1.00E+00	1.40E-01	RAGSE
1336-36-3	Polychlorinated Biphenyls (low risk)	4.00E-01	I	1.00E-07	I	-	-	-	-	YES	NO	1.00E+00	1.40E-01	RAGSE
12674-11-2	~Aroclor 1016	7.00E-02	S	2.00E-08	S	7.00E-05	I	-	-	YES	NO	1.00E+00	1.40E-01	RAGSE
11104-28-2	~Aroclor 1221	2.00E+00	S	5.71E-07	S	-	-	-	-	YES	NO	1.00E+00	1.40E-01	RAGSE
11141-16-5	~Aroclor 1232	2.00E+00	S	5.71E-07	S	-	-	-	-	YES	NO	1.00E+00	1.40E-01	RAGSE
53469-21-9	~Aroclor 1242	2.00E+00	S	5.71E-07	S	-	-	-	-	YES	NO	1.00E+00	1.40E-01	RAGSE
12672-29-6	~Aroclor 1248	2.00E+00	S	5.71E-07	S	-	-	-	-	YES	NO	1.00E+00	1.40E-01	RAGSE
11097-69-1	~Aroclor 1254	2.00E+00	S	5.71E-07	S	2.00E-05	I	-	-	YES	NO	1.00E+00	1.40E-01	RAGSE
11096-82-5	~Aroclor 1260	2.00E+00	S	5.71E-07	S	-	-	-	-	YES	NO	1.00E+00	1.40E-01	RAGSE

Table B.6a. Toxicity Values and Information (Continued)

Chemical Abstract Number	Analyte	Oral Slope Factor (SFo) ²	Sfo Ref	Inhalation Unit Risk (IUR) ³	IUR Ref	Oral RfD (RfDo) ⁴	RfDo Ref	Inhalation (RfCi) ⁵	RfCi Ref	Volatile Organic? ⁶	Muta gen? ⁷	Gastro intestinal (GI) Absorption Factor (Unitless)	EPA ABS (Unitless) ⁸	ABS Ref
	Polycyclic aromatic hydrocarbons													
50-32-8	(cPAH), Total Carcinogenic ^f	-	-	-	-	-	-	-	-	-	-	-	-	-
56-55-3	~Benz[a]anthracene	1.00E-01	Е	6.00E-08	Е	-	-	-	-	YES	YES	1.00E+00	1.30E-01	RAGSE
50-32-8	~Benzo[a]pyrene	1.00E+00	I	6.00E-07	I	3.00E-04	I	2.00E-06	I	NO	YES	1.00E+00	1.30E-01	RAGSE
205-99-2	~Benzo[b]fluoranthene	1.00E-01	Е	6.00E-08	Е	-	-	-	-	NO	YES	1.00E+00	1.30E-01	RAGSE
207-08-9	~Benzo[k]fluoranthene	1.00E-02	Е	6.00E-09	Е	-	-	-	-	NO	YES	1.00E+00	1.30E-01	RAGSE
218-01-9	~Chrysene	1.00E-03	Е	6.00E-10	Е	-	-	-	-	NO	YES	1.00E+00	1.30E-01	RAGSE
53-70-3	~Dibenz[a,h]anthracene	1.00E+00	E	6.00E-07	E	-	-	-	-	NO	YES	1.00E+00	1.30E-01	RAGSE
193-39-5	~Indeno[1,2,3-cd]pyrene	1.00E-01	Е	6.00E-08	Е	-	-	-	-	NO	YES	1.00E+00	1.30E-01	RAGSE
129-00-0	Pyrene	-	-	-	-	3.00E-02	I	-	-	YES	NO	1.00E+00	1.30E-01	RAGSE
127-18-4	Tetrachloroethylene	2.10E-03	I	2.60E-10	I	6.00E-03	I	4.00E-02	I	YES	NO	1.00E+00	-	RAGSE
108-88-3	Toluene ^c	-	-	-	1	8.00E-02	I	5.00E+00	I	YES	NO	1.00E+00	1	RAGSE
71-55-6	Trichloroethane, 1,1,1-	-	-	-	-	2.00E+00	I	5.00E+00	I	YES	NO	1.00E+00	-	RAGSE
79-00-5	Trichloroethane, 1,1,2-	5.70E-02	I	1.60E-08	I	4.00E-03	I	2.00E-04	PS	YES	NO	1.00E+00	-	RAGSE
79-01-6	Trichloroethylene	4.60E-02	I	4.10E-09	I	5.00E-04	T	2.00E-03	T	YES	YES	1.00E+00	-	RAGSE
76-13-1	Trichloro-1,2,2-trifluoroethane, 1,1,2- (Freon-113) ^c	-	-	-	1	3.00E+01	I	5.00E+00	P	YES	NO	1.00E+00	-	RAGSE
75-01-4	Vinyl Chloride	7.20E-01	I	4.40E-09	I	3.00E-03	I	8.00E-02	A	YES	YES	1.00E+00	1	RAGSE
1330-20-7	Xylene, Mixture	-	-	-	-	2.00E-01	I	1.00E-01	I	YES	NO	1.00E+00	-	RAGSE
108-38-3	Xylene, m-	-	-	-	-	2.00E-01	S	1.00E-01	S	YES	NO	1.00E+00	-	RAGSE
95-47-6	Xylene, o-	-	-	-	-	2.00E-01	S	1.00E-01	S	YES	NO	1.00E+00	1	RAGSE
106-42-3	Xylene, P-	-	-	-	-	2.00E-01	S	1.00E-01	S	YES	NO	1.00E+00	-	RAGSE

Table B.6a. Toxicity Values and Information (Continued)

		PEF		VF		Skin	Skin Perm.				
	PEF	Ind./	VF	Ind./	KY ABS	Permeability	Const.		τ _{event}	t	В
Analyte	Res.9	Comm.9	Res.10	Comm.10	(Unitless)11	Constant ¹²	Ref	FA^{13}	(hr/event) 14	(hr) 15	(Unitless) 16
Aluminum	1.36E+09	1.36E+09	-	-	5.00E-02	1.00E-03	RAGSE	1	1.49E-01	3.57E-01	2.00E-03
Antimony (metallic)	1.36E+09	1.36E+09	-	1	5.00E-02	1.00E-03	RAGSE	1	5.05E-01	1.21E+00	4.24E-03
Arsenic, Inorganic ^d	1.36E+09	1.36E+09	-	-	3.00E-02	1.00E-03	RAGSE	1	2.76E-01	6.63E-01	3.33E-03
Barium	1.36E+09	1.36E+09	-	-	5.00E-02	1.00E-03	RAGSE	1	6.18E-01	1.48E+00	4.51E-03
Beryllium and compounds	1.36E+09	1.36E+09	-	-	7.00E-03	1.00E-03	RAGSE	1	1.18E-01	2.83E-01	1.15E-03
Boron And Borates Only	1.36E+09	1.36E+09	-	-	5.00E-02	1.00E-03	RAGSE	1	1.26E-01	3.02E-01	1.43E-03
Cadmium (Diet)	1.36E+09	1.36E+09	-	-	1.00E-03	1.00E-03	RAGSE	1	4.48E-01	1.08E+00	4.08E-03
Cadmium (Water)	-	-	-	-	1.00E-03	1.00E-03	RAGSE	1	4.48E-01	1.08E+00	4.08E-03
Chromium (Total) ^a	1.36E+09	1.36E+09	-	-	1.30E-02	1.00E-03	RAGSE	1	2.06E-01	4.93E-01	2.77E-03
Chromium(III), Insoluble Salts	1.36E+09	1.36E+09	-	-	1.30E-02	1.00E-03	RAGSE	1	2.06E-01	4.93E-01	2.77E-03
Chromium(VI)	1.36E+09	1.36E+09	-	-	2.50E-02	2.00E-03	RAGSE	1	2.06E-01	4.93E-01	5.55E-03
Cobalt	1.36E+09	1.36E+09	-	-	5.00E-02	4.00E-04	RAGSE	1	2.25E-01	5.40E-01	1.18E-03
Copper	1.36E+09	1.36E+09	-	-	5.00E-02	1.00E-03	RAGSE	1	2.39E-01	5.73E-01	3.07E-03
Fluoride	1.36E+09	1.36E+09	-	-	5.00E-02	1.00E-03	RAGSE	1	1.72E-01	4.12E-01	2.37E-03
Iron	1.36E+09	1.36E+09	-	-	5.00E-02	1.00E-03	RAGSE	1	2.16E-01	5.19E-01	2.87E-03
Leade	-	-	-	-	-	-	-	-	-	1	-
Manganese	1.36E+09	1.36E+09	-	-	4.00E-02	1.00E-03	RAGSE	1	2.14E-01	5.13E-01	2.85E-03
Mercury, Inorganic Salts	1.36E+09	1.36E+09	-	-	5.00E-02	1.00E-03	RAGSE	0	-	-	-
Molybdenum	1.36E+09	1.36E+09	-	-	5.00E-02	1.00E-03	RAGSE	1	3.62E-01	8.70E-01	3.77E-03
Nickel Soluble Salts	1.36E+09	1.36E+09	-	-	4.00E-02	2.00E-04	RAGSE	1	2.24E-01	5.38E-01	5.89E-04
Selenium	1.36E+09	1.36E+09	-	-	5.00E-02	1.00E-03	RAGSE	1	2.91E-01	6.99E-01	3.42E-03
Silver	1.36E+09	1.36E+09	-	-	4.00E-02	6.00E-04	RAGSE	1	4.23E-01	1.01E+00	2.40E-03
Thallium (Soluble Salts)	1.36E+09	1.36E+09	-	-	5.00E-02	1.00E-03	RAGSE	1	1.47E+00	3.52E+00	5.50E-03
Uranium (Insoluble Compounds)	1.36E+09	1.36E+09	-	-	5.00E-02	1.00E-03	RAGSE	1	2.26E+00	5.43E+00	5.93E-03
Uranium (Soluble Salts)	1.36E+09	1.36E+09	-	-	5.00E-02	1.00E-03	RAGSE	1	2.26E+00	5.43E+00	5.93E-03
Vanadium and Compounds	1.36E+09	1.36E+09	-	-	2.60E-02	1.00E-03	RAGSE	1	2.03E-01	4.87E-01	2.75E-03
Zinc and Compounds	1.36E+09	1.36E+09	-	-	5.00E-02	6.00E-04	RAGSE	1	2.44E-01	5.86E-01	1.87E-03
Acenaphthene	1.36E+09	1.36E+09	1.41E+05	1.41E+05	1.30E-01	8.60E-02	EPI	1	7.68E-01	1.84E+00	4.11E-01
Acenaphthylene ^b	1.36E+09	1.36E+09	1.89E+05	1.89E+05	1.30E-01	9.11E-02	EPI	1	7.48E-01	1.80E+00	4.32E-01
Acrylonitrile	1.36E+09	1.36E+09	7.69E+03	7.69E+03	2.50E-01	1.16E-03	EPI	1	2.08E-01	5.00E-01	3.25E-03
Anthracene	1.36E+09	1.36E+09	5.23E+05	5.23E+05	1.30E-01	1.42E-01	EPI	1	1.05E+00	4.05E+00	7.29E-01
Benzene	1.36E+09	1.36E+09	3.54E+03	3.54E+03	2.50E-01	1.49E-02	EPI	1	2.88E-01	6.91E-01	5.07E-02
Bis(2-ethylhexyl)phthalate ^c	1.36E+09	1.36E+09	-	-	1.00E-01	1.13E+00	EPI	0.8	1.62E+01	7.29E+01	8.59E+00
Bromodichloromethane	1.36E+09	1.36E+09	3.97E+03	3.97E+03	2.50E-01	4.02E-03	EPI	1	8.70E-01	2.09E+00	1.98E-02
Carbazole	1.36E+09	1.36E+09	-	-	1.00E-01	5.36E-02	EPI	1	9.08E-01	2.18E+00	2.67E-02
Carbon Tetrachloride	1.36E+09	1.36E+09	1.49E+03	1.49E+03	2.50E-01	1.63E-02	EPI	1	7.64E-01	1.83E+00	7.78E-02
Chloroform	1.36E+09	1.36E+09	2.63E+03	2.63E+03	2.50E-01	6.83E-03	EPI	1	4.90E-01	1.18E+00	2.87E-02
Dichlorodifluoromethane (Freon-12) ^c	1.36E+09	1.36E+09	8.41E+02	8.41E+02	2.50E-01	8.95E-03	EPI	1	5.00E-01	1.20E+00	3.79E-02

B-24

Table B.6a. Toxicity Values and Information (Continued)

		PEF		VF		Skin	Skin Perm.		1		T
	PEF	Ind./	VF	Ind./	KY ABS	Permeability	Const.		$ au_{\mathrm{event}}$		В
Analyte	Res.9	Comm.9	Res.10	Comm. ¹⁰	(Unitless)11	Constant ¹²	Ref	FA^{13}	(hr/event) 14	t (hr) 15	(Unitless)16
Dichloroethane, 1,1-°	1.36E+09	1.36E+09	2.08E+03	2.08E+03	2.50E-01	6.75E-03	EPI	1	3.77E-01	9.04E-01	2.58E-02
Dichloroethane, 1,2-	1.36E+09	1.36E+09	4.57E+03	4.57E+03	2.50E-01	4.20E-03	EPI	1	3.77E-01	9.04E-01	1.61E-02
Dichloroethylene, 1,1-	1.36E+09	1.36E+09	1.16E+03	1.16E+03	2.50E-01	1.17E-02	EPI	1	3.67E-01	8.81E-01	4.43E-02
Dichloroethylene, 1,2- (Mixed Isomers)	1.36E+09	1.36E+09	2.51E+03	2.51E+03	2.50E-01	1.10E-02	EPI	1	3.67E-01	8.81E-01	4.17E-02
Dichloroethylene, 1,2-cis-	1.36E+09	1.36E+09	2.50E+03	2.50E+03	2.50E-01	1.10E-02	EPI	1	3.67E-01	8.81E-01	4.17E-02
Dichloroethylene, 1,2-trans-	1.36E+09	1.36E+09	1.75E+03	1.75E+03	2.50E-01	1.10E-02	EPI	1	3.67E-01	8.81E-01	4.17E-02
Dieldrin	1.36E+09	1.36E+09	-	-	1.00E-01	3.26E-02	EPI	0.8	1.43E+01	3.43E+01	2.45E-01
Dioxins/Furans, Total ^f	-	-	-	-	-	-	-	-	-	-	-
~HpCDD	1.36E+09	1.36E+09	2.43E+06	2.43E+06	3.00E-02	1.33E+00	EPI	0	2.53E+01	1.15E+02	1.05E+01
~HpCDF, 2,3,7,8-	1.36E+09	1.36E+09	6.27E+06	6.27E+06	3.00E-02	1.45E+00	EPI	0	2.06E+01	9.37E+01	1.13E+01
~HxCDD, 2,3,7,8-	1.36E+09	1.36E+09	-	-	3.00E-02	2.86E+00	EPI	0	1.62E+01	7.51E+01	2.17E+01
~HxCDF, 2,3,7,8-	1.36E+09	1.36E+09	-	-	3.00E-02	1.35E+00	EPI	0	1.32E+01	5.99E+01	1.01E+01
~OCDD	1.36E+09	1.36E+09	-	-	3.00E-02	1.16E+00	EPI	0	3.95E+01	1.79E+02	9.57E+00
~OCDF	1.36E+09	1.36E+09	-	-	1.00E-01	2.63E+00	EPI	0	3.21E+01	1.49E+02	2.13E+01
~PeCDD, 2,3,7,8-	1.36E+09	1.36E+09	-	-	3.00E-02	2.41E-01	EPI	0.7	1.04E+01	4.20E+01	1.75E+00
~PeCDF, 1,2,3,7,8-	1.36E+09	1.36E+09	-	-	1.00E-01	6.27E-01	EPI	0.4	8.48E+00	3.69E+01	4.45E+00
~PeCDF, 2,3,4,7,8-	1.36E+09	1.36E+09	-	-	1.00E-01	6.27E-01	EPI	0.4	8.48E+00	3.69E+01	4.45E+00
~TCDD, 2,3,7,8-	1.36E+09	1.36E+09	1.96E+06	1.96E+06	3.00E-02	8.08E-01	EPI	0.5	6.68E+00	2.95E+01	5.58E+00
~TCDF, 2,3,7,8-	1.36E+09	1.36E+09	2.49E+06	2.49E+06	1.00E-01	6.57E-01	EPI	0.6	5.44E+00	2.36E+01	4.42E+00
Ethylbenzene	1.36E+09	1.36E+09	5.67E+03	5.67E+03	2.50E-01	4.93E-02	EPI	1	4.13E-01	9.92E-01	1.95E-01
Fluoranthene	1.36E+09	1.36E+09	-	-	1.30E-01	3.08E-01	EPI	1	1.43E+00	5.73E+00	1.68E+00
Fluorene	1.36E+09	1.36E+09	2.81E+05	2.81E+05	1.30E-01	1.10E-01	EPI	1	8.97E-01	2.15E+00	5.45E-01
Hexachlorobenzene	1.36E+09	1.36E+09	6.80E+04	6.80E+04	1.00E-01	2.54E-01	EPI	0.9	4.14E+00	1.66E+01	1.65E+00
Naphthalene	1.36E+09	1.36E+09	4.63E+04	4.63E+04	1.30E-01	4.66E-02	EPI	1	5.49E-01	1.32E+00	2.03E-01
Nitroaniline, 2-	1.36E+09	1.36E+09	-	-	1.00E-01	4.46E-03	EPI	1	6.24E-01	1.50E+00	2.02E-02
Nitroso-di-N-propylamine, N-	1.36E+09	1.36E+09	-	-	1.00E-01	2.33E-03	EPI	1	5.64E-01	1.35E+00	1.02E-02
Pentachlorophenol	1.36E+09	1.36E+09	-	-	2.50E-01	1.27E-01	EPI	0.9	3.26E+00	1.25E+01	7.97E-01
Phenanthrene ^b	1.36E+09	1.36E+09	6.43E+05	6.43E+05	1.30E-01	1.44E-01	EPI	1	1.05E+00	4.05E+00	7.39E-01
Polychlorinated Biphenyls (high risk)	1.36E+09	1.36E+09	5.32E+05	5.32E+05	1.40E-01	5.45E-01	EPI	0.7	4.54E+00	1.94E+01	3.58E+00
Polychlorinated Biphenyls (low risk)	1.36E+09	1.36E+09	5.32E+05	5.32E+05	1.40E-01	5.45E-01	EPI	0.7	4.54E+00	1.94E+01	3.58E+00
~Aroclor 1016	1.36E+09	1.36E+09	5.86E+05	5.86E+05	1.40E-01	3.05E-01	EPI	0.9	1.26E+02	1.18E+01	1.88E+00
~Aroclor 1221	1.36E+09	1.36E+09	2.04E+05	2.04E+05	1.40E-01	1.68E-01	EPI	1	1.20E+00	4.60E+00	8.88E-01
~Aroclor 1232	1.36E+09	1.36E+09	1.12E+05	1.12E+05	1.40E-01	1.68E-01	EPI	1	1.20E+00	4.60E+00	8.88E-01
~Aroclor 1242	1.36E+09	1.36E+09	5.91E+05	5.91E+05	1.40E-01	5.45E-01	EPI	0.7	4.54E+00	1.94E+01	3.58E+00
~Aroclor 1248	1.36E+09	1.36E+09	5.14E+05	5.14E+05	1.40E-01	4.75E-01	EPI	0.7	3.06E+02	1.92E+01	3.12E+00
~Aroclor 1254	1.36E+09	1.36E+09	8.43E+05	8.43E+05	1.40E-01	7.51E-01	EPI	0.5	7.08E+00	3.11E+01	5.22E+00
~Aroclor 1260	1.36E+09	1.36E+09	1.31E+06	1.31E+06	1.40E-01	9.86E-01	EPI	0	1.72E+01	7.71E+01	7.54E+00

Table B.6a. Toxicity Values and Information (Continued)

Analyte	PEF Res. ⁹	PEF Ind./ Comm. ⁹	VF Res. ¹⁰	VF Ind./ Comm. ¹⁰	KY ABS (Unitless) ¹¹	Skin Permeability Constant ¹²	Skin Perm. Const. Ref	FA ¹³	τ _{event} (hr/event) 14	t (hr) 15	B (Unitless) 16
Polycyclic aromatic hydrocarbons (cPAH), Total Carcinogenic ^f	_	-	-	-	-	-	-	_	-	-	-
~Benz[a]anthracene	1.36E+09	1.36E+09	4.41E+06	4.41E+06	1.30E-01	5.52E-01	EPI	1	2.00E+00	8.48E+00	3.21E+00
~Benzo[a]pyrene	1.36E+09	1.36E+09	-	-	1.30E-01	7.13E-01	EPI	1	2.72E+00	1.18E+01	4.36E+00
~Benzo[b]fluoranthene	1.36E+09	1.36E+09	-	-	1.30E-01	4.17E-01	EPI	1	2.72E+00	1.13E+01	2.55E+00
~Benzo[k]fluoranthene	1.36E+09	1.36E+09	-	-	1.30E-01	6.91E-01	EPI	0.9	2.72E+00	1.18E+01	4.22E+00
~Chrysene	1.36E+09	1.36E+09	-	-	1.30E-01	5.96E-01	EPI	1	2.00E+00	8.53E+00	3.46E+00
~Dibenz[a,h]anthracene	1.36E+09	1.36E+09	-	-	1.30E-01	9.53E-01	EPI	0.6	3.81E+00	1.69E+01	6.12E+00
~Indeno[1,2,3-cd]pyrene	1.36E+09	1.36E+09	-	-	1.30E-01	1.24E+00	RAGSE	0.6	3.71E+00	1.67E+01	7.93E+00
Pyrene	1.36E+09	1.36E+09	2.38E+06	2.38E+06	1.30E-01	2.01E-01	EPI	1	1.43E+00	5.54E+00	1.10E+00
Tetrachloroethylene	1.36E+09	1.36E+09	2.35E+03	2.35E+03	2.50E-01	3.34E-02	EPI	1	8.92E-01	2.14E+00	1.65E-01
Toluene ^c	1.36E+09	1.36E+09	4.29E+03	4.29E+03	2.50E-01	3.11E-02	EPI	1	3.45E-01	8.28E-01	1.15E-01
Trichloroethane, 1,1,1-	1.36E+09	1.36E+09	1.65E+03	1.65E+03	2.50E-01	1.26E-02	EPI	1	5.87E-01	1.41E+00	5.60E-02
Trichloroethane, 1,1,2-	1.36E+09	1.36E+09	7.22E+03	7.22E+03	2.50E-01	5.04E-03	EPI	1	5.87E-01	1.41E+00	2.24E-02
Trichloroethylene	1.36E+09	1.36E+09	2.21E+03	2.21E+03	2.50E-01	1.16E-02	EPI	1	5.72E-01	1.37E+00	5.11E-02
Trichloro-1,2,2-trifluoroethane, 1,1,2-											
(Freon-113) ^c	1.36E+09	1.36E+09	1.29E+03	1.29E+03	2.50E-01	1.75E-02	EPI	1	1.18E+00	2.83E+00	9.21E-02
Vinyl Chloride	1.36E+09	1.36E+09	9.56E+02	9.56E+02	2.50E-01	8.38E-03	EPI	1	2.35E-01	5.65E-01	2.55E-02
Xylene, Mixture	1.36E+09	1.36E+09	5.74E+03	5.74E+03	2.50E-01	5.00E-02	EPI	1	4.13E-01	9.92E-01	1.98E-01
Xylene, m-	1.36E+09	1.36E+09	5.47E+03	5.47E+03	2.50E-01	5.32E-02	EPI	1	4.13E-01	9.92E-01	2.11E-01
Xylene, o-	1.36E+09	1.36E+09	6.45E+03	6.45E+03	2.50E-01	4.71E-02	EPI	1	4.13E-01	9.92E-01	1.87E-01
Xylene, P-	1.36E+09	1.36E+09	5.58E+03	5.58E+03	2.50E-01	4.93E-02	EPI	1	4.13E-01	9.92E-01	1.95E-01

Table B.6b. Toxicity Values and Information

Chemical Abstract		Inhalation Slope Factor	SFi	Oral Slope Factor for	SFow	Oral Slope Factor for Soil (SFos)	Oral Slope Factor for Soil (SFos)	SFos	Oral Slope Factor for	External Exposure Slope Factor	SFe		
Number	Analyte	(SFi) ³	Ref	Water (SFow)17	Ref	Res.17	Ind./Comm.17	Ref	Food (SFof)17	(SFe) ¹⁸	Ref	Lambda ¹⁹	Halflife ¹⁹
14596-10-2	Am-241	3.77E-08	О	1.04E-10	О	1.84E-10	9.10E-11	O	1.34E-10	2.77E-08	О	1.60E-03	4.32E+02
10045-97-3	Cs-137	1.12E-10	О	3.05E-11	О	4.26E-11	3.18E-11	O	3.74E-11	5.52E-10	О	2.30E-02	3.02E+01
13994-20-2	Np-237	2.87E-08	О	6.22E-11	О	1.25E-10	4.70E-11	O	8.29E-11	5.17E-08	О	3.23E-07	2.14E+06
13981-16-3	Pu-238	5.22E-08	О	1.31E-10	О	2.25E-10	1.17E-10	O	1.69E-10	6.91E-11	О	7.90E-03	8.77E+01
15117-48-3	Pu-239	5.55E-08	О	1.35E-10	О	2.28E-10	1.21E-10	O	1.74E-10	2.09E-10	О	2.87E-05	2.41E+04
14119-33-6	Pu-240	5.55E-08	О	1.35E-10	О	2.28E-10	1.21E-10	О	1.74E-10	7.12E-11	О	1.06E-04	6.56E+03
14133-76-7	Tc-99	3.81E-11	О	2.75E-12	О	7.25E-12	1.32E-12	О	4.00E-12	8.28E-11	О	3.28E-06	2.11E+05
14269-63-7	Th-230	3.41E-08	О	9.14E-11	О	1.66E-10	7.73E-11	O	1.19E-10	8.45E-10	О	9.19E-06	7.54E+04
7440-29-1	Th-232	4.33E-08	О	1.01E-10	O	1.84E-10	8.47E-11	O	1.33E-10	3.58E-10	О	4.93E-11	1.41E+10
13966-29-5	U-234	2.78E-08	О	7.07E-11	О	1.48E-10	5.11E-11	О	9.55E-11	2.53E-10	О	2.82E-06	2.46E+05
15117-96-1	U-235	2.50E-08	0	6.96E-11	О	1.48E-10	4.92E-11	О	9.44E-11	5.51E-07	О	9.84E-10	7.04E+08
7440-61-1	U-238	2.36E-08	О	6.40E-11	О	1.34E-10	4.66E-11	0	8.66E-11	1.24E-10	0	1.55E-10	4.47E+09

Information compiled from RAIS and EPA Radionuclide PRG chemical online calculators October 2022.

Note that the toxicity values and information is presented in a split table format. Cells containing "-" indicate no value or information is available.

Reference Codes:

- A Agency for Toxic Substances and Disease Registry (ATSDR) minimal risk levels
- C The California EPA Office of Environmental Health Hazard Assessment's (OEHHA) Chronic Reference Exposure Levels (RELS) from December 18, 2008, and the Cancer Potency Values from July 21, 2009
- E EPA/RPF
- EPI EPA's Estimation Programs Interface Suite
- H EPA's Health Effects Assessment Summary Tables (HEAST)
- I EPA's Integrated Risk Information System (IRIS)
- O Oak Ridge National Laboratory (ORNL) Technical Memorandum (TM) (ORNL/TM-2013/00), September 2014
- P The Provisional Peer Reviewed Toxicity Values (PPRTVs) derived by EPA's Superfund Health Risk Technical Support Center (STSC) for the EPA Superfund program
- PS EPA's PPRTVs Appendix Screening Levels
- RAGSE Risk Assessment Guidance for Superfund, Part E
- S Surrogate
- The RfDo and RfCi were derived as the midpoint of similar candidate RfDos and RfCis.
- W World Health Organization

^a Values for Chromium (Total) should use toxicity factors for Chromium VI, unless it is determined on a project-specific basis that chromium VI is not present. If chromium VI is not present, chromium III should be used. This approach is consistent with Screening Level note 9b (Appendix A).

^b Values for Acenaphthylene and Phenanthrene, if not available use toxicity factors for Acenaphthene.

^c Analytes are not PGDP-significant COPCs (Table 2.1), but are provided for project support.

d Calculations for arsenic include a relative bioavailability factor for soil ingestion of 0.6. For additional information, see the EPA document, OSWER 9200.1-113, December 2012.

^eLead toxicity values are not included because lead is evaluated using Integrated Exposure Uptake Biokinetic modeling and Adult Lead Model modeling, as appropriate. See Section B.3.

^fToxicity values for Total Dioxins/Furans use those for 2,3,7,8-TCDD and cPAHs use those for benzo[a]pyrene.

Notes on Tables B.6a. and B.6b.

Prior to using the values in this table, a risk assessor must be consulted to determine if any values need to be updated and to verify that the values are being used appropriately.

- 1. Information used to derive PRGs for COPCs at the PGDP is shown.
- 2. The "Oral Slope Factor" is the chronic oral slope factor used for the ingestion routes of exposure. The units on this value for chemicals is $[mg/(kg \times day)]^{-1}$. The units on this value for radionuclides is $(pCi)^{-1}$.
- The "Inhalation Unit Risk" is the chronic inhalation factor used for inhalation routes of exposure. The values listed for chemicals are in units of mg/m³, although they typically are expressed in μg/m³.
 For radionuclides, the inhalation slope factor continues to be used. The units on this value for radionuclides is (pCi)⁻¹.
- 4. The "Oral RfD" is the chronic oral reference dose used for ingestion routes of exposure. The units for Oral RfD are mg/(kg x day).
- 5. The "Inhalation RfC" is the chronic inhalation concentration used for inhalation routes of exposure. The units for Inhalation RfC are mg/m³.
- 6. "Volatile Organic?" is a flag used to specify if the chemical should be assessed as a vapor. A chemical is considered volatile in this context if it has a vapor pressure greater than 1 mm Hg or a Henry's Law constant greater than 0.00001 atm-m³/mole (RAIS 2021).
- 7. The column labeled "Mutagen?" is a flag used to specify if the chemical should be assessed as a mutagen. This assessment is made when PRGs are developed using the RAIS Chemical PRG online calculator. See Section 3.3.6.1 of the main text.
- 8. The "EPA ABS" is the dermal absorption value recommended by EPA in their guidance material, 2004 RAGs, Part E. The dermal absorption value is unitless.
- 9. The "Particle Emission Factor" is a value used to assess inhalation routes of exposure. The particle emission factor is in units of m³/kg. The values for residential (Res.) and industrial/commercial (Ind./Comm.) scenario listed are taken from the 2002 *Kentucky Risk Assessment Guidance*.
- 10. The "Volatilization Factor" is a value used to assess inhalation routes of exposure. Values are given for residential (Res.) and industrial/commercial (Ind./Comm.) scenarios. The volatilization factor is in units of m³/kg.
- 11. The "KY ABS" is the dermal absorption value recommended by the Commonwealth of Kentucky in their guidance material, 2002 *Kentucky Risk Assessment Guidance*. Dermal exposure to soil used default absorption values of 0.25 for volatiles, 0.1 for semivolatiles, and 0.05 for metals. The dermal absorption value is unitless.
 - In RAGS Part E, 2004, Exhibit 4-1, the following GI absorption efficiencies are listed that are below the 5% dermal absorption KDEP has recommended as a default value for inorganics. For these constituents, the dermal absorption value should be modified from 5% to mimic the GI absorption efficiencies, as follows: Beryllium 0.007 = 0.7%; Chromium III 0.013 = 1.3%; Chromium VI 0.025 = 2.5%; Manganese 0.04 = 4%; Nickel 0.04 = 4%; Silver 0.04 = 4%; Vanadium 0.026 = 2.6%
 - This is in addition to the chemical-specific dermal absorption fractions listed in RAGS, Part E, Exhibit 3-4, including: Arsenic 0.03 = 3% and Cadmium 0.001 = 0.1%
 - Additional deviations from the 2002 Kentucky Risk Assessment Guidance are documented in the August 14, 2007, meeting minutes (see DOE 2017, Appendix E).
- 12. The "Permeability Constant" is a chemical-specific value used to estimate dermal absorption of chemicals in water. The permeability constant is in units of cm/hr.
- 13. The "FA" is the fraction absorbed water. The chemical-specific value is unitless. Values were taken from RAIS.
- 14. The "tevent" is the lag time per event and is a chemical-specific value. The values were taken from RAIS and are in units of hours/event.
- 15. The variable "t" indicates time to reach steady-state. Values are chemical-specific in units of hours. Values were taken from RAIS.
- 16. "B" is the dimensionless ratio of the permeability coefficient of a compound through the stratum corneum, relative to its permeability coefficient across the viable epidermis. The chemical-specific values were taken from RAIS.
- 17. "Oral Slope Factor for Water," "Oral Slope Factor for Soil," and "Oral Slope Factor for Food" are the indicated values for radionuclides. The units for these factors are (pCi)-1.
- 18. The "External Exposure Slope Factor" is the slope factor used for external exposure to ionizing radiation emitted by radioactive chemicals. The units for external exposure slope factor are [(pCi x year)/g]⁻¹.
- 19. "Lambda" is a decay constant. It is equal to 0.693/half-life (year-1) where 0.693 = ln(2). The units for lambda are (year-1). "Half-life" is the time taken for the radioactivity of a specified isotope to fall to half its original value. The units for half-life are years.



APPENDIX C OUTLINE FOR BASELINE HUMAN HEALTH RISK ASSESSMENTS



OUTLINE FOR BASELINE HUMAN HEALTH RISK ASSESSMENTS

- *** Although the following outline can be used for baseline human health risk assessments for both source units and integrator units, not all sections may be relevant to all assessments and additional sections may be needed for some assessments. However, all baseline risk assessments completed for PGDP should include each of the first and second level headers listed below.
- *** The document should begin with an introduction that presents the scope and objectives of the baseline human health risk assessment. This should include a description of the general problem at the site and an overview of the design of the baseline human health risk assessment.
- 1. Results of Previous Studies
 - *** The section should begin with a brief summary of the previous studies that are relevant to the baseline human health risk assessment. All relevant previous risk evaluations should be summarized.
 - 1.1 Study #1
 - 1.2 Study #2

Etc.

- 2. Identification of Chemicals of Potential Concern
 - *** The section should begin with an introduction that describes the purpose of the section and the order in which the material is presented.
 - 2.1 Sources of Data
 - *** The sources of all data should be listed, and the projects in which the data were collected should be described.
 - 2.2 General Data Evaluation Considerations
 - *** The eight steps of data evaluation as applied to the baseline risk assessment should be discussed.
 - 2.2.1 Evaluation of Sampling Design
 - 2.2.2 Evaluation of Analytical Methods
 - 2.2.3 Evaluation of Sample Quantitation Limits
 - 2.2.4 Evaluation of Data Qualifiers and Codes
 - 2.2.5 Elimination of Chemicals not Detected
 - 2.2.6 Examination of Toxicity of Detected Analytes
 - 2.2.7 Examination of Essential Nutrients
 - 2.2.8 Comparison of Analyte Concentrations and Activities Detected in Site Samples to Background Concentrations
 - 2.3 Risk Assessment Specific Data Evaluation
 - *** This section should discuss in detail how the eight steps were applied to identify the chemicals of potential concern under both current and future conditions.
 - 2.3.1 Current Conditions
 - *** This section should discuss the evaluation of the data set.
 - 2.3.2 Future Conditions

*** This section should discuss any modeling performed to address potential future changes in the identity or concentration of contaminants.

2.4 Evaluation of Data from Other Sources

- *** The section should introduce any "special data," especially nonnumeric data (such as activities of visitors at a site or types of vegetables grown by Kentucky residents) used to develop the exposure assessment that are not used quantitatively in the baseline human health risk assessment. Examples of special data that may be used are found in the survey forms and responses in Appendix E.
- 2.4.1 Other Source #1
- 2.4.2 Other Source #2

Etc

2.5 Summary of Chemicals of Potential Concern

*** This section should present a summary of the quantitative data evaluation and its results.

3. Exposure Assessment

*** This section should begin with a description of the process used in exposure assessment, and the goal of the specific exposure assessment being performed.

3.1 Characterization of Exposure Setting

- *** This section should describe either by reference or directly the following:
- 3.1.1 Surface Features
- 3.1.2 Meteorology
- 3.1.3 Geology
- 3.1.4 Demography and Land Use
- 3.1.5 Ecology
- 3.1.6 Hydrology
- 3.1.7 Hydrogeology

3.2 Identification of Exposure Pathways

*** This section should begin by describing what a pathway is and how a pathway can be complete or incomplete.

3.2.1Land Use Considerations

*** The land use under current and expected and potential future conditions should be described.

3.2.2Potential Receptor Populations

*** The potential receptors under both current and future uses should be described and justified.

3.2.3 Delineation of Exposure Points/Exposure Routes

*** All possible exposure routes should be presented and justified. The number of possible exposure routes should be reduced, if possible, so that only probable exposure routes with significant risk or hazard are quantified. The exposure equations used in the assessment to quantify exposure should be presented. Justification for not quantifying a possible route should be presented.

3.2.4Development of Conceptual Site Models

*** Figures illustrating the pathways of exposure should be presented for each site under investigation. The model for each site should be justified.

3.3 Quantification of Exposure

*** The methods used to quantify exposure (i.e., estimate dose) should be described for each receptor. If modeling is used to determine concentration or activities of chemicals of potential concern in biota, the models should be presented.

3.4 Summary of Exposure Assessment

4. Toxicity Assessment

*** This section should begin by describing the goal and methods used for toxicity assessment. The source of all toxicity values should be discussed. Tables presenting the toxicity information should be presented.

4.1 Inorganics

- *** The toxicity of each chemical of potential concern should be profiled. Each profile should include a listing of the carcinogenic and noncarcinogenic toxicity values used in the baseline human health risk assessment.
- 4.1.1 Chemical 1
- 4.1.2 Chemical 2

Etc

4.2 Organics

- *** The toxicity of each chemical of potential concern should be profiled. Each profile should include a listing of the toxicity values used in the baseline human health risk assessment.
- 4.2.1 Chemical 1
- 4.2.2 Chemical 2

Etc.

4.3 Radionuclides

- *** The toxicity of each chemical of potential concern should be profiled. Each profile should include a listing of the toxicity values used in the baseline human health risk assessment.
- 4.3.1 Radionuclide 1
- 4.3.2 Radionuclide 2

Etc.

4.4 Chemicals for Which No EPA Toxicity Values Are Available

*** The chemicals of potential concern that fall in this class should be listed. If the baseline human health risk assessment is evaluating multiple units or areas, these chemicals should be listed by unit or area. This section should include the procedure for evaluating potential surrogate chemicals that may be available for some of the chemicals without toxicity values.

4.5 Uncertainties Related to Toxicity Assessment

*** A brief presentation of the uncertainties related to all toxicity assessments and toxicity values should be made.

4.6 Summary

*** The amount of toxicity information for the chemicals of potential concern should be discussed. If the baseline human health risk assessment is evaluating multiple units or areas, this information should be presented by unit or area.

5. Risk Characterization

*** The section should begin with a brief discussion of the purpose and goals of risk characterization and what will result from this step of the assessment.

5.1 Determination of Noncancer Effects

*** The methods used to quantify systemic toxicity for each chemical, both within and across pathways should be presented. If exposure over multiple scenarios or areas is possible, this should be noted.

5.2 Determination of Excess Cancer Risk

*** The methods used to quantify excess lifetime cancer risk for each chemical, both within and across pathways should be presented. If exposure over multiple scenarios or areas is possible, this should be noted.

5.3 Risk Characterization for Current Use Scenario(s)

*** Risk results for each unit or area should be presented in two-way tables and in a narrative summary. If subchronic effects are characterized, they should be presented separately from the chronic effects.

5.3.1 Systemic Toxicity

5.3.2 Excess Lifetime Cancer Risk

5.4 Risk Characterization for Future Use Scenario(s)

*** Risk results for each unit or area should be presented in two-way tables and in a narrative summary. If more than one future time is quantitatively evaluated, the results should be presented for each time period. If subchronic effects are characterized, they should be presented separately from the chronic effects.

5.4.1 Systemic Toxicity

5.4.2 Excess Lifetime Cancer Risk

5.5 Risk Characterization for Lead (if needed)

*** The special problems associated with risk characterization for lead should be discussed. Results from lead modeling and from comparisons against EPA and Kentucky screening values should be presented by unit or area.

5.6 Identification of Use Scenarios, Chemicals, Pathways, and Media of Concern

*** The section should begin with a listing of the rules used to identify use scenarios, chemicals, pathways and media of concern.

5.6.1 Use Scenarios of Concern

*** These should be listed within area or unit under investigation.

5.6.2 Chemicals of Concern

*** These should be listed within area or unit under investigation.

5.6.3 Pathways of Concern

*** These should be listed within area or unit under investigation.

5.6.4 Media of Concern

*** These should be listed within area or unit under investigation

5.7 Summary of Risk Characterization

*** This section should describe and present the risk characterization summary tables.

6. Uncertainty in the Risk Assessment

***This section should begin with a general discussion of uncertainty. If a qualitative uncertainty analysis is being performed, "small," "moderate," and "large" uncertainties should be defined and the following subsections should be included. If a quantitative uncertainty analysis is being performed, the methods and results should be described in detail. Normally, a qualitative analysis, including sensitivity analyses, will be sufficient. Regardless, this section should continue with a discussion of each of the uncertainties affecting the major portions of the risk assessment. (Note, the uncertainties listed below are some of those found in past assessments. The uncertainties to be addressed in future assessments must be determined on a case-by-case basis.)

6.1 Uncertainties Associated with Data

*** The uncertainties to be discussed should be summarized in the introduction of this section. Categories of uncertainties to discuss are presented in the following.

- 6.1.1 Selection of Chemicals of Potential Concern
- 6.1.2 Determination of Exposure Point Concentrations—Current Conditions
- 6.1.3 Determination of Exposure Point Concentrations—Future Conditions
- 6.1.4 Use of Unfiltered versus Filtered Water Samples

6.2 Uncertainties Associated with Exposure Assessment

*** The uncertainties to be discussed should be summarized in the introduction of this section. Categories of uncertainties to discuss are presented in the following.

6.2.1 Uncertainties in Fate and Transport Modeling

6.2.2 Uncertainties in Use of Reasonable Maximum Exposure (RME) Scenarios

- 6.2.3 Uncertainties Related to Development of Conceptual Site Models
- 6.2.4 Uncertainties Related to Use of Default Values When Estimating Dermal Absorbed Dose

6.3 Uncertainties Associated with Toxicity Assessment

- *** The uncertainties to be discussed should be summarized in the introduction of this section. Categories of uncertainties to discuss are presented in the following.
- 6.3.1 Uncertainties Due to Lack of Toxicity Values for Some Chemicals
- 6.3.2 Uncertainties in Deriving Toxicity Values
- 6.3.3 Uncertainties Due to Calculation of Absorbed Dose Toxicity Values from Administered Toxicity Values
- 6.3.4 Uncertainties Due to Use of Toxicity Values for Chronic Exposure for Subchronic Exposure Times

6.4 Uncertainties Associated with Risk Characterization

- *** The uncertainties to be discussed should be summarized in the introduction of this section. Categories of uncertainties to discuss are presented in the following.
- 6.4.1 Uncertainties in Combining Chemical-Specific Risk and Hazard Estimates and Pathway-Specific Risk and Hazard Estimates
- 6.4.2 Uncertainties in Combining Risk Estimated for Chemical Exposure to those for Risk Estimated for Radioisotope Exposure

6.5 Summary of Uncertainties

*** This section should summarize the uncertainties discussed earlier in the section and present a table reviewing all uncertainties.

7. Conclusions and Summary

*** The purpose of this section is to review the results of the risk assessment without the use of tables and explanations and provide significant observations interpreting the results of the assessment for use by risk managers. When properly presented, it should be possible to insert this section as written into the feasibility study.

7.1 Chemicals of Potential Concern

*** A brief description of the screening process should be provided, and the chemicals of potential concern for each area or unit listed either by name (if the list is short) or by class.

7.2 Exposure Assessment

*** The exposure pathways quantitatively evaluated should be listed for each use scenario

7.3 Toxicity Assessment

*** The amount of available toxicity data for the chemicals of potential concern for each area should be listed. Chemicals of potential concern lacking toxicity values should be highlighted.

7.4 Risk Characterization

*** The use scenarios, chemicals, pathways, and media of concern should be listed for each area or unit, and the rules used to delineate the use scenarios, chemicals, pathways, and media of concern should be presented.

7.5 Observations

*** This section should integrate the risk estimates and the uncertainties to develop a list of salient issues to be considered by risk managers when making decisions in risk management documents. This includes a discussion for each of the chemicals of concern identified in the risk assessment. In addition, the results of the baseline human health risk assessment should be compared to results of previous risk evaluations, if any.

8 Remedial Goal Options

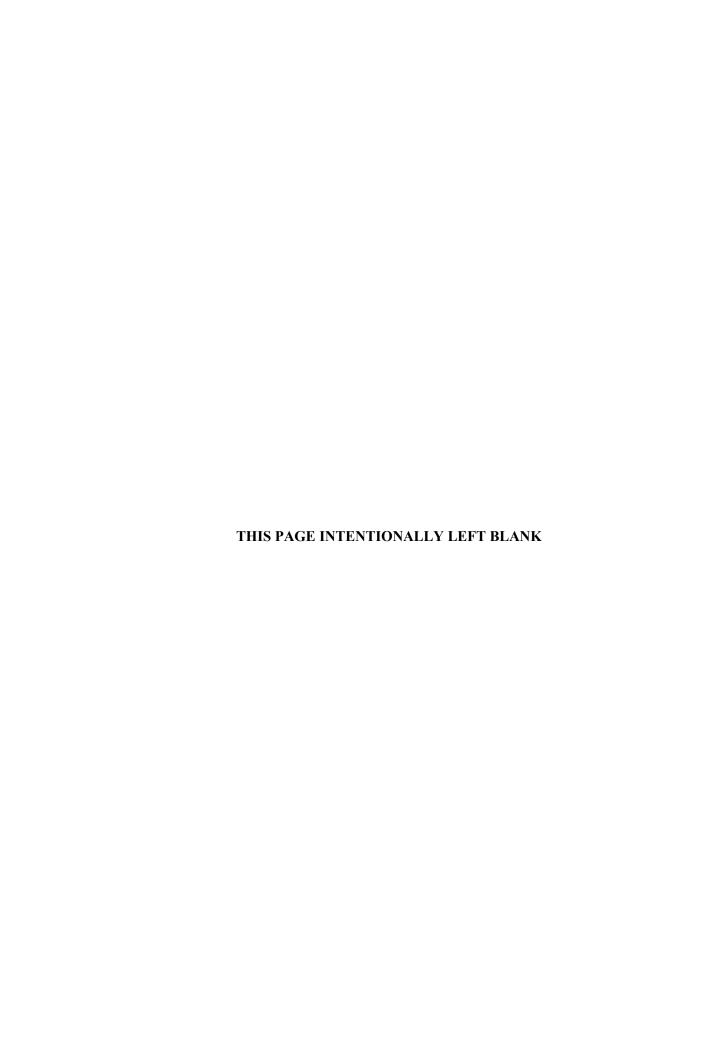
*** This section should present the methods used to derive the remedial goal options and list the remedial goal options for each chemical of concern. Because remedial goal options are medium- and scenario-specific, a separate list should be presented for each area (or unit), scenario, and medium combination.

8.1 Derivation of RGOs

*** This presentation should be as brief as possible.

8.2 Presentation of RGOs

*** These should be presented in tables. Very little narrative beyond directing the reader to the appropriate tables is needed.



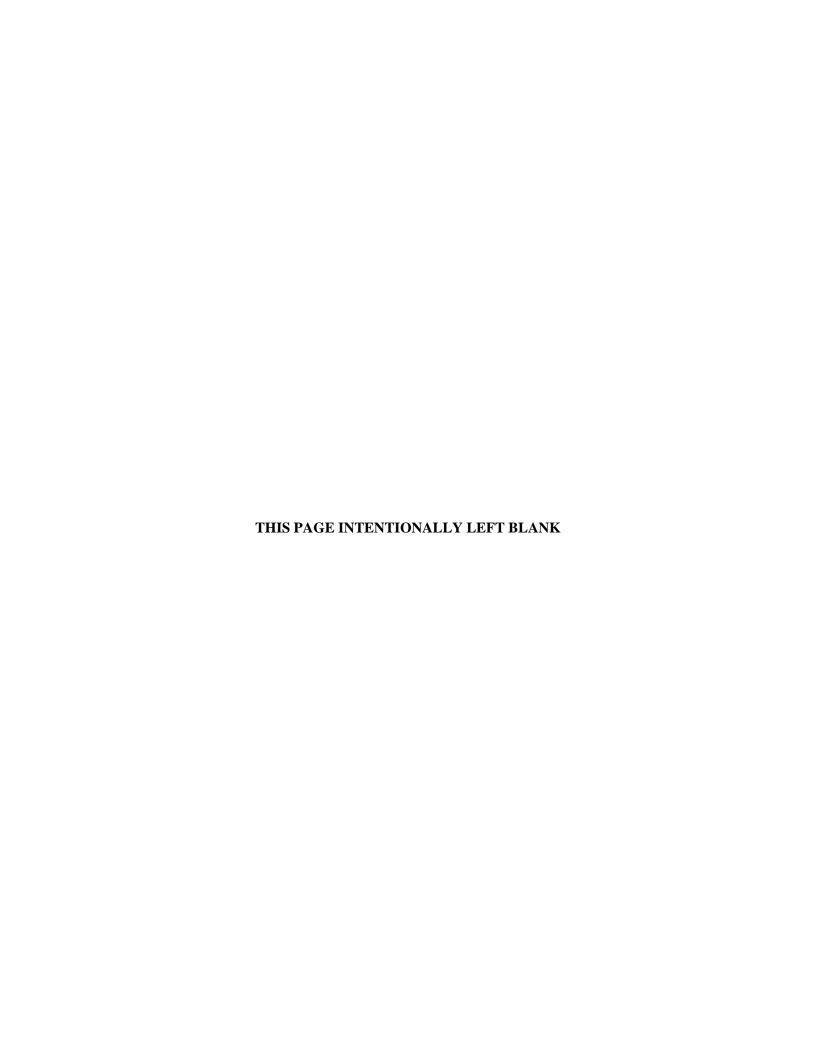
APPENDIX D EXPOSURE EQUATIONS



EXPOSURE EQUATIONS

This appendix contains the exposure equations used in environmental human health risk assessments for Department of Energy sites located at the Paducah Gaseous Diffusion Plant (PGDP). It should be noted that the equations shown in this appendix may not be the same as those used in preliminary remediation goal (PRG) calculations. PRG calculations were taken from the Risk Assessment Information System (RAIS) Chemical PRG online calculator (available at http://rais.ornl.gov/) and the U.S. Environmental Protection Agency (EPA) Radionuclide PRG calculator (available at https://epa-prgs.ornl.gov/radionuclides/).

The equations in this appendix are consistent with all Region 4 EPA and Commonwealth of Kentucky guidance materials. Unless otherwise noted, equations are from EPA's Risk Assessment Guidance for Superfund, Volume 1 (EPA 1989). The exposure parameters are those used to produce daily intakes and absorbed doses used to complete environmental risk assessments performed for PGDP only. These exposure parameters are for a default reasonable maximum exposure (RME). While these exposure parameters generally are consistent with the exposure parameters recommended by Region 4 EPA, they do differ in some cases, as determined by the PGDP Risk Assessment Working Group. The source of each value is provided below the equation. Equations to complete dose assessments and to derive dose conversion factors are not presented; however, these can be derived from the information provided here.



TABLES

D.1.	Reasonable Maximum Exposure Assumptions for Ingestion of Water	D-7
D.2.	Reasonable Maximum Exposure Assumptions for Inhalation of Volatile Organic	
	Compounds in Water during Household Use (including Showering)	D-8
D.3.	Reasonable Maximum Exposure Assumptions for Dermal Contact with Water while	
		D-9
D.4.	Reasonable Maximum Exposure Assumptions for Incidental Ingestion of Soil/Sediment	D-10
D.5.	Reasonable Maximum Exposure Assumptions for Dermal Contact with Soil/Sediment	
D.6.	Reasonable Maximum Exposure Assumptions for Inhalation of Vapors and Particulates	
	Emitted from Soil/Sediment	D-12
D.7.	Reasonable Maximum Exposure Assumptions for External Exposure to Ionizing	
	Radiation from Soil/Sediment	D-13
D.8.	Reasonable Maximum Exposure Assumptions for Incidental Ingestion of Surface Water	
	while Swimming	D-13
D.9.	Reasonable Maximum Exposure Assumptions for Dermal Contact with Surface Water	
	(Wading)	D-14
D.10.	Reasonable Maximum Exposure Assumptions for Dermal Contact with Surface Water	
	(Swimming)	D-15
D.11.	Reasonable Maximum Exposure Assumptions for Consumption of Fish	D-16
D.12.	Reasonable Maximum Exposure Assumptions for Concentration or Activity	
	Concentration of COPCs in Fish	D-16
D.13.	Reasonable Maximum Exposure Assumptions for Consumption of Deer	D-17
D.14.	Reasonable Maximum Exposure Assumptions for Concentration or Activity	
	Concentration of COPCs in Deer	D-18
D.15.	Reasonable Maximum Exposure Assumptions for Consumption of Rabbit	D-19
D.16.	Reasonable Maximum Exposure Assumptions for Concentration or Activity	
	Concentration of COPCs in Rabbit	D-20
D.17.	Reasonable Maximum Exposure Assumptions for Consumption of Quail	D-21
D.18.	Reasonable Maximum Exposure Assumptions for Concentration or Activity	
	Concentration of COPCs in Quail	D-22
D.19.	Reasonable Maximum Exposure Assumptions for Consumption of Homegrown	
	Vegetables	D-23
D.20.	Reasonable Maximum Exposure Assumptions for Concentration or Activity	
	Concentration of COPCs in Homegrown Vegetables	
D.21.	Reasonable Maximum Exposure Assumptions for Consumption of Beef	D-25
D.22.	Reasonable Maximum Exposure Assumptions for Concentration or Activity	
	Concentration of COPCs in Beef	D-26
D.23.	Reasonable Maximum Exposure Assumptions for Consumption of Milk	D-27
D.24.	Reasonable Maximum Exposure Assumptions for Concentration or Activity	
	Concentration of COPCs in Milk	
D.25.	Reasonable Maximum Exposure Assumptions for Consumption of Poultry	D-29
D.26.	Reasonable Maximum Exposure Assumptions for Concentration or Activity	
	Concentration of COPCs in Poultry	
D.27.	Reasonable Maximum Exposure Assumptions for Consumption of Pork	D-31
D.28.	Reasonable Maximum Exposure Assumptions for Concentration or Activity	
	Concentration of COPCs in Pork	
D.29.	Reasonable Maximum Exposure Assumptions for Consumption of Eggs	D-33

D.30.	Reasonable Maximum Exposure Assumptions for Concentration or Activity	
	Concentration of COPCs in Egg	D-33
D.31.	Reasonable Maximum Exposure Assumptions for Inhalation of Vapors in Ambient Air	D-34
REFE	RENCES FOR EXPOSURE EQUATIONS	D-35

Table D.1. Reasonable Maximum Exposure Assumptions for Ingestion of Water

Chemical Intake [mg/(kg × day)] =
$$\frac{C_w \times IR \times EF \times ED}{BW \times AT}$$

Radionuclide Intake (pCi) =
$$A_w \times IR \times EF \times ED$$

Parameter	Units	Value used	Referencesa
Chemical concentration in water = $\mathbf{C}_{\mathbf{w}}$	mg/L	Chemical-specific	
Radiological activity concentration in water = $\mathbf{A}_{\mathbf{w}}$	pCi/L	Chemical-specific	
Ingestion rate = \mathbf{IR}	L/day	See Table B.5	
Exposure frequency = $\mathbf{E}\mathbf{F}$	days/year	See Table B.5	
Exposure duration $=$ ED	years	See Table B.5	
Body weight = \mathbf{BW}	kg	See Table B.5	
Averaging time = AT	year × day/year	See Table B.5	

^a References (noted in brackets []) follow Table D.31.

NOTE: Because future use of groundwater at the Paducah Gaseous Diffusion Plant (PGDP) is uncertain, the industrial worker exposure to groundwater scenario is provided for informational purposes only. This hypothetical future exposure pathway (i.e., the industrial worker) should represent in most, if not all, locations an incomplete exposure pathway. See Table D.31 for information regarding industrial worker exposure to vapors.

Table D.2. Reasonable Maximum Exposure Assumptions for Inhalation of Volatile Organic Compounds in Water during Household Use (including Showering)^a

Exposure Concentration
$$\left(\frac{\mu g}{m^3}\right)$$

$$= \frac{\left[\left(C_{shower} \times EF \times ET_{shower}\right) + \left(C_{house} \times EF \times ET_{house}\right)\right] \times ED}{AT} \times CF$$

$$C_{shower} \left(mg/m^3\right) = \frac{\left[\left(C_{amax}/2\right) \times t_1\right] + \left[C_{amax} \times t_2\right]}{t_1 + t_2}$$

$$C_{amax} \left(mg/m^3\right) = \frac{C_{gw} \times f_{shower} \times F_w \times t_1}{V_a}$$

$$C_{house} \left(mg/m^3\right) = \frac{C_{gw} \times WHF \times f_{house}}{HV \times ER \times MC}$$

Parameter	Units	Value used	Referencesb
Time-adjusted concentration in shower = $\mathbf{C}_{\text{shower}}$	mg/m ³	Chemical-specific	Calculated
Indoor inhalation rate = IR_{air}	m ³ /hour	See Table B.5	
Exposure frequency = $\mathbf{E}\mathbf{F}$	day/year	See Table B.5	
Exposure duration = $\mathbf{E}\mathbf{D}$	years	See Table B.5	
Conversion factor = \mathbf{CF}	μg/mg	1,000	
Exposure Time = $\mathbf{ET}_{\mathbf{shower}}$	hours/day	See Table B.5	
Exposure Time = \mathbf{ET}_{house}	hours/day	See Table B.5	
Averaging time = \mathbf{AT}	$h/day \times year \times day/year$	See Table B.5	
Maximum air concentration = C_{amax}	mg/m^3	Chemical-specific	Calculated
Time of shower $= \mathbf{t_1}$	hour	See Table B.5	
Time after shower = \mathbf{t}_2	hour	See Table B.5	
Concentration in groundwater = C_{gw}	mg/L	Chemical-specific	
Fraction volatilized = $\mathbf{f}_{\text{shower}}$	unitless	See Table B.5	
Water flow rate = $\mathbf{F}_{\mathbf{w}}$	L/h	See Table B.5	
Bathroom volume = V_a	m^3	See Table B.5	
Concentration in household air = C_{house}	mg/m^3	Chemical-specific	Calculated
Water flow rate = \mathbf{WHF}	L/day	See Table B.5	
Fraction volatilized = \mathbf{f}_{house}	unitless	See Table B.5	
House volume = HV	m³/change	See Table B.5	
Exchange rate = $\mathbf{E}\mathbf{R}$	changes/day	See Table B.5	
Mixing coefficient = MC	unitless	See Table B.5	

^a Equations from [1], [14], [33], and [38].

NOTE: Because future use of groundwater at the PGDP is uncertain, the industrial worker exposure to groundwater scenario is provided for informational purposes only. This hypothetical future exposure pathway (i.e., the industrial worker) should represent in most, if not all, locations an incomplete exposure pathway. Household use for the industrial worker is assumed to be zero.

^b References (noted in brackets []) follow Table D.31.

Table D.3. Reasonable Maximum Exposure Assumptions for Dermal Contact with Water while Showering

Absorbed Dose Inorganic
$$[mg/(kg \times day)] = \frac{C_w \times SA \times Kp \times CF \times ED \times EF \times ET \times EV}{BW \times AT}$$
Absorbed Dose Organic $[mg/(kg \times day)] = \frac{DAevent \times SA \times CF \times ED \times EF \times EV}{BW \times AT}$

DA_{event} (mg/cm³-event) is calculated for organic compounds as follows:

$$\begin{split} & \text{If } t_{\text{event}} \leq t^*, \text{then: DA}_{\text{event}} = 2 \text{ FA} \times K_p \times C_w \sqrt{\frac{6\tau_{\text{event}} \times t_{\text{event}}}{\pi}} \\ & \text{If } t_{\text{event}} > t^*, \text{then: DA}_{\text{event}} = \text{FA} \times K_p \times C_w \left[\frac{t_{\text{event}}}{1+B} + 2\tau_{\text{event}} \left(\frac{1+3B+3B^2}{(1+B)^2} \right) \right] \end{split}$$

Parameter	Units	Value used	Referencesa
Chemical concentration in water = $\mathbf{C}_{\mathbf{w}}$	mg/L	Chemical-specific	
Skin surface area exposed = SA	m^2	See Table B.5	
Skin permeability constant = $\mathbf{K}_{\mathbf{p}}$	cm/hour	See Table B.6a	[40]
Absorbed dose per event = $\mathbf{D}\mathbf{A}_{event}$	mg/cm ² -event	Chemical-specific× C_w	[34]
Fraction absorbed= FA	unitless	See Table B.5	
Event time = \mathbf{t}_{event}	hours/event	Corresponds to ET	
Time to reach steady-state = \mathbf{t}^*	hour	Chemical-specific	
Lag time per event = τ_{event}	hour/event	Chemical-specific	
Dimensionless ratio of the permeability coefficient of a compound through the stratum corneum relative to its permeability coefficient across the viable epidermis $=$ \bf{B}	dimensionless	Chemical-specific	
Conversion Factor = \mathbf{CF}	$(L-m)/(cm-m^3)$	10	
Exposure duration = $\mathbf{E}\mathbf{D}$	years	See Table B.5	
Exposure frequency = \mathbf{EF}	days/year	See Table B.5	
Exposure time = $\mathbf{E}\mathbf{T}$	hours/event	See Table B.5	
$Event = \mathbf{EV}$	events/day	See Table B.5	
Body weight = \mathbf{BW}	kg	See Table B.5	
Averaging time = AT	$year \times day/year$	See Table B.5	

^a References (noted in brackets []) follow Table D.31.

NOTE: Because future use of groundwater at the PGDP is uncertain, the industrial worker exposure to groundwater scenario is provided for informational purposes only. This hypothetical future exposure pathway (i.e., the industrial worker) should represent in most, if not all, locations an incomplete exposure pathway.

Table D.4. Reasonable Maximum Exposure Assumptions for Incidental Ingestion of Soil/Sediment

Chemical Intake [mg/(kg × day)] =
$$\frac{C_s \times CF \times EF \times FI \times ED \times IR}{BW \times AT}$$

$$\mbox{Radionuclide Intake (pCi)} \, = \, \frac{\mbox{$A_s \times CF_{rad} \times EF \times FI \times ED \times IR \times (1 - e^{-\lambda \times ED})$}}{\mbox{$ED \times \lambda$}}$$

Parameter	Units	Value used	References ^a
Chemical concentration in soil/sediment = C_s	mg/kg	Chemical-specific	
Radiological activity concentration in soil/sediment = \mathbf{A}_s	pCi/g	Chemical-specific	
Conversion factor = \mathbf{CF}	kg/mg	0.000001	
Conversion factor = $\mathbf{CF}_{\mathbf{rad}}$	g/mg	0.001	
Exposure frequency = \mathbf{EF}	days/year	See Table B.5	
Fraction ingested = \mathbf{FI}	unitless	See Table B.5	
Exposure duration = ED	years	See Table B.5	
Ingestion rate = \mathbf{IR}	mg/day	See Table B.5	
Body weight = $\mathbf{B}\mathbf{W}$	kg	See Table B.5	
Averaging time = \mathbf{AT}	year × day/year	See Table B.5	
Decay constant = λ	unitless	See Table B.5	

^a References (noted in brackets []) follow Table D.31.

NOTE: For the construction/excavation worker scenario, the ED and EF can be reduced and documented on a site-specific basis, based on guidance from the Exposure Factors Handbook or similar PGDP Risk Assessment Working Group (RAWG)-approved guidance, and included in the uncertainties section of the baseline human health risk assessment.

Table D.5. Reasonable Maximum Exposure Assumptions for Dermal Contact with Soil/Sediment

Absorbed Dose
$$[mg/(kg \times day)] = \frac{C_s \times CF_d \times SA \times AF \times ABS \times EF \times ED}{BW \times AT}$$

Parameter	Units	Value used	Referencesa
Chemical concentration in soil/sediment = C_s	mg/kg	Chemical-specific	
Conversion factor = $\mathbf{CF_d}$	$(kg-cm^2)/(mg-m^2)$	0.01	
Surface area = SA	m^2/day	See Table B.5	
Adherence factor = \mathbf{AF}	mg/cm^2	See Table B.5	
Absorption factor = ABS	unitless	See Table B.6	[14]
Exposure frequency = $\mathbf{E}\mathbf{F}$	day/year	See Table B.5	
Exposure duration $=$ ED	years	See Table B.5	
Body weight = $\mathbf{B}\mathbf{W}$	kg	See Table B.5	
Averaging time = \mathbf{AT}	year × day/year	See Table B.5	

^a References (noted in brackets []) follow Table D.31.

Dermal absorbed dose is not applicable to radionuclides per guidance found in [1].
 For the construction/excavation worker scenario, the ED and EF can be reduced and documented on a site-specific basis, based on guidance from the Exposure Factors Handbook or similar PGDP RAWG-approved guidance, and included in the uncertainties section of the baseline human health risk assessment.

Table D.6. Reasonable Maximum Exposure Assumptions for Inhalation of Vapors and Particulates Emitted from Soil/Sediment^a

Exposure Concentration (
$$\mu g/m^3$$
) = $\frac{C_s \times EF \times ED \times ET \times \left(\frac{1}{VF} + \frac{1}{PEF}\right)}{AT} \times CF_1$

Radionuclide Intake (pCi) =
$$\frac{A_s \times EF \times ED \times ET \times CF_2 \times \left(\frac{1}{PEF}\right) \times (1 - e^{-\lambda \times ED})}{ED \times \lambda}$$

Parameter	Units	Value used	Referencesb
Chemical concentration in soil/sediment = C_s	mg/kg	Chemical-specific	
Activity concentration in soil/sediment = A_s	pCi/g	Chemical-specific	
Exposure frequency = \mathbf{EF}	days/year	See Table B.5	
Exposure duration = ED	years	See Table B.5	
Exposure time = \mathbf{ET}	hours/day	See Table B.5	
Conversion factor = $\mathbf{CF_1}$	μg/mg	1,000	
Conversion factor = $\mathbf{CF_2}$	g/kg	1,000	
Volatilization factor = \mathbf{VF}	m^3/kg	Chemical-specific	[18]
Particulate emission factor = PEF	m^3/kg	See Table B.5	
Averaging time = \mathbf{AT}	$\begin{array}{c} \text{hours/day} \times \text{year} \times \\ \text{day/year} \end{array}$	See Table B.5	
Decay constant = λ	unitless	See Table B.5	

^a Equation from [38].

NOTE: For the construction/excavation worker scenario, the ED and EF can be reduced and documented on a site-specific basis, based on guidance from the Exposure Factors Handbook or similar PGDP RAWG-approved guidance, and included in the uncertainties section of the baseline human health risk assessment.

^b References (noted in brackets []) follow Table D.31.

Table D.7. Reasonable Maximum Exposure Assumptions for External Exposure to Ionizing Radiation from Soil/Sediment^a

Absorbed Dose [(pCi × year)/g] =
$$\frac{A_s \times ED \times EF \times (1 - S_e) \times T_e \times (1 - e^{-\lambda \times ED})}{ED \times \lambda}$$

Parameter	Units	Value used	References ^b
Radiological activity concentration in soil/sediment = A_s	pCi/g	Chemical-specific	
Exposure duration = $\mathbf{E}\mathbf{D}$	year	See Table B.5	
Exposure frequency = \mathbf{EF}	(days/year)/(days/year)	See Table B.5	
Gamma shielding factor = S_e	unitless	See Table B.5	
Gamma exposure time factor = T_e	(hour/day)/(hour/day)	See Table B.5	
Decay constant = λ	unitless	See Table B.5	

^a Equation from [19].

NOTE: For the construction/excavation worker scenario, the ED and EF can be reduced and documented on a site-specific basis, based on guidance from the Exposure Factors Handbook or similar PGDP RAWG-approved guidance, and included in the uncertainties section of the baseline human health risk assessment.

Table D.8. Reasonable Maximum Exposure Assumptions for Incidental Ingestion of Surface Water while Swimming^a

Equations:

Chemical Intake [mg/(kg × day)] =
$$\frac{C_w \times IR \times ET \times EF \times ED}{BW \times AT}$$

Radionuclide Intake (pCi) = $A_w \times IR \times ET \times EF \times ED$

Parameter	Units	Value used	Referencesb
Chemical concentration in water = $\mathbf{C}_{\mathbf{w}}$	mg/L	Chemical-specific	
Radiological activity concentration in water = $\mathbf{A}_{\mathbf{w}}$	pCi/L	Chemical-specific	
Ingestion rate = \mathbf{IR}	L/hour	See Table B.5	
Exposure time = \mathbf{ET}	hours/day	See Table B.5	
Exposure frequency = \mathbf{EF}	days/year	See Table B.5	
Exposure duration $=$ ED	years	See Table B.5	
Body weight = \mathbf{BW}	kg	See Table B.5	
Averaging time = \mathbf{AT}	$year \times day/year$	See Table B.5	

^a Equation intended for recreational users.

^b References (noted in brackets []) follow Table D.31.

^b References (noted in brackets []) follow Table D.31.

Table D.9. Reasonable Maximum Exposure Assumptions for Dermal Contact with Surface Water (Wading)^a

Absorbed Dose Inorganic [mg/(kg × day)] =
$$\frac{C_w \times SA \times Kp \times CF \times ED \times EF \times ET}{BW \times AT}$$
Absorbed Dose Organic[mg/(kg × day)] =
$$\frac{DA_{event} \times SA \times CF \times ED \times EF \times EV}{BW \times AT}$$

DA_{event} (mg/cm³-event) is calculated for organic compounds as follows:

$$\begin{split} & \text{If } t_{\text{event}} \leq t^*, \text{then: DA}_{\text{event}} = 2 \; \text{FA} \; \times K_p \; \times \; C_w \sqrt{\frac{6\tau_{\text{event}} \; \times t_{\text{event}}}{\pi}} \\ & \text{If } t_{\text{event}} \; > t^*, \text{then: DA}_{\text{event}} = \text{FA} \times K_p \times C_w \left[\frac{t_{\text{event}}}{1+B} + 2\tau_{\text{event}} \left(\frac{1+3B+3B^2}{(1+B)^2} \right) \right] \end{split}$$

Parameter	Units	Value used	Referencesb
Chemical concentration in water = $\mathbf{C}_{\mathbf{w}}$	mg/L	Chemical-specific	
Surface area = SA	m^2	See Table B.5	
Conversion factor = \mathbf{CF}	$L/(cm - m^2)$	10	
Skin permeability constant = $\mathbf{K}_{\mathbf{p}}$	cm/hour	See Table B.6	[40]
Absorbed dose per event = $\mathbf{D}\mathbf{A}_{event}$	mg/cm ² -event	Chemical-specific× C_w	[34]
Fraction absorbed= FA	unitless	See Table B.5	
Event time = \mathbf{t}_{event}	hours/event	Corresponds to ET	
Time to reach steady-state = \mathbf{t}^*	hour	Chemical-specific	
Lag time per event = τ_{event}	hour/event	Chemical-specific	
Dimensionless ratio of the permeability coefficient of a compound through the stratum corneum relative to its permeability coefficient across the viable epidermis $=$ \bf{B}	dimensionless	Chemical-specific	
Exposure duration = $\mathbf{E}\mathbf{D}$	years	See Table B.5	
Exposure Frequency = \mathbf{EF}	days/year	See Table B.5	
Exposure time = \mathbf{ET}	hours/day	See Table B.5	
Event = EV	events/day	See Table B.5	
Body weight = \mathbf{BW}	kg	See Table B.5	

^a Equation intended for recreational users, industrial workers, outdoor workers, and excavation workers.

NOTES:

Averaging time = AT

year × day/year

See Table B.5

^b References (noted in brackets []) follow Table D.31.

^{1.} Dermal absorbed dose is not applicable to radionuclides per guidance found in [1].

For the construction/excavation worker scenario, the ED and EF can be reduced and documented on a site-specific basis, based on guidance from the Exposure Factors Handbook or similar PGDP RAWG-approved guidance, and included in the uncertainties section of the baseline human health risk assessment.

Table D.10. Reasonable Maximum Exposure Assumptions for Dermal Contact with Surface Water (Swimming)^a

Absorbed Dose Inorganic
$$[mg/(kg \times day)] = \frac{C_w \times SA \times Kp \times CF \times ED \times EF \times ET}{BW \times AT}$$

Absorbed Dose Organic $[mg/(kg \times day)] = \frac{DAevent \times SA \times CF \times ED \times EF \times EV}{BW \times AT}$

DA_{event} (mg/cm3-event) is calculated for organic compounds as follows:

$$\begin{split} & \text{If } t_{event} \leq t^*, then: DA_{event} = 2 \text{ FA} \times K_p \times C_w \sqrt{\frac{6\tau_{event} \times t_{event}}{\pi}} \\ & \text{If } t_{event} \, > t^*, then: DA_{event} = FA \times K_p \times C_w \left[\frac{t_{event}}{1+B} + 2\tau_{event} \left(\frac{1+3B+3B^2}{(1+B)^2} \right) \right] \end{split}$$

Parameter	Units	Value used	Referencesb
Chemical concentration in water = C_w	mg/L	Chemical-specific	
Surface area = SA	m^2	See Table B.5	
Conversion factor = \mathbf{CF}	$L/(cm - m^2)$	10	
Skin permeability constant = $\mathbf{K}_{\mathbf{p}}$	cm/hour	See Table B.6	[40]
Absorbed dose per event = $\mathbf{D}\mathbf{A}_{event}$	mg/cm ² -event	Chemical-specific× C_w	[34]
Fraction absorbed= FA	unitless	See Table B.5	
Event time = t_{event}	hrs/event	Corresponds to ET	
Time to reach steady-state = \mathbf{t}^*	hour	Chemical-specific	
Lag time per event = τ_{event}	hour/event	Chemical-specific	
Dimensionless ratio of the permeability coefficient of a compound through the stratum corneum relative to its permeability coefficient across the viable epidermis = $\bf B$	dimensionless	Chemical-specific	
Exposure duration = $\mathbf{E}\mathbf{D}$	years	See Table B.5	
Exposure Frequency = \mathbf{EF}	day/year	See Table B.5	
Exposure time = \mathbf{ET}	hour/day	See Table B.5	
$Event = \mathbf{EV}$	event/day	See Table B.5	
Body weight = \mathbf{BW}	kg	See Table B.5	
Averaging time = \mathbf{AT}	$year \times day/year$	See Table B.5	

^a Equation intended for recreational users.

NOTE: Dermal absorbed dose is not applicable to radionuclides per guidance found in [1].

^b References (noted in brackets []) follow Table D.31.

Table D.11. Reasonable Maximum Exposure Assumptions for Consumption of Fish

$$Chemical\ Intake\ [mg/(kg \times day)] = \frac{C_{fish}\ \times\ FI_f\ \times\ IR_f\ \times\ EF\ \times\ ED}{BW\ \times\ AT}$$

Radionuclide Intake (pCi) = $A_{fish} \times FI_f \times IR_f \times EF \times ED$

Parameter	Units	Value used	Referencesa
Chemical concentration in fish = $\mathbf{C}_{\mathbf{fish}}$	mg/kg	Chemical-specific	See Table D.12
Radiological activity concentration in fish = A_{fish}	pCi/kg	Chemical-specific	See Table D.12
Fish ingestion rate = $\mathbf{IR_f}$	kg/day	See Table B.5	
Diet fraction = $\mathbf{FI}_{\mathbf{f}}$	unitless	See Table B.5	
Exposure frequency = $\mathbf{E}\mathbf{F}$	days/year	See Table B.5	
Exposure duration = $\mathbf{E}\mathbf{D}$	years	See Table B.5	
Body weight = \mathbf{BW}	kg	See Table B.5	
Averaging time = \mathbf{AT}	$year \times day/year$	See Table B.5	

^a References (noted in brackets []) follow Table D.31.

Table D.12. Reasonable Maximum Exposure Assumptions for Concentration or Activity Concentration of COPCs in Fish

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$$C_{fish} = C_{sw} \times BAF_{fish}$$

Parameter	Units	Value used	References
Chemical concentration or radiological	mg/kg or	Chemical-specific	Calculated
activity concentration in fish = C_{fish}	pCi/kg		
Chemical concentration or radiological	mg/L or	Chemical-specific	
activity concentration in water = C_{sw}	pCi/L		
Bioaccumulation factor = BAF_{fish}	L/kg	Chemical-specific	[40]
	_	See Appendix E	

Table D.13. Reasonable Maximum Exposure Assumptions for Consumption of Deer

$$Chemical\ Intake\ [mg/(kg \times day)] = \frac{C_{deer} \times FI_d \times IR_d \times EF \times ED}{BW \times AT}$$

Radionuclide Intake (pCi) = $A_{deer} \times FI_d \times IR_d \times EF \times ED$

Parameter	Units	Value used	Referencesa
Chemical concentration in deer = $\mathbf{C}_{\mathbf{deer}}$	mg/kg	Chemical-specific	See Table D.14
Radiological activity concentration in $deer = A_{deer}$	pCi/kg	Chemical-specific	See Table D.14
Deer ingestion rate = $\mathbf{IR}_{\mathbf{d}}$	kg/day	See Table B.5	
Diet fraction = \mathbf{FI}_d	unitless	See Table B.5	
Exposure frequency = \mathbf{EF}	day/year	See Table B.5	
Exposure duration = $\mathbf{E}\mathbf{D}$	years	See Table B.5	
Body weight = \mathbf{BW}	kg	See Table B.5	
Averaging time = \mathbf{AT}	$year \times day/year$	See Table B.5	

^a References (noted in brackets []) follow Table D.31.

Table D.14. Reasonable Maximum Exposure Assumptions for Concentration or Activity Concentration of COPCs in Deera

$$C_{\text{deer}} = F_{\text{deer}} \times \left[\left(C_{\text{forage}} \times CF_{\text{rad}} \times AC \times f_{\text{s}} \times Q_{\text{f}} \right) + \left(C_{\text{s}} \times CF_{\text{rad}} \times AC \times Q_{\text{s}} \right) + \left(C_{\text{sw}} \times Q_{\text{sw}} \right) \right]$$

$$C_{\text{forage}} = \left(C_{\text{s}} \times R_{\text{upp}} \right) + \left(C_{\text{s}} \times R_{\text{es}} \right)$$

Parameter	Units	Value used	Referencesb
Chemical concentration or radiological activity	mg/kg or	Chemical-specific	Calculated
concentration in deer = $\mathbf{C}_{\text{dee}_r}$	pCi/kg		
Forage-deer transfer factor = \mathbf{F}_{deer}	day/kg	Chemical-specific	use F_{beef} values
Chemical concentration or radiological activity	mg/kg or	Chemical-specific	Calculated
concentration in forage = $\mathbf{C}_{\mathbf{forage}}$	pCi/g		
Area of contact ^{c} = AC	unitless	AS/AD	
Area of SWMU = AS	acres	SWMU-specific	
Area of deer range = AD	acres	494	[31]
Fraction of deer's food from site when on-site = \mathbf{f}_s	unitless	1.0	[5]
Quantity of forage ingested daily by deer = $\mathbf{Q}_{\mathbf{f}}$	kg/day	1.74	[7]
Chemical concentration or radiological activity	mg/kg or	Chemical-specific	
concentration in soil or sediment $= C_s$	pCi/g		
Quantity of soil ingested daily by deer = \mathbf{Q}_s	kg/day	0.034	[6]; 2% of forage
Contaminant concentration or radiological activity	mg/L or	Chemical-specific	
concentration in surface water = C_{sw}	pCi/L		
Conversion factor for radionuclides = \mathbf{CF}_{rad}	g/kg	1,000	
Quantity of surface water ingested daily by deer $^{\!\! d} = Q_{sw}$	L/day	3.61	[8]
Soil to plant uptake (dry) = \mathbf{R}_{upp}	unitless	Chemical-specific or $38 \times K_{ow}^{-0.58}$	[8]
Soil resuspension multiplier = \mathbf{R}_{es}	unitless	0.25	[3]

^a Equations after [1], [2], [3], [4]. ^b All references (noted in brackets []) follow Table D.31.

^c AC cannot be greater than 1.

^d All ingested water is assumed to be from SWMU or SWMU area.

Table D.15. Reasonable Maximum Exposure Assumptions for Consumption of Rabbit

Chemical Intake [mg/(kg × day)] =
$$\frac{C_{rabbit} \times FI_r \times IR_r \times EF \times ED}{BW \times AT}$$

Radionuclide Intake (pCi) = $A_{rabbit} \times FI_r \times IR_r \times EF \times ED$

Parameter	Units	Value used	Referencesa
Chemical concentration in rabbit = C_{rabbit}	mg/kg	Chemical-specific	See Table D.16
Radiological activity concentration in rabbit = A_{rabbit}	pCi/kg	Chemical-specific	See Table D.16
Rabbit ingestion rate = $\mathbf{IR}_{\mathbf{r}}$	kg/meal	See Table B.5	
Diet fraction = $\mathbf{FI_r}$	unitless	See Table B.5	
Exposure frequency = \mathbf{EF}	meals/year	See Table B.5	
Exposure duration = $\mathbf{E}\mathbf{D}$	years	See Table B.5	
Body weight = $\mathbf{B}\mathbf{W}$	kg	See Table B.5	
Averaging time = \mathbf{AT}	$year \times day/year$	See Table B.5	

^aReferences (noted in brackets []) follow Table D.31.

Table D.16. Reasonable Maximum Exposure Assumptions for Concentration or Activity Concentration of COPCs in Rabbita

$$C_{\text{rabbit}} = F_{\text{rabbit}} \times [(C_{\text{forage}} \times CF_{\text{rad}} \times AC \times f_{\text{s}} \times Q_{\text{f}}) + (C_{\text{s}} \times CF_{\text{rad}} \times AC \times Q_{\text{s}}) + (C_{\text{sw}} \times Q_{\text{sw}})]$$

$$C_{\text{forage}} = (C_{\text{s}} \times R_{\text{upp}}) + (C_{\text{s}} \times R_{\text{es}})$$

Parameter	Units	Value used	Referencesb
Chemical concentration or radiological activity concentration in rabbit = $\mathbf{C}_{\text{rabbit}}$	mg/kg or pCi/kg	Chemical-specific	Calculated
Forage-rabbit transfer factor = $\mathbf{F_{rabbit}}$	day/kg	Chemical-specific	use \mathbf{F}_{beef} values
Chemical concentration or radiological activity concentration in forage = \mathbf{C}_{forage}	mg/kg or pCi/g	Chemical-specific	Calculated
Area of contact ^c = \mathbf{AC}	unitless	AS/AR	
Area of $SWMU = AS$	acres	SWMU-specific	
Area of rabbit range = $\mathbf{A}\mathbf{R}$	acres	3.6	[28]
Fraction of rabbit's food from site when on-site = $\mathbf{f_s}$	unitless	1.0	
Quantity of forage ingested daily by rabbit = $\mathbf{Q}_{\mathbf{f}}$	kg/day	0.237	[29]
Chemical concentration or radiological activity concentration in soil or sediment = C_s	mg/kg or pCi/g	Chemical-specific	
Quantity of soil ingested daily by rabbit = \mathbf{Q}_s	kg/day	0.0149	[29] 6.3% of forage
Contaminant concentration in surface water = \mathbf{C}_{sw}	mg/L or pCi/L	Chemical-specific	
Conversion factor for radionuclides = \mathbf{CF}_{rad}	g/kg	1,000	
Quantity of surface water ingested daily by rabbit \mathbf{Q}_{sw}	L/day	0.116	[29]
Soil to plant uptake $(dry) = \mathbf{R}_{upp}$	unitless	Chemical-specific or $38 \times K_{ow}^{-0.58}$	[8]
Soil resuspension multiplier = \mathbf{R}_{es}	unitless	0.25	[3]

^a Equations after [1], [2], [3], [4]. ^b All references (noted in brackets []) follow Table D.31.

^c AC cannot be greater than 1.

^d All ingested water is considered to be from SWMU or SWMU area.

Table D.17. Reasonable Maximum Exposure Assumptions for Consumption of Quail

Chemical Intake [mg/(kg × day)] =
$$\frac{C_{quail} \times FI_q \times IR_q \times EF \times ED}{BW \times AT}$$

Radionuclide Intake (pCi) = $A_{quail} \times FI_q \times IR_q \times EF \times ED$

Parameter	Units	Value used	Referencesa
Chemical concentration in quail = \mathbf{C}_{quail}	mg/kg	Chemical-specific	See Table D.18
Radiological activity concentration in quail = \mathbf{A}_{quail}	pCi/kg	Chemical-specific	See Table D.18
Quail ingestion rate = $\mathbf{IR}_{\mathbf{q}}$	kg/meal	See Table B.5	
Diet fraction = $\mathbf{FI}_{\mathbf{q}}$	unitless	See Table B.5	
Exposure frequency = $\mathbf{E}\mathbf{F}$	meals/year	See Table B.5	
Exposure duration = $\mathbf{E}\mathbf{D}$	years	See Table B.5	
Body weight = \mathbf{BW}	kg	See Table B.5	
Averaging time = AT	year × day/year	See Table B.5	

^a References (noted in brackets []) follow Table D.31.

Table D.18. Reasonable Maximum Exposure Assumptions for Concentration or Activity Concentration of COPCs in Quail^a

$$\begin{aligned} C_{quail} &= F_{quail} \\ &\times \left[\left(C_{forage} \times CF_{rad} \times AC \times f_{s} \times Q_{f} \right) + \left(C_{s} \times CF_{rad} \times AC \times Q_{s} \right) + \left(C_{sw} \times Q_{sw} \right) \right. \\ &\left. + \left(C_{i} \times CF_{rad} \times AC \times Q_{i} \right) \right] \\ &C_{forage} &= \left(C_{s} \times R_{upp} \right) + \left(C_{s} \times R_{es} \right) \end{aligned} \qquad C_{i} &= \left(C_{s} \times BAF_{i} \right) \end{aligned}$$

Parameter	Units	Value used	References ^b
Chemical concentration or radiological activity concentration in quail = $\mathbf{C}_{\mathbf{quail}}$	mg/kg or pCi/kg	Chemical-specific	Calculated
Forage-quail transfer factor = \mathbf{F}_{quail}	day/kg	Chemical-specific	use $\mathbf{F}_{poultry}$ values
Chemical concentration or radiological activity concentration in forage = C_{forage}	mg/kg or pCi/g	Chemical-specific	Calculated
Area of contact ^c = \mathbf{AC}	unitless	AS/AQ	
Area of $SWMU = AS$	acres	SWMU-specific	
Area of quail range = \mathbf{AQ}	acres	15.4	[28]
Fraction of quail's food from site when on-site = \mathbf{f}_s	unitless	1.0	
Quantity of forage ingested daily by quail = $\mathbf{Q}_{\mathbf{f}}$	kg/day	0.01499	[28] 88.2% of total food
Chemical concentration or radiological activity concentration in invertebrates = C_i	mg/kg or pCi/g	Chemical-specific	
Quantity of invertebrates ingested daily by quail = \mathbf{Q}_i	kg/day	0.002006	[28] 11.8% of total food
Chemical concentration or radiological activity concentration in soil or sediment = \mathbf{C}_s	mg/kg or pCi/g	Chemical-specific	
Quantity of soil ingested daily by quail = Q_s	kg/day	0.00195	[17]
Contaminant concentration or radiological activity concentration in surface water = \mathbf{C}_{sw}	mg/L or pCi/L	Chemical-specific	
Conversion factor for radionuclides = \mathbf{CF}_{rad}	g/kg	1,000	
Quantity of surface water ingested daily by quail ^d = \mathbf{Q}_{sw}	L/day	0.02	[17]
Soil to plant uptake (dry) = \mathbf{R}_{upp}	unitless	Chemical-specific or $38 \times K_{ow}^{-0.58}$	[8]
Soil resuspension multiplier = \mathbf{R}_{es}	unitless	0.25	[3]

^a Equations after [1], [2], [3], [4].

^b All references (noted in brackets []) follow Table D.31.

^c AC cannot be greater than 1.

^d All ingested water is considered to be from SWMU or SWMU area.

Table D.19. Reasonable Maximum Exposure Assumptions for Consumption of Homegrown Vegetables

$$\text{Chemical Intake } [\text{mg/(kg} \times \text{day})] = \frac{C_{\text{vegetables}} \times \text{FI}_{\text{v}} \times \text{IR}_{\text{v}} \times \text{EF} \times \text{ED}}{\text{BW} \times \text{AT}}$$

Radionuclide Intake (pCi) = $A_{\text{vegetables}} \times FI_v \times IR_v \times EF \times ED$

Parameter	Units	Value used	Referencesa
Chemical concentration in vegetables = C _{vegetables}	mg/kg	Chemical-specific	See Table D.20
Radiological activity concentration in vegetables = $\mathbf{A}_{\text{vegetables}}$	pCi/kg	Chemical-specific	See Table D.20
Diet fraction = $\mathbf{FI}_{\mathbf{v}}$	unitless	See Table B.5	
Vegetable ingestion rate = $\mathbf{IR}_{\mathbf{v}}$	kg/day	See Table B.5	
Exposure frequency = $\mathbf{E}\mathbf{F}$	days/year	See Table B.5	
Exposure duration $=$ ED	years	See Table B.5	
Body weight = \mathbf{BW}	kg	See Table B.5	
Averaging time = \mathbf{AT}	$year \times day/year$	See Table B.5	

^a References (noted in brackets []) follow Table D.31.

Table D.20. Reasonable Maximum Exposure Assumptions for Concentration or Activity Concentration of COPCs in Homegrown Vegetables^a

$$\begin{split} C_{vegetables} &= \left(C_w \times Irr_{rup} \right) + \left(C_s \times CF_{rad} \times AC \times R_{upv} \right) + \left(C_w \times Irr_{res} \right) \\ &+ \left(C_s \times CF_{rad} \times AC \times R_{es} \right) + \left(C_w \times Irr_{dep} \right) \\ C_{vegetables} &= \left(C_w \times Irr_{rup} \right) + \left(C_s \times CF_{rad} \times AC \times R_{upv} \right) + \left(C_w \times Irr_{res} \right) + \left(C_s \times CF_{rad} \times AC \times R_{es} \right) \\ &+ \left(C_w \times Irr_{dep} \right) \\ Irr_{rup} &= \frac{Ir \times F \times Bv_{wet} \times [1 - exp(-\lambda_B \times t_b)]}{P \times \lambda_B} \qquad Irr_{dep} &= \frac{Ir \times F \times I_f \times T \times [1 - exp(-\lambda_E \times t_v)]}{Y_v \times \lambda_E} \\ Irr_{res} &= \frac{Ir \times F \times MLF \times [1 - exp(-\lambda_B \times t_b)]}{P \times \lambda_B} \end{split}$$

Parameter	Units	Value used	References ^b
Chemical concentration or radiological	mg/kg or pCi/kg	Chemical-specific	Calculated
activity concentration in vegetable = $\mathbf{C}_{\text{vegetables}}$			
Chemical concentration or radiological	mg/L or pCi/L	Chemical-specific	
activity concentration in groundwater = $\mathbf{C}_{\mathbf{w}}$			
Root uptake from irrigation = Irr_{rup}	L/kg	Chemical-specific	Calculated
Conversion factor for radionuclides = \mathbf{CF}_{rad}	g/kg	1,000	
Chemical concentration or radiological	mg/kg or pCi/g	Chemical-specific	
activity concentration in soil = C_s Area of contact ^c = AC	unitless	AS/AG	
Area of SWMU = AS	acres	SWMU-specific	
Area of garden = \mathbf{AG}	acres	0.25	[30]
Wet root uptake for leafy vegetables = \mathbf{R}_{upv}	kg/kg	Chemical-specific	[41]
Resuspension from irrigation = Irr _{res}	L/kg	Chemical-specific	Calculated
Resuspension multiplier = \mathbf{R}_{es}	unitless	0.26	[9]
Aerial deposition from irrigation = Irr_{dep}	L/kg	Chemical-specific	Calculated
Irrigation rate = \mathbf{Ir}	L/m²-day	3.62	[10]
Irrigation period = \mathbf{F}	unitless	0.25	[10]; 3 months a year
Soil to plant uptake, wet weight = $\mathbf{B}\mathbf{v}_{wet}$	kg/kg	Chemical-specific or $7.7 \times K_{ow}^{-0.58}$	[11]
Effective rate for removal = λ_B	1/day	$\lambda_i + \lambda_{HL}$	[11]
$Decay = \lambda_i$	1/day	$0.693/T_{r}$	[11]
$Half$ -life = $\mathbf{T_r}$	day	Chemical-specific	[40]
Soil leaching rate = λ_{HL}	1/day	2.7×10^{-5}	[11]
Long-term deposition and build-up = \mathbf{t}_{b}	day	10,950	[2]
Area density for root zone = \mathbf{P}	kg/m^2	240	[8], [12], [13]
Plant mass leading factor = MLF	unitless	0.26	[9]
Interception fraction = I_f	unitless	0.42	[7]
Translocation factor = \mathbf{T}	unitless	1	[2]
Decay for removal on produce = λ_E	1/day	$\lambda_i + (0.693/t_w)$	[11]
Weathering half-life = t_w	day	14	[2]
Above ground exposure time = $\mathbf{t}_{\mathbf{v}}$	day	60	[2]
Plant yield (wet) = $\mathbf{Y}_{\mathbf{v}}$	kg/m^2	2	[2]

^a Equations after [1], [2], [3], [4].

^b References (noted in brackets []) follow Table D.31.

^c AC cannot be greater than 1.

Table D.21. Reasonable Maximum Exposure Assumptions for Consumption of Beef

Chemical Intake [mg/(kg × day)] =
$$\frac{C_{beef} \times FI_b \times IR_b \times EF \times ED}{BW \times AT}$$

Radionuclide Intake (pCi) = $A_{beef} \times FI_b \times IR_b \times EF \times ED$

Parameter	Units	Value used	Referencesa
Chemical concentration in beef = $\mathbf{C}_{\mathbf{beef}}$	mg/kg	Chemical-specific	See Table D.22
Radiological activity concentration in beef = A _{beef}	pCi/kg	Chemical-specific	See Table D.22
Beef ingestion rate = IR_b	kg/day	See Table B.5	
Diet fraction = $\mathbf{FI}_{\mathbf{b}}$	unitless	See Table B.5	
Exposure frequency = \mathbf{EF}	days/year	See Table B.5	
Exposure duration $=$ ED	years	See Table B.5	
Body weight = $\mathbf{B}\mathbf{W}$	kg	See Table B.5	
Averaging time = \mathbf{AT}	$year \times day/year$	See Table B.5	

^a References (noted in brackets []) follow Table D.31.

Table D.22. Reasonable Maximum Exposure Assumptions for Concentration or Activity Concentration of COPCs in Beefa

$$\begin{aligned} C_{beef} &= F_{beef} \times \left[(C_{forage} \times AC \times f_{s} \times Q_{f}) + (C_{s} \times AC \times Q_{s}) + (C_{w} \times CF_{rad} \times Q_{w}) \right] \\ C_{forage} &= (C_{s} \times R_{upp}) + (C_{s} \times R_{es}) \end{aligned}$$

Parameter	Units	Value used	References ^b
Chemical concentration or radiological activity concentration in beef = \mathbf{C}_{beef}	mg/kg or pCi/kg	Chemical-specific	Calculated
Forage-beef transfer factor = \mathbf{F}_{beef}	day/kg	Chemical-specific	[41]
Chemical concentration or radiological activity concentration in pasture = C_{forage}	mg/kg or pCi/g	Chemical-specific	Calculated
Area of contact ^c = \mathbf{AC}	unitless	AS/AD	
Area of $SWMU = AS$	acres	SWMU-specific	
Area of beef range = AD	acres	2	[27]
Fraction of beef's food from site when on-site = f_s	unitless	1.0	[5]
Quantity of pasture ingested daily by beef = \mathbf{Q}_f	kg/day	25	[23]
Chemical concentration or radiological activity concentration in soil or sediment = C_s	mg/kg or pCi/g	Chemical-specific	
Quantity of soil ingested daily by beef = \mathbf{Q}_s	kg/day	1	[24]
Contaminant concentration or radiological activity concentration in water = $\mathbf{C}_{\mathbf{w}}$	mg/L or pCi/L	Chemical-specific	
Conversion factor for radionuclides = \mathbf{CF}_{rad}	g/kg	1,000	
Quantity of water ingested daily by $beef^d = \mathbf{Q}_w$	L/day	50	[23]
Soil to plant uptake (dry) = \mathbf{R}_{upp}	unitless	Chemical-specific or $38 \times K_{ow}^{-0.58}$	[8]
Soil resuspension multiplier = \mathbf{R}_{es}	unitless	0.25	[3]

^a Equations after [1], [2], [3], [4]. ^b All references (noted in brackets []) follow Table D.31.

^c AC cannot be greater than 1.
^d All ingested water is considered to be from SWMU or SWMU area.

Table D.23. Reasonable Maximum Exposure Assumptions for Consumption of Milk

$$Chemical\ Intake\ [mg/(kg \times day)] = \frac{C_{milk}\ \times\ FI_{m}\ \times\ IR_{m}\ \times\ EF\ \times\ ED}{BW \times AT}$$

Radionuclide Intake (pCi) = $A_{milk} \times FI_m \times IR_m \times EF \times ED$

Parameter	Units	Value used	Referencesa
Chemical concentration in $milk = C_{milk}$	mg/kg	Chemical-specific	See Table D.24
Radiological activity concentration in $milk = A_{milk}$	pCi/kg	Chemical-specific	See Table D.24
Milk ingestion rate = IR_m	kg/day	See Table B.5	
Diet fraction = $\mathbf{FI}_{\mathbf{m}}$	unitless	See Table B.5	
Exposure frequency = \mathbf{EF}	day/year	See Table B.5	
Exposure duration $=$ ED	years	See Table B.5	
Body weight = \mathbf{BW}	kg	See Table B.5	
Averaging time = \mathbf{AT}	$year \times day/year$	See Table B.5	

^a References (noted in brackets []) follow Table D.31.

Table D.24. Reasonable Maximum Exposure Assumptions for Concentration or Activity Concentration of COPCs in Milk^a

$$C_{\text{milk}} = F_{\text{milk}} \times \left[(C_{\text{forage}} \times CF_{\text{rad}} \times AC \times f_{\text{s}} \times Q_{\text{f}}) + (C_{\text{s}} \times CF_{\text{rad}} \times AC \times Q_{\text{s}}) + (C_{\text{w}} \times Q_{\text{w}}) \right]$$

$$C_{\text{forage}} = (C_{\text{s}} \times R_{\text{upp}}) + (C_{\text{s}} \times R_{\text{es}})$$

Parameter	Units	Value used	Referencesb
Chemical concentration or radiological activity concentration in $milk = C_{milk}$	mg/kg or pCi/kg	Chemical-specific	Calculated
Forage-milk transfer factor = \mathbf{F}_{milk}	day/kg	Chemical-specific	[41]
Chemical concentration or radiological activity concentration in pasture = \mathbf{C}_{forage}	mg/kg or pCi/g	Chemical-specific	Calculated
Area of contact ^c = \mathbf{AC}	unitless	AS/AD	
Area of $SWMU = AS$	acres	SWMU-specific	
Area of dairy range = AD	acres	2	[27]
Fraction of dairy's food from site when on-site = $\mathbf{f_s}$	unitless	1.0	[5]
Quantity of pasture ingested daily by dairy = $\mathbf{Q}_{\mathbf{f}}$	kg/day	25	[23]
Chemical concentration or radiological activity concentration in soil or sediment = C_s	mg/kg or pCi/g	Chemical-specific	
Quantity of soil ingested daily by dairy = \mathbf{Q}_s	kg/day	1	[24]
Contaminant concentration or radiological activity concentration in water = $\mathbf{C}_{\mathbf{w}}$	mg/L or pCi/L	Chemical-specific	
Conversion factor for radionuclides = \mathbf{CF}_{rad}	g/kg	1,000	
Quantity of water ingested daily by dairy ^d = $\mathbf{Q}_{\mathbf{w}}$	L/day	60	[23]
Soil to plant uptake (dry) = \mathbf{R}_{upp}	unitless	Chemical-specific or $38 \times K_{ow}^{-0.58}$	[8]
Soil resuspension multiplier = \mathbf{R}_{es}	unitless	0.25	[3]

^a Equations after [1], [2], [3], [4].

^b All references (noted in brackets []) follow Table D.31.

^c AC cannot be greater than 1.

^d All ingested water is considered to be from SWMU or SWMU area.

Table D.25. Reasonable Maximum Exposure Assumptions for Consumption of Poultry

$$\text{Chemical Intake } [\text{mg/(kg} \times \text{day})] = \frac{C_{\text{poultry}} \times \text{FI}_{\text{p}} \times \text{IR}_{\text{p}} \times \text{EF} \times \text{ED}}{\text{BW} \times \text{AT}}$$

Radionuclide Intake (pCi) = $A_{poultry} \times FI_p \times IR_p \times EF \times ED$

Parameter	Units	Value used	Referencesa
Chemical concentration in poultry = $C_{poultry}$	mg/kg	Chemical-specific	See Table D.26
Radiological activity concentration in poultry = $\mathbf{A}_{poultry}$	pCi/kg	Chemical-specific	See Table D.26
Poultry ingestion rate = $\mathbf{IR}_{\mathbf{p}}$	kg/day	See Table B.5	
Diet fraction = $\mathbf{FI}_{\mathbf{p}}$	unitless	See Table B.5	
Exposure frequency = \mathbf{EF}	day/year	See Table B.5	
Exposure duration = $\mathbf{E}\mathbf{D}$	years	See Table B.5	
Body weight = \mathbf{BW}	kg	See Table B.5	
Averaging time = \mathbf{AT}	$year \times day/year$	See Table B.5	

^a References (noted in brackets []) follow Table D.31.

Table D.26. Reasonable Maximum Exposure Assumptions for Concentration or Activity Concentration of COPCs in Poultry^a

$$C_{poultry} = F_{poultry} \times [(C_{forage} \times CF_{rad} \times AC \times f_s \times Q_f) + (C_s \times CF_{rad} \times AC \times Q_s) + (C_w \times Q_w)]$$

$$C_{forage} = (C_s \times R_{upp}) + (C_s \times R_{es})$$

Parameter	Units	Value used	References ^b
Chemical concentration or radiological activity concentration in poultry = $\mathbf{C}_{poultry}$	mg/kg or pCi/kg	Chemical-specific	Calculated
Forage-poultry transfer factor = $\mathbf{F}_{poultry}$	day/kg	Chemical-specific	[32], [39]
Chemical concentration or radiological activity concentration in pasture = $\mathbf{C}_{\mathbf{forage}}$	mg/kg or pCi/g	Chemical-specific	Calculated
Area of contact ^c = \mathbf{AC}	unitless	AS/AD	
Area of SWMU = AS	Acres	SWMU-specific	
Area of poultry range = $\mathbf{A}\mathbf{D}^{d}$	Acres	1	[27]
Fraction of poultry's food from site = \mathbf{f}_s	unitless	0.5	[27] assumes broilers get 50% bought grain
Quantity of pasture ingested daily by poultry = $\mathbf{Q}_{\mathbf{f}}$	kg/day	0.12 (chicken) 0.35 (turkey)	[22] 20 wk old male turkey
Chemical concentration or radiological activity concentration in soil or sediment = C_s	mg/kg or pCi/g	Chemical-specific	
Quantity of soil ingested daily by poultry = \mathbf{Q}_s	kg/day	0.0024 (chicken 0.007 (turkey)	[8] same ratio for chicken
Contaminant concentration in water = $\mathbf{C}_{\mathbf{w}}$	mg/L or pCi/L	Chemical-specific	
Conversion factor for radionuclides = \mathbf{CF}_{rad}	g/kg	1,000	
Quantity of water ingested daily by poultry $^{e} = \mathbf{Q}_{w}$	L/day	0.24 (chicken) 1.0 (turkey)	[22] 1:2 ratio of 20 wk old male turkey
Soil to plant uptake (dry) = \mathbf{R}_{upp}	unitless	Chemical-specific or 38×K _{ow} -0.58	[8]
Soil resuspension multiplier = $\mathbf{R_{es}}$	unitless	0.25	[3]

NOTE: Under this model, poultry raised for use as broilers by subsistence farmers are allowed to forage on pasture where they ingest pasture and

^a Equations after [1], [2], [3], [4]. ^b All references (noted in brackets []) follow Table D.31.

^c AC cannot be greater than 1.

^d Assumes 1 acre of pasture for 200 adult birds with a three year rotation.

^e All ingested water is considered to be from SWMU or SWMU area.

Table D.27. Reasonable Maximum Exposure Assumptions for Consumption of Pork

$$Chemical\ Intake\ [mg/(kg \times day)] = \frac{C_{pork} \times FI_{pork} \times IR_{pork} \times EF \times ED}{BW \times AT}$$

Radionuclide Intake (pCi) = $A_{pork} \times FI_{pork} \times IR_{pork} \times EF \times ED$

Parameter	Units	Value used	Referencesa
Chemical concentration in pork = \mathbf{C}_{pork}	mg/kg	Chemical-specific	See Table D.28
Radiological activity concentration in pork = \mathbf{A}_{pork}	pCi/kg	Chemical-specific	See Table D.28
Pork ingestion rate = IR_{pork}	kg/day	See Table B.5	
Diet fraction = \mathbf{FI}_{pork}	unitless	See Table B.5	
Exposure frequency = \mathbf{EF}	days/year	See Table B.5	
Exposure duration = $\mathbf{E}\mathbf{D}$	years	See Table B.5	
Body weight = \mathbf{BW}	kg	See Table B.5	
Averaging time = \mathbf{AT}	$year \times day/year$	See Table B.5	

^a References (noted in brackets []) follow Table D.31.

Table D.28. Reasonable Maximum Exposure Assumptions for Concentration or Activity Concentration of COPCs in Pork^a

$$C_{\text{pork}} = F_{\text{pork}} \times [(C_{\text{forage}} \times CF_{\text{rad}} \times AC \times f_{\text{s}} \times Q_{\text{f}}) + (C_{\text{s}} \times CF_{\text{rad}} \times AC \times Q_{\text{s}}) + (C_{\text{w}} \times Q_{\text{w}})]$$

$$C_{\text{forage}} = (C_{\text{s}} \times R_{\text{upp}}) + (C_{\text{s}} \times R_{\text{es}})$$

Parameter	Units	Value used	References ^b
Chemical concentration or radiological activity concentration in pork = C_{pork}	mg/kg or pCi/kg	Chemical-specific	Calculated
Forage-pork transfer factor = \mathbf{F}_{pork}	day/kg	Chemical-specific	[32], [39]
Chemical concentration or radiological activity concentration in pasture = \mathbf{C}_{forage}	mg/kg or pCi/g	Chemical-specific	Calculated
Area of contact $^{c} = AC$	unitless	AS/AD	
Area of SWMU = AS	acres	SWMU-specific	
Area of swine range = AD	acres	1	[27]
Fraction of swine's food from site = \mathbf{f}_s	unitless	0.4	[27]
Quantity of pasture ingested daily by swine = $\mathbf{Q}_{\mathbf{f}}$	kg/day	2.4	[32]
Chemical concentration or radiological activity concentration in soil or sediment = C_s	mg/kg or pCi/g	Chemical-specific	
Quantity of soil ingested daily by swine = Q_s	kg/day	0.034	[26]
Chemical concentration or radiological activity concentration in water = $\mathbf{C}_{\mathbf{w}}$	mg/L or pCi/L	Chemical-specific	
Conversion factor for radionuclides $= \mathbf{CF}_{rad}$	g/kg	1,000	
Quantity of water ingested daily by swine $^{d}=\mathbf{Q}_{w}$	L/day	6.14	[25] 2.56 to 1, water to feed ratio
Soil to plant uptake (dry) = \mathbf{R}_{upp}	unitless	Chemical-specific or $38 \times K_{\rm ow}^{-0.58}$	[8]
Soil resuspension multiplier = \mathbf{R}_{es}	unitless	0.25	[3]

^a Equations after [1], [2], [3], [4].

NOTE: According to Morrison (1956), subsistence farmers allow 20% to 40% of the swine's diet to come from pasture, while the remaining comes from store bought grain.

^b All references (noted in brackets []) follow Table D.31.

^c AC cannot be greater than 1.

^d All ingested water is considered to be from SWMU or SWMU area.

Table D.29. Reasonable Maximum Exposure Assumptions for Consumption of Eggs

Chemical Intake [mg/(kg × day)] =
$$\frac{C_{egg} \times FI_e \times IR_e \times EF \times ED}{BW \times AT}$$

Radionuclide Intake (pCi) =
$$A_{egg} \times FI_{e} \times IR_{e} \times EF \times ED$$

Parameter	Units	Value used	Referencesa
Chemical concentration in egg = $\mathbf{C}_{\mathbf{egg}}$	mg/kg	Chemical-specific	See Table D.30
Radiological activity concentration in egg = \mathbf{A}_{egg}	pCi/kg	Chemical-specific	See Table D.30
Egg ingestion rate = \mathbf{IR}_{e}	kg/day	See Table B.5	
Diet fraction = \mathbf{FI}_{e}	unitless	See Table B.5	
Exposure frequency = \mathbf{EF}	day/year	See Table B.5	
Exposure duration = $\mathbf{E}\mathbf{D}$	years	See Table B.5	
Body weight = \mathbf{BW}	kg	See Table B.5	
Averaging time = \mathbf{AT}	year × day/year	See Table B.5	

^a References (noted in brackets []) follow Table D.31.

Table D.30. Reasonable Maximum Exposure Assumptions for Concentration or Activity Concentration of COPCs in Egg^a

Equations:

$$C_{egg} = F_{egg} \times (C_w \times Q_w)$$

Parameter	Units	Value used	References ^b
Chemical concentration or radiological activity concentration in egg = \mathbf{C}_{egg}	mg/kg or pCi/kg	Chemical-specific	Calculated
Forage-egg transfer factor = \mathbf{F}_{egg}	day/kg	Chemical-specific	[32], [39]
Chemical concentration or radiological activity concentration in water = $\mathbf{C}_{\mathbf{w}}$	mg/L or pCi/L	Chemical-specific	
Quantity of water ingested daily by poultry = $\mathbf{Q}_{\mathbf{w}}$	L/day	0.24 (chicken) 1.0 (turkey)	[22] 1:2 ratio of 20 wk old male turkey

^a Equations after [1], [2], [3], [4].

NOTE: Model assumes that laying hens are in a hutch and are not allowed to forage on pasture. Therefore, they eat only store bought grain and are not exposed to pasture or soil. Drinking water is assumed to come from the SWMU or SWMU area.

^b All references (noted in brackets []) follow Table D.31.

Table D.31. Reasonable Maximum Exposure Assumptions for Inhalation of Vapors in Ambient Aira

Exposure Concentration (
$$\mu g/m^3$$
) = $\frac{C_a \times EF \times ED \times ET}{AT}$

Parameter	Units	Value used	Referencesb
Chemical concentration in $Air = C_a$	$\mu g/m^3$	Chemical-specific	
Exposure frequency = \mathbf{EF}	days/year	See Table B.5	
Exposure duration = \mathbf{ED}	years	See Table B.5	
Exposure time = \mathbf{ET}	hours/day	See Table B.5	
Averaging time = \mathbf{AT}	$\begin{array}{c} \text{hours/day} \times \text{year} \times \\ \text{day/year} \end{array}$	See Table B.5	

^a Equation from [38].

NOTE: For the construction/excavation worker scenario, the ED and EF can be reduced and documented on a site-specific basis, based on guidance from the Exposure Factors Handbook or similar PGDP RAWG-approved guidance, and can be included in the uncertainties section of the baseline human health risk assessment.

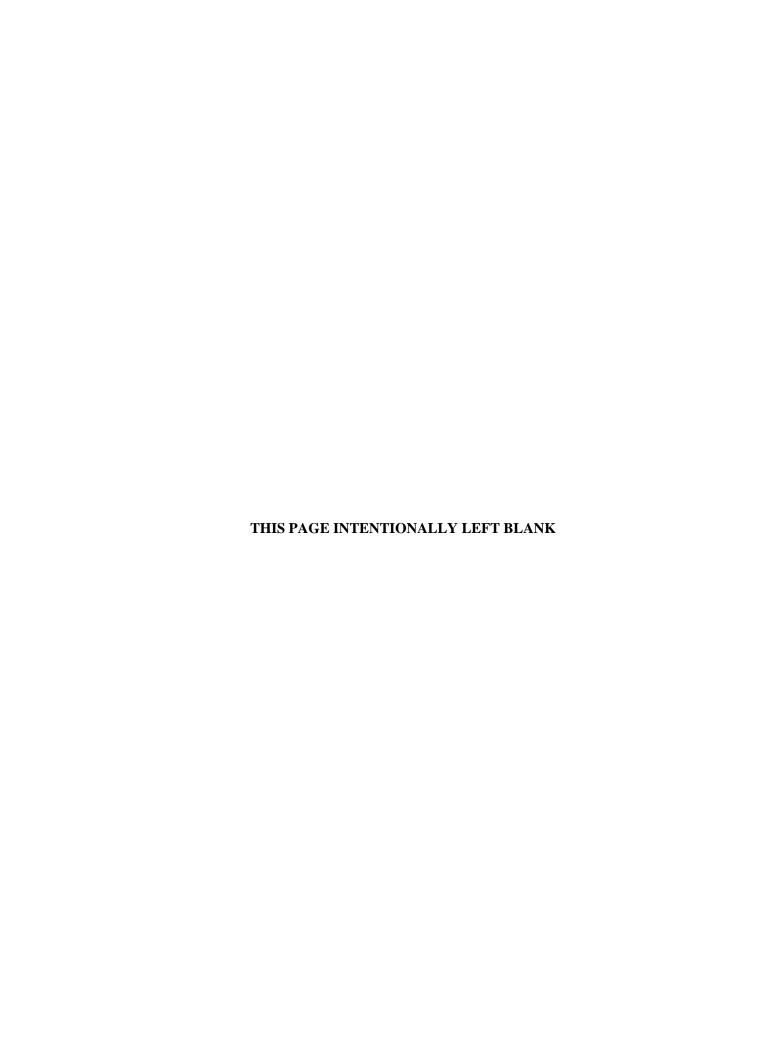
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APPENDIX E TECHNICAL INFORMATION



APPENDIX E TECHNICAL INFORMATION (CD)

